Exhibit 508

Expert Report of Judy S. LaKind, Ph.D.
In the Matter of *Keller v. United States*

Prepared by
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8 April 2025

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ACRONYMS/ABBREVIATIONS

ADD: Average Daily Dose
AT: averaging time
ATSDR: Agency for Toxic Substances and Disease Registry
BW: body weight (kg)
C: contaminant air concentration (μg/m^3)
CASRN: Chemical Abstracts Service Registry Number
cm^2: square centimeter
CTE: central tendency exposure
D: age-specific dose (mg/kg-day)
DA_{event}: absorbed dose per event (μg/cm^2/event)
DAD: dermal absorbed dose (μg/kg/day)
DCE: *trans*-1,2-dichloroethylene
ED: exposure duration (year)
EDG: Exposure Data Guidance
EF (intermediate or chronic): exposure factor (unitless) = (F x ED)/AT
EPC: exposure point concentration, contaminant concentration (mg/L)
EV: event frequency
F: exposure frequency (day/week x week/year)
FM: Field Manual
ft: feet
hr: hour
ID: inhalation dose (μg/kg/day)
IR: intake rate of water (L/day) or air (m^3/day)
kg: kilogram
L: liter
LADD: Lifetime Average Daily Dose
L/min: liters air breathed per minute
m^3: cubic meter
mg: milligram
mg/kg-day: milligram chemical per kilogram body weight per day
N: number
NHANES: National Health and Nutrition Examination Survey
PCE: perchloroethylene
PHAST: Public Health Assessment Site Tool
RME: reasonable maximum exposure
SA: dermal surface area (cm^2)
SHOWER: Shower and Household Water-use Exposure
TCE: trichloroethylene
μg/L: microgram per liter
US: United States
US DOJ: United States Department of Justice

US EPA: United States Environmental Protection Agency
v: version
VC: vinyl chloride
WTP: water treatment plant

1. QUALIFICATIONS

I am Judy S. LaKind, MS, Ph.D. I am President of LaKind Associates, LLC, a human health risk science firm specializing in exposure science and the evaluation of scientific data for regulatory decision-making. I have over 30 years of experience in the fields of exposure science and risk assessment. I have expertise in assessing child and adult exposures to environmental chemicals, risk assessment and the implications of uncertainty in the risk assessment process, evaluation of data quality, use of environmental epidemiology research in public health decision-making, weighing potential risks and benefits related to chemical use, and systematic review. I am an adjunct Associate Professor in the Department of Epidemiology and Public Health, University of Maryland School of Medicine. I am also a Fellow by Courtesy, Department of Applied Mathematics and Statistics, The Johns Hopkins University.

I have a B.A. from The Johns Hopkins University, an MS from University of Wisconsin, Madison in geology and a Ph.D. from The Johns Hopkins University in environmental engineering. My dissertation research was on the kinetics of reductive dissolution of iron oxyhydroxides by phenolic compounds. In 1988, I was a scientist at the US Environmental Protection Agency (US EPA) where one of my main activities was reviewing environmental impact assessments produced under the National Environmental Policy Act. I was a scientist at consulting firms from 1988 to 1998 during which time my work focused on the conduct of exposure and risk assessments (e.g., field, computational, and communication aspects). From 1998 until the present, I have been a self-employed scientist specializing in exposure science, assessment of human health risks, biomonitoring, scientific analysis for regulatory support, and state-of-the-science and systematic reviews. I have extensive experience in speaking and publishing on exposure- and risk-related issues, including children's exposures to environmental chemicals, the implications of uncertainty in the risk assessment process, data quality, use of environmental epidemiology research in public health decision-making, weighing potential risks and benefits related to chemical use, the presence of environmental chemicals in human milk, and time-dependence and distributional analysis of exposure. I have evaluated the use of human health risk assessments in the development of water quality criteria and have critically analyzed the environmental fate, behavior, and bioavailability of pollutants in the context of setting regulatory criteria. I have developed risk assessments for a variety of urban industrial sites, military bases, and firing ranges, and have utilized state-of-the-science models for estimating blood lead levels in adults and children.

I have taught or co-taught courses on aquatic chemistry (Johns Hopkins University) and risk assessment (Johns Hopkins University, the University of Maryland School of Law and the University of Maryland, Baltimore County). I also co-taught a short course on biomonitoring and have developed an on-line course for continuing medical education credit on chemical exposures and health effects.

From 2008 to 2009, I served as Environmental Health Advisor to the Maryland Department of the Environment, Science Services Administration. One of my many activities was to develop standard operating procedures for developing risk-based fish consumption advisories.

I am a past President of the International Society of Exposure Science and served on the Executive Committee of the Exposure Specialty Section of the Society of Toxicology. I am also a member of the American Chemical Society, Environmental Division, and the Society for Risk Analysis. I was a founding member of the International Society for Children's Health and the Environment (2009-2015). I am a former member of the Health Effects Institute Energy Research Committee. I previously served on the Board of the Coalition Against Childhood Lead Poisoning (with a term as president). I was also a member of Maryland's Children's Environmental Health and Protection Advisory Council, the Maryland Lead Poisoning Prevention Commission, the Maryland Pesticide Reporting and Information Workgroup, the Maryland Department of Health and Mental Hygiene Cancer Cluster Advisory Committee, the Health and Environmental Sciences Institute (HESI) RISK21 Advisory Board, and the World Health Organization (WHO) Survey Coordinating Committee for the WHO Global Survey of Human Milk for Persistent Organic Pollutants. I also served on the Institute of Medicine Committee on Blue Water Navy Vietnam Veterans and Agent Orange Exposure and the US Environmental Protection Agency Science Advisory Board Panel on Perchlorate - Approaches for Deriving Maximum Contaminant Level Goals for Drinking Water.

I have published over 100 papers in the peer-reviewed literature, and these have been cited over 5,600 times (h-index = 41). I serve on the editorial boards of *Environment International* (where I am Insights Editor) and the *Journal of Environmental Exposure Assessment*. I am a past editorial board member of the *International Journal of Environmental Research and Public Health* and the *Journal of Toxicology and Environmental Health* and past Associate Editor for the *Journal of Exposure Science and Environmental Epidemiology*. I have conducted peer review of manuscripts and reports for numerous scientific journals and governmental agencies.

My curriculum vitae is attached to this Report as Appendix 1.

I am compensated at a rate of $575 per hour for my time consulting on these matters, preparing this Report, and, if called upon to do so, providing testimony in this case. I have not previously testified as an expert witness. The Materials Considered Appendix lists all the materials I considered in the preparation of this Report.

2. CASE OVERVIEW

This Report was prepared at the request of the United States Department of Justice (US DOJ). As part of my engagement in this case, I have been asked to review materials relevant to the *Keller v. United States* case and to develop opinions regarding Mr. Keller's exposure to five chemicals in treated water used by people at Marine Corps Base Camp Lejeune, North Carolina (referred to in this Report as "Camp Lejeune" or "Base"): perchloroethylene (PCE, tetrachloroethylene, CASRN: 127-18-4), trichloroethylene (TCE, CASRN: 79-01-6), *trans*-1,2-dichloroethylene (DCE, CASRN: 156-60-5), vinyl chloride (VC, CASRN: 75-01-4), and benzene (CASRN: 71-43-2). These five chemicals are referred to in this Report as "chemicals of interest." My overall opinion is based on results from the modeling of exposures.

2.1 Summary of opinion

In this Report, I use two models to estimate Mr. Keller's past exposures to the Agency for Toxic Substances and Disease Registry's (ATSDR) modeled monthly concentration estimates of PCE, TCE, DCE, VC, and benzene in water at Camp Lejeune: one for the dermal/inhalation routes of exposure and one for the oral route of exposure (water ingestion). Based on my review and analysis of the information produced in this case, as well as my exposure and risk assessment education, training, and experience, I have formed the following opinion. My opinion herein is held to a reasonable degree of scientific certainty considering my use of ATSDR's modeled chemical concentrations in water. I reserve the right to modify or supplement my opinion if additional information is made available to me, including information from reports and testimony of other experts in this matter.

SUMMARY OF OPINION

People living and working at Camp Lejeune from the 1950's to the 1980's may have been exposed to PCE, TCE, DCE, VC, and/or benzene due to the presence of these chemicals in finished water at Camp Lejeune. Finished water is "[w]ater that has passed through a water treatment plant. All the treatment processes are completed or finished. This water is the product from the water treatment plant and is ready to be delivered to consumers" (https://owp.csus.edu/glossary/finished-water.php). In this Report, either "water" or "finished water" is used to indicate the water used in residences or for drinking water at Camp Lejeune.

Note that in this Report, I use mean monthly chemical concentration estimates modeled by ATSDR, who state that their modeled data are for finished water at Camp Lejeune (Maslia et al. 2007, 2013). In Dr. Alexandros Spiliotopoulos' Expert Report (2024, pgs. 68-69), he states that "For Hadnot Point, as with Tarawa Terrace, ATSDR assumed concentrations of contaminants in the influent to the WTP [water treatment plant] were equal to the concentrations of contaminants in the 'finished water' that was delivered to consumers...This assumption is incorrect, as treatment of the influent to the treatment plant resulted in evaporative and other losses, reducing contaminant concentrations in the 'finished' water." Based on this opinion, the concentrations of

chemicals of interest used in this Report, derived from ATSDR modeling, would be an overestimate of actual chemical concentrations in water used by people at Camp Lejeune[1].

The routes of exposure for people living and working at Camp Lejeune could have included:

- Ingestion (for example, drinking the finished water, using the water for cooking, drinking small amounts of water during swimming)
- Inhalation (breathing the chemicals that volatilized from the finished water during activities such as showering, bathing, swimming, or using appliances such as washing machines)
- Skin contact (dermal exposure from contacting the water during activities such as showering, bathing, hand washing, or swimming)

There were very few measurements made of chemicals in the water at Camp Lejeune during the overall time-period of interest (1953 -1987; https://www.navy.mil/Camp-Lejeune-Justice-Act-Claims/Claim-Eligibility/); measurements of the chemicals of interest in the water began in the 1980's (Maslia et al. 2007, 2013). However, ATSDR estimated mean monthly water concentrations for the time-period of interest (Maslia et al. 2007, 2013). The US DOJ requested that I rely on ATSDR's mean monthly chemical concentration data for estimating exposures at Camp Lejeune as these are the values reported in the Expert Report of Morris L. Maslia, P.E. (2024).

Similarly, no measurements of chemicals in indoor air at Camp Lejeune were identified for the time-period of interest. Therefore, a model that can estimate indoor air concentrations based on chemical concentrations in water was used in this Report. Note that in this Report, the modeled indoor air concentrations are from use of finished water in the residence and not from vapor intrusion; the potential presence of chemicals in indoor air from vapor intrusion is not addressed in this Report.

Finally, I did not identify detailed contemporaneous documentation related to daily behaviors and activities for people on Base decades ago. Information from various sources - including Mr. Keller's deposition transcript - was used to describe behaviors and activities leading to likely contact with chemicals in water and air.

These information sources were used in conjunction with various exposure models to estimate exposures to people at Camp Lejeune (see Section 5.1 for additional information). The exposure models used in this Report were developed by ATSDR. The underlying approaches (described in Sections 7 and 8) are well-established and have been used in assessments of ingestion, inhalation, and dermal exposures for many years by regulatory agencies, consultants,

[1] Drs. Hennet and Spiliotopoulos explain in their Expert Reports that ATSDR's modeled exposure estimates are unreliable and likely biased high as a result of several conservative assumptions used in ATSDR's modeling due to limited historical data available about the start and the extent of contaminant source releases, as well as the absence of concentration data prior to 1980 (Expert Reports of Drs. Hennet [2024, pgs. 5-35 – 5-38] and Spiliotopoulos [2024, pgs. 36-45, 70-87]).

and academicians. The models were employed to estimate ranges of possible exposures that reflect the time that Mr. Keller was on Base and his general likely behaviors and activities on Base.

Using these existing data and models, I was able to draw conclusions about Mr. Keller's likely exposures to PCE, TCE, DCE, VC, and benzene to a reasonable degree of scientific certainty, considering my use of ATSDR's modeled chemical concentrations in water, as detailed in this Report.

It is important to note that, where possible and scientifically supportable, conservative assumptions were used for determining model inputs. Conservative assumptions are those that tend to produce higher estimates of exposure. They are used to avoid underestimating exposures. In other words, conservative assumptions produce "[a]n estimate that tends to err on the side of caution or gives a 'worst case scenario'" and are "[o]ften used in risk assessment to ensure that as much risk as possible is taken into account" (https://www.efsa.europa.eu/en/glossary/conservative-assumption#:~:text=Description:,possible%20is%20taken%20into%20account).

Specific aspects of this Report that contribute to the conservative nature of the exposure estimates are described throughout the Report and summarized in Section 9.

Therefore, Mr. Keller's actual exposures are unlikely to be higher than the exposure estimates produced by these models. These exposure estimates can be used in risk assessments to determine whether people who resided at Camp Lejeune during the time-period that Mr. Keller was there, who lived in similar areas, and engaged in similar activities had an increased risk of disease (this is addressed in the Expert Report of Dr. Lisa Bailey for Scott Keller).

3. METHODOLOGY

The opinions in this Report are based on my training and experience in exposure science and on a review of documents available as of the date of this Report. Specific documents that I have reviewed are presented in the Materials Considered Appendix. In addition, there are numerous documents that I have reviewed in my professional history that are not referenced specifically, but that have supported my understanding of this case.

I have reviewed the Expert Reports of Dr. Remy Hennet (2024) and Dr. Spiliotopoulos (2024) regarding information related to groundwater, contaminant fate and transport, and water distribution modeling for Camp Lejeune, Mr. Keller's deposition transcript, and certain Military or Service Records of the Plaintiff. I have also reviewed the ATSDR's water modeling reports for Camp Lejeune and housing and other drawings for Camp Lejeune.

The specific activities I performed for my evaluation are briefly stated below:

- I reviewed Mr. Keller's deposition transcript and documents related to Mr. Keller's Military Service history (these documents are included in the Materials Considered Appendix).
- I reviewed the ATSDR's estimated monthly mean concentrations in finished water from the Hadnot Point and Tarawa Terrace water systems, specifically modeled concentrations for TCE, PCE, DCE, VC, and benzene.
- I applied an exposure science method (ATSDR's Shower and Household Water-use Exposure [SHOWER] model) to conduct an exposure assessment for dermal contact with – and inhalation of – chemicals of interest for a population using communal bathroom/shower facilities in barracks with parameters (e.g., time at Camp Lejeune, shower duration) substantially similar to Mr. Keller.
- I applied a standard exposure science method to conduct a drinking water exposure assessment for people with parameters similar to Mr. Keller (e.g., time at Camp Lejeune, drinking water consumption rates) using the ATSDR Public Health Assessment Site Tool (PHAST) for drinking water ingestion.

The following sections provide more information about methodologies for conducting exposure assessments and specifically for conducting exposure assessments for people living or working at Camp Lejeune.

4. BACKGROUND ON CHEMICAL EXPOSURE ASSESSMENT

The chemical risk assessment approach currently in use was initially put forth several decades ago (NRC 1983). The purpose was to provide a structure for estimating the possible health effects of chemical exposures to humans. Risk assessment is comprised of four basic elements as shown in **Figure 1** (US EPA 2022).

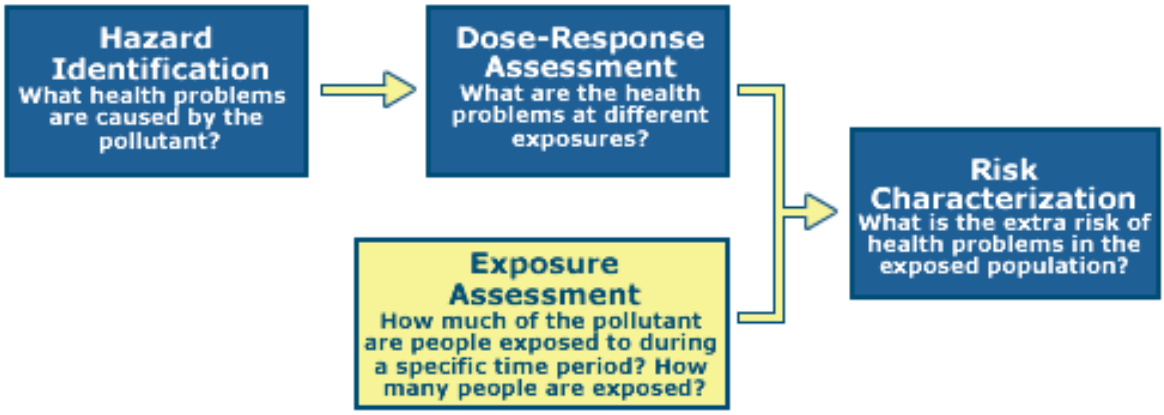

Figure 1. The 4-step risk assessment process (US EPA 2022).

One of the four basic elements is exposure assessment, defined as "[t]he process of estimating or measuring the magnitude, frequency, and duration of exposure to an agent, along with the number and characteristics of the population exposed. Ideally, it describes the sources, routes, pathways, and uncertainty in the assessment" (IPCS 2004, pg. 101).

Simply put, in conducting an exposure assessment, we seek to understand how much of a chemical people are exposed to during a specific time-period (e.g., a period of days, weeks, months, or years). When combined with information about a chemical's toxicity, the health risks associated with exposure to one or more chemicals can be assessed, or "characterized." Therefore, the assessment of human exposure is an essential component of any risk assessment.

When considering exposure to chemicals from water, three routes of exposure are generally evaluated: oral, inhalation, and dermal (**Figure 2**).

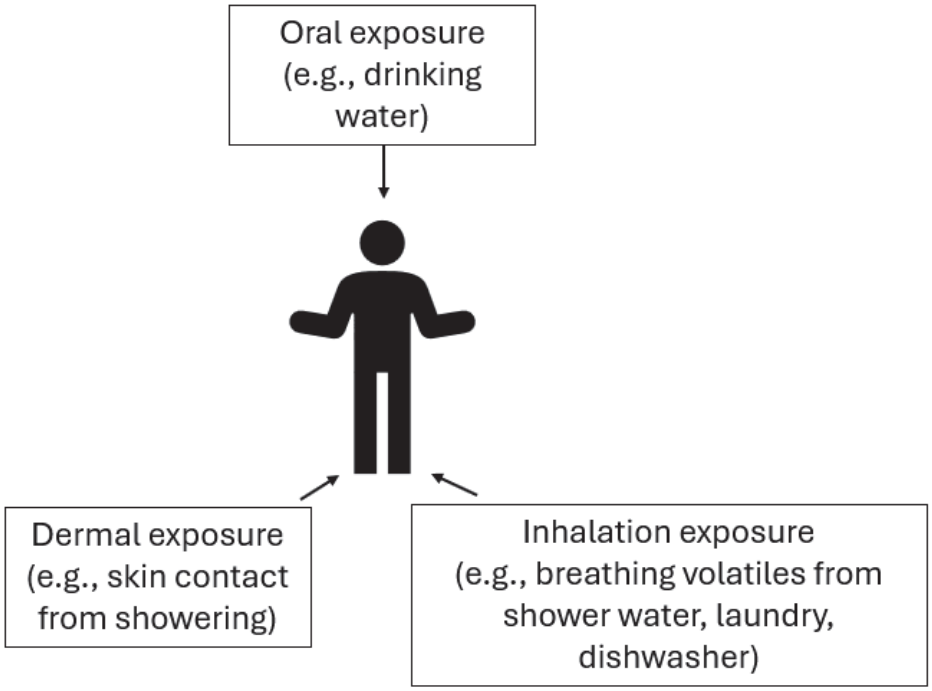

Figure 2. Routes of exposure: chemicals in water

To assess human exposures to chemicals, one needs information on chemical concentrations in environmental media such as water and air, on human behaviors, and on aspects of the environment in which people reside. These can include data on the duration of exposure (e.g., how many years a person comes into contact with the air or water), the frequency of exposure (e.g., how many days per week, hours per day), the volume of water consumed (how many liters per day), and many other factors, as well. The exposure assessor obtains site- and population-specific information where possible. When this information is not available, exposure assessors rely on information obtained from sources such as general population studies, governmental data, and scientific literature. We then make determinations regarding how to use that information to conduct site-specific exposure assessments.

The types of information described in the preceding paragraph are used as inputs to models to derive quantitative estimates of exposure. These estimates are generally expressed in units of milligram chemical per kilogram body weight per day, or mg/kg-day. The quantitative estimates describe how much of a chemical enters the body per day. A model can be a simple equation requiring at most a hand calculator or can be very complex.

In this Report, various parameters needed to estimate past human exposures to chemicals at Camp Lejeune are described and numerical values are assigned to these parameters. These parameters are more fully described in Sections 7 and 8 but can include, for example, the number of minutes spent showering each day, the size of the bathroom, and the volume of daily water consumption.

It is important to recognize that model inputs are derived from different sources and can include well-supported site-specific values, "default" values, and values based on best professional judgment. Well-supported site-specific data are generally the preferred source of information for an exposure assessment. Examples could include information collected at – or close to – the time that a plaintiff was on Base. The information could be obtained from interviews or diaries, for example, and could be related to activities such as daily shower durations or exact amounts of daily water consumption. Unfortunately, in studies of past exposures, it is often the case that these kinds of data are not available.

A standard practice for assigning a parameter value in the case of missing or limited information is to use a default value (ATSDR 2022a; Health Canada 1999; US EPA 2011, pg. 1-16). The European Food Safety Authority describes the use of default values as follows (EFSA 2012; pg. 2): "A number of assumptions and default values are usually applied at the various steps of the risk assessment process. These can…compensate for the absence of data, in which case the risk assessor may have to refer to default values to be able to perform the assessment. These default values should be scientifically justified and, where possible, be based on existing data and represent typical values for the missing parameter."

For the exposure assessments in this Report, various default values are used. These values are often based on data from the published literature for the general population (e.g., body weights; body surface area) or other representative types of data. For certain parameters, both "typical" values and more conservative (e.g., 95th percentiles) values are used in the models.

Some default values in this Report were obtained from the US EPA Exposure Factors Handbook (US EPA 2011). The Handbook "…has become a key source of exposure factor information and has served to promote consistency among risk assessments conducted by the [Environmental Protection] Agency and others. It provides a unique synthesis of exposure factor data for the US population that is unavailable in any other single source. It has been cited in numerous EPA Reports and peer-reviewed publications… The Exposure Factors Handbook has also been widely used by researchers outside the United States" (Phillips and Moya 2011, pg. 13). Most of the Exposure Factors Handbook data come from studies of the general population (e.g., the National Health and Nutrition Examination Survey [NHANES]) or from studies on sample populations that focus on specific groups (e.g., children). The Exposure Factors Handbook was reviewed internally by individuals within the US EPA and also underwent peer review by an external panel of experts. Thus, the default values from this source are scientifically well-supported and appropriate for use in exposure assessment. Default values from the Exposure Factors Handbook can also be supplemented with site-specific information, if available.

The most recent complete compilation of default exposure values is the 2011 Exposure Factors Handbook (US EPA 2011). Since that time, the US EPA has updated certain chapters and made them available online (https://www.epa.gov/expobox/about-exposure-factors-handbook).

In the absence of well-supported site- and plaintiff-specific data or default values, another approach to addressing missing or limited data is to use professional judgment. Professional judgment is an accepted aspect of risk assessment. For example, in the US EPA's Guidelines for

Carcinogen Risk Assessment (2005, pg. 2-51), EPA notes that "Choosing a descriptor [for weight of evidence for carcinogenic potential] is a matter of judgment and cannot be reduced to a formula." Further, the US EPA (1992a, pg. 92) has stated that "professional judgment comes into play in virtually every aspect of the exposure assessment process, from defining the appropriate exposure scenarios, to selecting the proper environmental fate models, to determining representative environmental conditions, etc...." As noted by the US Army Corps of Engineers (2010, pg. 1-5) "...there will be unavoidable data gaps and uncertainties where scientific and professional judgment is needed to predict or infer certain outcomes under certain scientific principles (Federal Focus Inc. 1994). The application of such judgment requires that the risk assessor provide the rationale or basis for the judgment."

Use of professional judgment is not unique to risk assessment but is used in various scientific disciplines. For example, professional judgment has been described as "one of the most important aspects of evidence-based practice" in psychology (Wilczynski 2017, pg. 65): "Good professional judgment is based on accessing all relevant information about the best available evidence and the clients (target/stakeholder/ leader) as well as the context, so the best clinical decision is made." In the field of biology, “[i]t has long been recognized that there are relatively few absolutes in biology, and that any interpretation of observed phenomena must be tempered by sound scientific judgment” (Weed 2007, pg. 138). As noted by Weed (2007, pg. 139), "science would not be science without judgment."

For the exposure assessments in this Report, values derived from professional judgment are based on a combination of (i) information derived from plaintiff depositions, (ii) military and other expert Reports, (iii) the peer-reviewed published literature, and (iv) experience and education. While the information from these sources may not be specific to the plaintiff or to Camp Lejeune, for it to be used, it should be considered relevant to one or both. Where necessary and scientifically supportable, values based on professional judgment were selected to be able to derive both typical and conservative (in other words, designed to avoid under-estimating) estimates of exposure.

In summary, exposure assessment is an essential component of risk assessment and methods for estimating human exposures to chemicals have been used by exposure and risk assessors for several decades. Despite advances in exposure assessment methods, uncertainties and limitations are an inherent part of the exposure assessment process. Exposure assessments require assumptions because site-specific information is often unavailable, and individuals may not be able to accurately recall (or may not know) exposure-related information. Further, exposure varies from day to day (e.g., shower duration, amount of water consumed, water sources and concentrations, etc.) and, in particular for retrospective assessments, data describing this variability are generally not available. Because of this, where possible and where scientifically supportable, I have chosen to utilize values and assumptions for the exposure assessment in this Report that would tend to overestimate exposure (i.e., provide conservative exposure estimates).

5. CONCEPTS AND TERMINOLOGY FOR THE EXPOSURE ASSESSMENT IN THIS REPORT

As with any scientific discipline, exposure science is replete with concepts and terminology that may be unfamiliar to those who are not experts in the field. I describe here several concepts and words/phrases that are used throughout the Report and that may be unfamiliar to the reader.

5.1 Concepts

Plaintiff activities and behaviors: The exposure assessment in this Report is intended to capture exposures experienced by people residing and/or working at Camp Lejeune during a time-period specific to the Plaintiff's actual time on Base. A necessary component of this assessment is an understanding of a plaintiff's activities and behaviors (e.g., amount of water consumed, time spent in the shower). The exposure assessment in this Report is not a perfectly accurate representation of exposure to a specific individual because Plaintiff-specific information on activities and behaviors necessary to develop such a representation is not available. For example, no contemporaneous documentation (e.g., diaries) describing day-to-day activities was identified. However, exposures can still be assessed by making assumptions derived from information from depositions, other sources of information related to the United States population, the military in general, Camp Lejeune specifically, and my best professional judgment. These various sources of information are used to gain a better understanding of data uncertainties (e.g., lack of data from the time-period of interest, uncertain recall) and variability (e.g., spatial and temporal changes in a person's activities and other factors) for the exposure parameters used in the exposure assessment.

Models: Two types of models are referenced in this Report: models used to estimate concentrations of chemical of interest in the water at Camp Lejeune and models used to estimate plaintiff exposures.

- The first type of model (i.e., models used to estimate chemical concentrations in water) is referred to as water modeling, which ATSDR describes as a "...scientific method that helps ATSDR estimate past water-system conditions that no longer exist today" (https://www.atsdr.cdc.gov/camp-lejeune/php/water-modeling/meetings-faq.html?CDC_AAref_Val=https://www.atsdr.cdc.gov/sites/lejeune/water-modeling-meetings-and-faqs.html). In this Report, I use the results from ATSDR models to describe concentrations of chemicals of interest in water from the Hadnot Point and Tarawa Terrace water systems. The US DOJ requested that I rely on ATSDR's mean monthly chemical concentration data for estimating exposures at Camp Lejeune as these are the values reported in the Expert Report of Morris L. Maslia, P.E. (2024). Details regarding water modeling are provided in a separate Expert Report by Dr. Spiliotopoulos (2024) and are not described here.
- The second type of model (i.e., models used to estimate plaintiff exposures) is central to this Report. Two different exposure models are used. These models: (i) estimate human exposures to chemicals from consumption of drinking water, and (ii) estimate

human exposures to chemicals from inhalation of volatiles from water and dermal contact with water related to using communal bathroom/shower facilities in barracks.

The basic models (i.e., equations) for estimating inhalation, dermal, and oral exposures to chemicals are well-established and have been used by various agencies, consultants, and academicians (e.g., ATSDR 2023a; Baier-Anderson et al. 2006; Chowdhury 2015; EarthCon 2019; Health Canada 2021; Huerta et al. 2023; Khan et al. 2024; Lowe and Jamall 1994; Oregon Department of Environmental Quality 2010; Ramirez-Andreotta et al. 2013; Salhotra 2011; USEPA 1989, 1992b, 2009).

For the models used to estimate human exposures to chemicals of interest, it is important to note that the estimates are for a single 24-hour period. The process of converting a one-day exposure to an estimate of long-term exposure - and the results of that process for individual plaintiffs - are described in a separate Expert Report (Expert Report of Dr. Lisa Bailey for Scott Keller).

Exposure pathways: The water at Camp Lejeune was used for a variety of purposes including drinking, use for food preparation, appliance use such as laundry and dishwashing, and showering and bathing as well as various occupational, recreational, and cleaning purposes. For use as drinking water, in this Report I consider the total amount of water that may have been consumed over the course of a 24-hour period, assuming the water source is either the Hadnot Point or Tarawa Terrace water system. For dermal and inhalation contact, I consider exposures to the chemicals of interest over a 24-hour period from using water in a bathroom (e.g., showering, sink use). I recognize that other on-Base activities could have resulted in dermal or inhalation exposures. For example, these exposures could have occurred during mess hall activities, swimming, or car washing. In the case of the use of the Base swimming pool(s), for plaintiffs who specifically noted the use of a pool, indoor air concentrations were modeled. However, based on information in Mr. Keller's deposition (Scott Keller April 15, 2024 Deposition Transcript, pg. 171), he did not use a pool during his time on Base. Therefore, I did not model swimming pool air concentrations of chemicals of interest for the time Mr. Keller was on Base. For outdoor activities such as car washing, in my professional judgment, inhalation exposures would be minimal due to dilution with the surrounding air. Dermal exposure would similarly likely be low due to off-gassing of volatile chemicals and minimal amounts of exposed skin surface area.

Many plaintiffs described cleaning activities such as "field days" in the barracks. Cleaning activities are not addressed in this Report for two reasons. First, I did not have a model that would specifically address this activity in a barracks. Second, for cleaning the barracks, based on my professional judgment, use of warm water from buckets in a large building would not provide a substantial contribution to a person's exposure compared to their showering-related exposures.

5.2 Terminology

Dose: This is the amount of a chemical that is taken into a person's body. Dose is usually estimated for a certain amount of time (for example, how much of a chemical enters the body in a

day). The amount that enters the body is also adjusted for the body weight of the person (i.e., the amount of a chemical that enters the body for each kilogram of body weight). Thus, the units to describe dose are milligram of a chemical per kilogram body weight per day, or mg/kg-day.

Extent of exposure: In the human exposure models used in this Report, there are options to assess two types of exposure: central tendency exposure (CTE) and reasonable maximum exposure (RME). These are defined by ATSDR as follows (https://www.atsdr.cdc.gov/pha-guidance/resources/ATSDR-EDG-Body-Weight-508.pdf):

> Central Tendency Exposure (CTE): CTE refers to persons who have average or typical intake factors.
>
> Reasonable Maximum Exposure (RME): RME refers to persons at the upper end of the exposure distribution (approximately the 95th percentile). The RME scenario assesses exposures that are higher than average but still within a realistic exposure range.

The model used to estimate exposure to chemicals of interest via drinking the water produces both CTE and RME results and these are included in this Report. In addition, much higher estimates of exposure from drinking water are included (higher than the RME). CTE and RME results are also provided for inhalation and dermal exposures associated with use of communal bathrooms/showers.

Intake rate: Intake rate is defined by ATSDR (https://www.atsdr.cdc.gov/pha-guidance/glossary/index.html#I_definitions) as: "The amount of a contaminated medium to which a person is exposed during a specified period of time. The amount of water, soil, and food ingested on a daily basis; the amount of air inhaled; or the amount of water or soil that a person may contact through dermal exposures are all examples of intake rates." If the medium is water, then the drinking water intake rate is expressed in units of liters per day (L/day). If the medium is air, then the air inhalation intake rate is expressed in units of cubic meters of air per day (m^3/day). Intake rates refer to the medium (e.g., air, water) as opposed to dose, which is the intake of the chemical of interest.

Exposure Factor: The Exposure Factor, or EF, is "[a]n expression of how often (frequency) and how long (duration) a person may be contacting a substance in the environment. In many instances, the exposure factor (EF) will equal 1, representing a daily exposure to the contaminant. However, some exposures may occur on an intermittent or irregular basis. For these exposures, an EF can be used to average the dose over the exposure interval" (ATSDR 2018, pg. 4). The equation for EF (unitless) is (F [frequency] x ED [exposure duration])/AT [averaging time]. In this Report, I estimate exposures for a single day, and do not consider frequency, duration, or averaging time. These parameters are addressed in a separate Expert Report (Expert Report of Dr. Lisa Bailey for Scott Keller). For a single day exposure, the parameter EF reduces to a value of 1.

Oral exposure: Oral – or ingestion - exposure occurs from consumption of contaminants in, for example, food or water. In this Report, I estimate the Plaintiff's oral exposures from ingestion of finished water.

Dermal exposure: This Report includes consideration of dermal exposure, or exposure from skin contact with the chemicals of interest in the water. The primary equations for estimating dermal exposure are provided in a later chapter of this Report. These equations are more complex than the equations for exposure via water ingestion or for inhalation of volatiles from the air. This is because dermal exposure assessment requires information not only on the amount of skin contact that occurs, but also on the extent to which the chemical is absorbed by the skin. The reader is referred to the references in the relevant chapters in this Report for information on additional equations and equation parameters. For dermal exposure, the dose is described as the dermally absorbed dose, or the dose of the chemical absorbed through the skin and into the body (ATSDR 2023a). This dose can be converted to what is referred to as an "administered dose." For the chemicals of interest in this Report, the dermally absorbed dose and the administered dose are equivalent. As stated by ATSDR (2023a, pg. 7): "For most chemicals, the absorbed dermal dose is the same as the oral administered dose because we assume 100% of the chemical is absorbed through the GI tract, thus [the gastrointestinal absorption factor] equals 1. Therefore, no adjustment from absorbed dermal dose to administered oral dose is needed for VOCs [volatile organic compounds], SVOCs [semi-volatile organic compounds], pesticides, PAHs [polycyclic aromatic hydrocarbons], and PCBs [polychlorinated biphenyls]. For these chemicals the absorbed dose calculated from dermal uptake is also an administered dose."

Inhalation exposure associated with finished water: While the available ATSDR estimated monthly mean chemical concentration data are for water, the chemicals of interest are volatile, meaning that they can evaporate from the water and enter the air. Therefore, this Report includes an assessment of inhalation of air containing chemicals that have volatilized from the water. The concentrations in air are modeled with approaches described in later sections of this Report.

6. CHEMICAL CONCENTRATION INFORMATION FOR CAMP LEJEUNE

In the following sections of this Report, I describe two models that I used to estimate Mr. Keller's past exposures to PCE, TCE, DCE, VC, and benzene in water at Camp Lejeune: one for the dermal/inhalation routes of exposure (SHOWER model, Section 7) and one for the oral route of exposure (PHAST, Section 8). In Sections 7 and 8, I describe the models themselves as well as the available information used to select values for the model parameters. Finally, I describe the results from each of these models. Where Plaintiff-specific information was available, this is shown in **bold font**.

The exposure models in this Report require information on concentrations of the chemicals of interest in water. In the following sections, I describe the sources of the water concentration data at Camp Lejeune (Section 6.1) and the water concentrations of PCE, TCE, DCE, VC, and benzene used in this Report (Section 6.2).

6.1 Background on available chemical concentration data for water at Camp Lejeune

Chemical concentrations in water (and in air from volatilization of chemicals from water to air) can be determined from measuring those chemicals in samples of the water. In the case of past exposures for which few or no measurements of chemicals were obtained, models can be used to estimate water concentrations. Modeling the chemical concentrations in water is often the only approach that can yield the information needed to conduct an exposure assessment.

There are a limited number of historical measurements of PCE, TCE, DCE, VC, and benzene in the water in the impacted areas of Camp Lejeune (Maslia et al. 2016) and these measurements were not made until the 1980's. Reconstructions (or modeling) of estimated mean monthly water concentrations of these chemicals were done by ATSDR. ATSDR modeled monthly average concentrations of PCE, TCE, DCE, VC, and benzene for the years of interest (1953-1987). They provided the results in publicly available reports (Maslia et al. 2007, 2013). These reports include modeled monthly mean concentrations of the chemicals of interest in the areas of Camp Lejeune served by the Tarawa Terrace and Hadnot Point water systems (the water systems that are the focus of this Report).

ATSDR reconstructed monthly mean concentration values (Maslia et al. 2016) for finished water from January 1952 to May 1996 for Hadnot Point (Maslia et al. 2013) and from January 1952 to February 1987 for Tarawa Terrace (Maslia et al. 2007). I relied on estimated mean monthly concentrations of PCE, TCE, DCE, VC, and benzene (benzene for Hadnot Point only) in water for Hadnot Point[2] and Tarawa Terrace extracted and compiled into Excel spreadsheets by S.S. Papadopulos & Associates, Inc. It is my understanding that the data were extracted from the ATSDR Reports (Maslia et al. 2013, Appendix A7 and Maslia et al. 2007, Appendix A2, respectively). These compiled data were used as the basis for the analyses in this Report. Reconstructed concentration minima for all chemicals were equal to 0 µg/L (micrograms per liter). While chemical concentrations in the water could have varied from day to day, only monthly

[2] The Maslia et al. (2013) report refers to this as the Hadnot Point–Holcomb Boulevard study area. For detailed information on the locations of interest, see the Expert Report by Dr. Spiliotopoulos (2024).

average modeled concentrations were available; these were used as the basis for determining overall average water concentrations for the time the Plaintiff spent on Base.

According to the Expert Report of Dr. Spiliotopoulos (pgs. 68-69): "For Hadnot Point, as with Tarawa Terrace, ATSDR assumed concentrations of contaminants in the influent to the WTP were equal to the concentrations of contaminants in the 'finished water' that was delivered to consumers...This assumption is incorrect, as treatment of the influent to the treatment plant resulted in evaporative and other losses, reducing contaminant concentrations in the 'finished' water." Further, according to the Expert Report of Dr. Hennet (pg. 5-43): "A substantial portion of [contaminants of concern] that may have been present in the treated water used to fill a water buffalo would have unavoidably been lost to evaporation during filling, use, and variations of temperature. These [contaminants of concern] reductions between the raw water and the water in the water buffaloes would have been in the order of 52% to 73% based on my estimation." *Based on the information in these Expert Reports, the ATSDR concentrations described in this Report, as well as the associated estimates of Plaintiff exposure, would be overly conservative (too high).*

6.2 Water concentration data relevant to Mr. Keller

Assumptions for assessment of plaintiff-specific time on Base are:

(i) If a plaintiff was on Base for part of the calendar month, I assumed that the plaintiff was there for the entire month (the exception to this was if the plaintiff was only on Base for one day for that month).

(ii) Plaintiffs may have been off-Base for part of their time assigned to Camp Lejeune (e.g., leave, weekends away, time spent on parts of the Base where water was not impacted). Unless they were off Base for at least one calendar month (e.g., January 1 to January 31) and the exact dates were known, it was assumed that they were on Base and exposed to the chemicals of interest for the entire time-period.

According to Mr. Keller's Military Records, he arrived at Camp Lejeune in November 1985 (01501_KELLER_NARA_0000000041). He lived and worked on Base until January 1987, at which time he was deployed to the Mediterranean (01501_KELLER_NARA_0000000073). He returned to Base in July 1987 and lived there until September 1987, at which time he moved to Jacksonville, NC to be with his family (Scott Keller April 15, 2024 Deposition Transcript, pg. 45). However, he continued to work and shower on Base until January 1988 (Scott Keller April 15, 2024 Deposition Transcript, pgs. 46-47, 158). Mr. Keller did not specify where on Base he showered during this last time-period; I assumed that he continued to shower in locations served by water from the Hadnot Point water system. He then moved to Palms, California (01501_KELLER_NARA_0000000041). While he was on Base until January 1988, the end of the statutory period for this case is December 1987. Therefore, for the purposes of the exposure assessment in this Report, the dates for which Mr. Keller lived and/or

worked/showered on Base are: November 1985 - January 1987 and July 1987 – December 1987.

Water source(s) and time-periods for residential (barracks) dermal and inhalation exposure: **Mr. Keller lived in the Mainside barracks at Hadnot Point and presumably showered at Hadnot Point during his time at Camp Lejeune (Scott Keller April 15, 2024 Deposition Transcript, pgs. 17, 149).** I estimated his dermal and inhalation exposures for the time-periods November 1985 - January 1987 and July 1987 – December 1987 using the Hadnot Point water source.

Water source(s) and time-periods for exposure via water ingestion: I assumed that Mr. Keller could have spent time on parts of the Base consuming water from either the Hadnot Point or Tarawa Terrace water systems. I therefore modeled his exposure via drinking water consumption using data from both the Hadnot Point and Tarawa Terrace water systems for the time-periods November 1985 - January 1987 and July 1987 – December 1987.

The monthly mean modeled values for Hadnot Point and Tarawa Terrace used for estimating the overall mean water concentrations for Mr. Keller are shown in **Table 1**. To estimate exposures for those – including Mr. Keller - at Camp Lejeune during this time-period, the overall mean value for each chemical at each location is used (**Table 2**). This is consistent with ATSDR's use of a three-year rolling average for estimating exposures in its Camp Lejeune Public Health Assessment (ATSDR 2017a). Estimation of the average dose is also consistent with the risk assessment paradigm that includes the use of an Average Daily Dose (ADD) or Lifetime Average Daily Dose (LADD) (US EPA 1992a). Further, the US DOJ requested that I rely on ATSDR's mean monthly chemical concentration data for estimating exposures at Camp Lejeune as these are the values reported in the Expert Report of Morris L. Maslia, P.E. (2024).

Table 1. Monthly mean modeled water concentrations (µg/L) of PCE, TCE, DCE, VC, and benzene at Hadnot Point and PCE, DCE, TCE, and VC at Tarawa Terrace from November 1985 - January 1987 and July 1987 – December 1987.

Hadnot Point	**Water concentrations (µg/L)**				
Month/Year	**PCE**	**TCE**	**DCE**	**VC**	**Benzene**
Nov-85	0	0	0	0	3
Dec-85	0	0	0	0	3
Jan-86	0	0	0	0	3
Feb-86	0	0	0	0	3
Mar-86	0	0	0	0	3
Apr-86	0	0	0	0	4
May-86	0	0	0	0	3

Jun-86	0	0	0	0	3
Jul-86	0	0	0	0	3
Aug-86	0	0	0	0	3
Sep-86	0	0	0	0	3
Oct-86	0	0	0	0	3
Nov-86	0	0	0	0	3
Dec-86	0	0	0	0	3
Jan-87	0	0	0	0	2
Jul-87	0	0	0	0	3
Aug-87	0	0	0	0	3
Sep-87	0	0	0	0	3
Oct-87	0	0	0	0	3
Nov-87	0	0	0	0	2
Dec-87	0	0	0	0	2

Tarawa Terrace*	**Water concentrations (µg/L)**			
Month/Year	**PCE**	**DCE**	**TCE**	**VC**
Nov 1985	3.24	0.87	0.15	0.71
Dec 1985	3.58	0.95	0.16	0.76
Jan 1986	3.95	1.04	0.18	0.82
Feb 1986	4.24	1.08	0.19	0.83
Mar 1986	5.40	1.34	0.24	1.01
Apr 1986	4.93	1.20	0.22	0.89
May 1986	5.25	1.25	0.23	0.91
June 1986	5.61	1.30	0.25	0.92
July 1986	5.97	1.35	0.26	0.94
Aug 1986	6.36	1.39	0.28	0.96
Sept 1986	6.75	1.44	0.30	0.97
Oct 1986	7.12	1.48	0.31	0.99
Nov 1986	7.52	1.52	0.33	1.00
Dec 1986	7.89	1.56	0.34	1.01
Jan 1987	8.28	1.59	0.36	1.01

*Benzene was not included for Tarawa Terrace as it was not included in the modeled water results (Maslia et al. 2007). The final month with water concentration data for Tarawa Terrace is February 1987 due to the closure of the Tarawa Terrace water treatment plant in March 1987 (Maslia et al. 2007).

Table 2. Overall mean concentrations (µg/L) of PCE, TCE, DCE, VC, and benzene at Hadnot Point and Tarawa Terrace over the time-periods November 1985 - January 1987 and July 1987 – December 1987. These concentration data were used to estimate chemical exposures in this Report. (The final month with water concentration data for Tarawa Terrace is February 1987 due to the closure of the Tarawa Terrace water treatment plant in March 1987 [Maslia et al. 2007]).

Hadnot Point (µg/L)					Tarawa Terrace (µg/L)			
PCE	**TCE**	**DCE**	**VC**	**Benzene**	**PCE***	**DCE**	**TCE**	**VC**
0.0	0.0	0.0	0.0	2.9	5.7	1.3	0.3	0.9

*The Tarawa Terrace value for PCE is based on the results using the TechFlowMP model. The modeled values using the TechFlowMP model are lower than those generated using the MT3DMS model; the reasons for this are given in Jang and Aral (2008), pg. G-14. Because TCE, VC, and DCE were modeled using the TechFlowMP model only, for consistency, values for all four chemicals at Tarawa Terrace generated with that model are used in this Report.

A description of the uncertainties in the ATSDR mean monthly concentration data is outside of the scope of this Report, but information is available on this topic in the Expert Reports by Dr. Hennet (2024) and Dr. Spiliotopoulos (2024).

According to the Expert Report of Dr. Spiliotopoulos (pgs. 68-69): "For Hadnot Point, as with Tarawa Terrace, ATSDR assumed concentrations of contaminants in the influent to the WTP were equal to the concentrations of contaminants in the 'finished water' that was delivered to consumers...This assumption is incorrect, as treatment of the influent to the treatment plant resulted in evaporative and other losses, reducing contaminant concentrations in the 'finished' water." Further, according to the Expert Report of Dr. Hennet (pg. 5-43): "A substantial portion of [contaminants of concern] that may have been present in the treated water used to fill a water buffalo would have unavoidably been lost to evaporation during filling, use, and variations of temperature. These [contaminants of concern] reductions between the raw water and the water in the water buffaloes would have been in the order of 52% to 73% based on my estimation." *Based on these expert opinions, the chemical concentrations used in this Report as well as the associated estimates of Plaintiff exposure would be overly conservative (too high).*

7. DERMAL AND INHALATION EXPOSURE – THE SHOWER MODEL

7.1 SHOWER model: Background

Due to the volatile nature of PCE, TCE, DCE, VC, and benzene, inhalation of these chemicals deriving from water can occur during showering and bathing and via the use of appliances that use water (e.g., washing machines, dishwashers). In addition, dermal exposure to PCE, TCE, DCE, VC, and benzene can occur during showering and bathing or during faucet use.

While the basic models for estimating inhalation and dermal exposures to chemicals are well-established, addressing the time-varying concentrations of these chemicals in a residence is complex. ATSDR's SHOWER model addresses this complexity. The model, first released (version 1.0) in May 2018, includes the basic components of the models described by the US EPA (US EPA 1989) and used for decades but adds model components that allow for a rapid evaluation of inhalation and dermal exposures from volatile chemicals in household water (ATSDR 2022b). It was developed with the following objectives (list taken directly from ATSDR 2022c):

- providing an easy-to-navigate platform that requires minimal input to obtain results,
- providing standardized scenarios based on characteristic parameter values,
- allowing users to develop custom scenarios for site-specific simulations, and
- allowing users to evaluate the effects of changing model parameters on model outputs.
- simulating the most common water sources that contribute to indoor exposure,
- evaluating exposure from water use in bathrooms and the main house in addition to exposures from showering,
- evaluating exposure throughout the day and night,
- accounting for non-exposure when persons are away from the home, and
- accounting for exposure contributions from water use by all household members.

To run the SHOWER model, the user needs a chemical name and chemical concentration in the finished water to obtain estimates of household daily air concentrations, dermal doses, and inhalation doses (ATSDR 2022b). Since the release of version 1.0, ATSDR has developed additional versions that give the model user more flexibility in terms of the behaviors of people in the household (e.g., number and timing of showers) and the layout of the modeled residence (e.g., number of bathrooms, size of the house, aspects of appliances in the residence) and improve the underlying model equations (ATSDR 2022c).

SHOWER model v2.0, released in February 2020, had several changes including the ability to evaluate the sensitivity of simulation results to changes in model parameters. It also expanded upon the functionality of the first model by, for example, allowing the user to customize several model parameters (e.g., number of bathrooms, activity sequence and duration for each household member, size and layout of the house, and household appliance parameters) (ATSDR 2022c).

Version 3.0, released in May 2022 (ATSDR 2022c), allowed estimation of both central tendency (or "typical") exposure (CTE) and reasonable maximum exposure (RME) results for households

with 1, 2, 3, and 4 persons. In addition, the effects of contaminant saturation in air were incorporated into the SHOWER model's governing equations and an export function was added permitting the import and analysis of SHOWER model data within PHAST.

Version 4.0.0 (v4) was released 26 September 2024. The main change is that v4 "...adds the ability to simulate inhalation and dermal exposures from contaminated water in public showers and bathrooms" (e-mail from PHAST, CDC, 26 September 2024). Specifically, it "...includes default exposure scenarios for gyms, offices, schools, daycares, and dorms or barracks" (e-mail, 26 September 2024, David Mellard, ATSDR/OAD/OCDAPS). There was also "...a change in the equations for calculating chemical volatility" resulting in higher inhalation concentrations compared to Version 3.0, although for volatile chemicals the difference is considered by ATSDR to be minimal (< 5%) (e-mail, 26 September 2024, David Mellard, ATSDR/OAD/OCDAPS). Version 4.0.1 (v4.0.1) was released on 19 November 2024. In terms of relevance to the modeling for this Report, the new version produces model reports with more information on facility visits, corrects a report bug in v4 regarding peak times, and includes revised algorithms for assigning activity patterns in the public shower and bathroom scenarios and for determining the number of people who shower in scenarios with small facilities where only a low percentage of people take showers (e-mail, 19 November 2024, David Mellard, ATSDR/OAD/OCDAPS).

I note here that other models that estimate indoor air concentrations have various limitations regarding their utility for assessing indoor air human exposures at Camp Lejeune. For example, various models have been developed to estimate indoor air concentrations but do not include a component that estimates human exposures (NRC 1981). Also, some available models were designed for a different purpose (e.g., assessing the effect of use of indoor stoves on air quality [WHO 2020]).

The ATSDR model relies on standard inhalation and dermal exposure equations used by exposure scientists for many years, estimates time-varying indoor air/water concentrations and human exposures, and allows for various modifications to better represent site-specific features of the indoor environmental and human behaviors.

7.2 SHOWER model: General methodology

In this Report, SHOWER Model v4.0.1 was used to estimate human exposures to PCE, TCE, DCE, VC, and benzene at Camp Lejeune via inhalation and dermal contact associated with the use of water in communal bathroom facilities in a barracks. Facilities are defined in the model as "...either a single communal shower area with adjoining locker room and bathroom area or a single communal bathroom area consisting of toilets and sinks" (ATSDR 2024a, pg. 32). Detailed descriptions of the model algorithms and parameters were given for SHOWER Model version 3.0 in ATSDR (2022c), which was specific to exposure in house/apartment-like residences. A technical document for SHOWER v4.0.1 was not available at the time of the preparation of this Report. I assume that the basic algorithms for estimating inhalation and dermal doses used in version 3.0 are also used in v4.0.1. This can be confirmed once the technical document is released. The information in this section describes those algorithms.

Inhalation route of exposure: The SHOWER model predicts daily inhalation doses derived from age-specific breathing rates and the average daily exposure concentrations calculated for each person in each Monte Carol simulation (see Reports – edited for length - in Attachment 1). Based on information from the SHOWER model version 3.0 (ATSDR 2022c), inhalation doses are estimated as follows:

$$ID=(C\times IR\times EF)/BW$$

Where:
ID = inhalation dose (µg/kg/day)
C = contaminant air concentration (µg/m^3)
IR = intake rate (m^3/day)
BW = body weight (kg)
EF = exposure factor (equal to 1)

Dermal route of exposure: The SHOWER model version 3.0 uses the following equation to estimate dermal dose (ATSDR 2022c; US EPA 2004):

$$DAD=(DA_{event}\times SA\times EV\times EF)/BW$$

Where:
DAD = dermal absorbed dose (µg/kg/day)
DA_{event} = absorbed dose per event (µg/cm^2/event)
SA = skin surface area available for contact with water (cm^2)
BW = body weight (kg)
EV = event frequency (events/day)
EF = exposure factor

The default exposure factor in the SHOWER model version 3.0 is set to 1 because the model assumes that the activities leading to exposure (e.g., showering, handwashing) occur daily. For organic compounds such as the chemicals of interest at Camp Lejeune, the equation used to estimate DA_{event} includes the chemical concentration in the water and other factors and depends on the time required for the chemical to reach steady state when passing through the skin compared to the duration of the human activity.[3]

A single run of the SHOWER model represents modeled exposure for a single 24-hour day.

ATSDR has noted that an uncertainty related to modeling the dermal permeability coefficient for certain halogenated chemicals, including PCE, TCE, DCE, and VC, can result in an underestimate of dermal doses (ATSDR 2022c, pg. B3). I did not identify any information on the extent of underestimation.

[3] Detailed information on the equations and parameters used to estimate DA_{event} can be found in ATSDR 2022c.

7.3 SHOWER Model v4.0.1: Communal facilities

In this Report, exposures to chemicals of interest in the water via dermal and inhalation exposure are based on Mr. Keller's water use while living in the barracks as well as showering on Base during the period when he lived off Base. **Mr. Keller lived in the Mainside barracks at Hadnot Point and continued to shower on Base during the time he lived off Base (Scott Keller April 15, 2024 Deposition Transcript, pgs. 17, 137, 149).** The SHOWER model (v4.0.1) for estimating exposures associated with the use of communal facilities (showers and locker room or bathrooms) relies on Monte Carlo methods[4] to "...randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure" (see Model report in Attachment 1).

The communal facilities include commercial gyms (non-school), commercial daycares (non-residential), dorms or barracks, offices, or schools (with an option for "Other building type"). The facility type can be either a shower and locker room or a bathroom with no shower. For this Report, the former facility type is considered. **Figure 3** shows an example facility. The shower area includes all showers and connecting spaces. The bathroom/locker room area includes the toilets, sinks, and connecting areas. Note that there is also a "main building" in the SHOWER model v4.0.1 which in this case would represent the part of the living quarters of the barracks not including the shower/bathroom area. In the model, this part of the building does not contribute to people's exposures because "[t]he default parameters assume that the facility operates under negative pressure when it is open and that air does not flow out from the facility (e.g., the locker room/showers) into the main building. Flow of air out from the facility into the main building may occur if you change the air exchange rates and airflow parameters from their defaults, but in that case the model still does not simulate the variation in the main building concentration over time, since it's treated as a constant" (e-mail, David Mellard, ATSDR/ OAD/OCDAPS, 30 September 2024).

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[4] "Monte Carlo simulation is a type of simulation that relies on repeated random sampling and statistical analysis to compute the results. This method of simulation is very closely related to random experiments, experiments for which the specific result is not known in advance. In this context, Monte Carlo simulation can be considered as a methodical way of doing so-called what-if analysis... In Monte Carlo simulation, we identify a statistical distribution which we can use as the source for each of the input parameters. Then, we draw random samples from each distribution, which then represent the values of the input variables. For each set of input parameters, we get a set of output parameters. The value of each output parameter is one particular outcome scenario in the simulation run. We collect such output values from a number of simulation runs. Finally, we perform statistical analysis on the values of the output parameters..." (Raychaudhuri 2008).

Figure 3. Layout of an example public bathroom facility (not to scale) (ATSDR SHOWER Model v4.0.1).

The SHOWER model default value for the percent of facility users who use the showers depends on the building type. The CTE and RME inhalation and dermal values included in this Report are for those who shower (exposures are lower for those using the facility but not the shower). The inhalation values represent daily exposure concentrations from exposures while in the building. The dermal values are daily doses from showering and hand washing in the facility.

7.4 SHOWER model – Communal facilities: Parameters and model inputs

The SHOWER model requires information on parameters that describe population characteristics and features related to a communal facilities environment. The SHOWER model includes numerous default values that are applied to the model parameters, unless the user specifies that alternative values are to be used. The default communal facilities scenarios require that the model user provide information on the following parameters: type of the building (e.g., gym, barracks, dorms), facility type (showers and locker room, bathroom with no showers), number of people using the facility, and the chemical name and water concentration. Custom scenarios allow for the model user to modify values for various other parameters.

The default values or plaintiff-specific values used in this Report for key parameters are described here. As noted above, default values are often used when site- and situation-specific information is not available. Appendix 2 shows the SHOWER model parameters and options for modifications.

7.4.1 Population-based parameters

Number of people using the facility: The model can simulate up to 1,000 people. For the depositions that I reviewed, the number of people living in the barracks ranged from approximately 30 to over 100 (although many of the depositions did not include information on this parameter).

Mr. Keller recalled that there were 40 people in his squad bay (Scott Keller April 15, 2024 Deposition Transcript, pg. 168).

In the SHOWER Model, as the number of people using the communal facility increases, the default number of showers, toilets, and sinks increases, with subsequent increases in room size and ventilation rates. This, in turn, results in lower exposure concentrations (e-mail, David Mellard, ATSDR/OAD/OCDAPS, 2 October 2024). Because of this, a larger number of people using the facility corresponds with lower modeled exposures. This can be observed in **Figure 4**, which shows the results from a hypothetical barracks (water PCE concentration equal to 10 µg/L and default values for all other parameters) with varying numbers of people using the showers and locker room (see supporting information in Appendix 3.1; SHOWER v4).

It is not clear from Mr. Keller's deposition transcript whether the communal bathroom facility served only one squad bay. To provide both reasonable (based on Mr. Keller's deposition) and conservative (protective) exposure estimates, for this Report I assume that there were 40 people living in the barracks and using the communal facility.

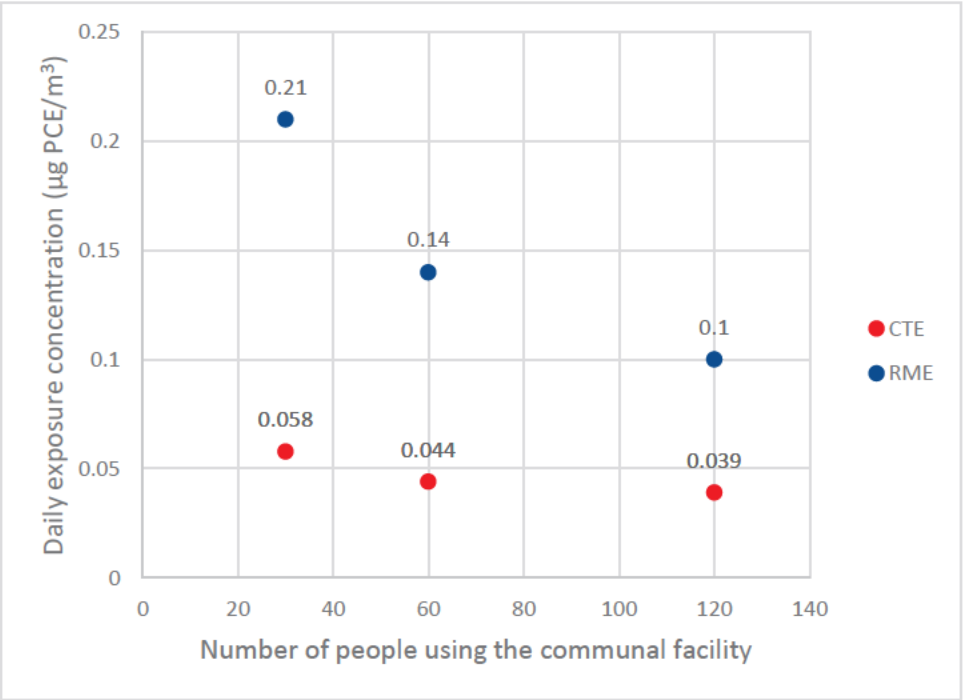

Figure 4. The relationship between the number of people in a barracks using the communal facility and daily air concentrations (see Appendix 3.1 for supporting information).

Peak shower times: The SHOWER model includes two scenarios regarding timing of shower use: (i) people in the barracks use the facility at a constant rate over the course of the day, or (ii) there are one or more peak times for shower use. It is not unreasonable to assume that early morning might be a peak time for shower/bathroom usage. The assumption of a 1-hr morning or evening peak period for bathroom/shower usage yields more conservative (higher) results compared to a constant usage scenario (see results for a hypothetical scenario in **Table 3**). The daily exposure concentrations are the same regardless of whether the peak occurred in the morning or evening. For this Report, I used the more conservative assumption that the facility utilized by Mr. Keller experienced a peak usage period from 6:00-7:00 am.

Table 3. Constant versus peak usage of bathroom/shower facility and daily exposure concentration. The results are based on a hypothetical barracks facility with a PCE water concentration of 10 µg/L, 30 people using the facility during the day, and a shower duration (standard deviation) of 20 minutes (5.3 minutes). All other parameters are set to default values. See Appendix 3.2 for supporting information (SHOWER v4).

Peak time(s)	**Daily exposure concentration (µg/m³)**	
	CTE	**RME**
None	0.25	0.56
6:00-7:00 am	0.28	0.62
6:00-7:00 pm	0.28	0.62

CTE=central tendency exposure; RME=reasonable maximum exposure

Many plaintiffs recalled that while there may have been peak shower times, the barracks showers were often used throughout the day. The SHOWER model accounts for this. As described in the model, when assuming a peak period for shower use (as was done in this Report), while modeled people arrive more often during the peak than during the rest of the barrack's operating hours, the rate of arrival increases linearly up until the peak time, stays constant during the peak period, and then decreases linearly afterwards.

Percent of people showering: The model default for a barracks sets the percentage of people who use the facility and who also use the showers to 100%. This value was utilized for this Report.

Shower duration and timing: The SHOWER model default value for shower duration is 8.0 minutes, with a standard deviation of 5.3 minutes. The model uses a lognormal distribution to determine the time that each simulated person spends in the shower. With these default values, the 95th percentile shower duration is 18 minutes for all the people that shower.

Mr. Keller reported showering one to two times per day for 5-10 minutes (Scott Keller April 15, 2024 Deposition Transcript, pg. 168). A mean 20-minute shower duration was used in this Report. Note that this value represents the average shower duration for those in Mr. Keller's barracks. For context, the average reported shower duration for North America is 7.8 minutes per shower (DeOreo et al. 2016). The recommended mean values for showering time for adults in the US EPA Exposure Factors Handbook (2011; Table 16-1) are 20 minutes for 16 - < 21-year-olds and 17 minutes for adults 18 - < 65 years of age.

Regarding the number of showers, in this Report, Mr. Keller reported showering one to two times per day. The model does not accommodate an individual showering multiple times. I therefore assumed that he took one longer shower. The high end of his total shower number and duration estimates (two showers, 10 minutes each) results in a total shower duration of 20 minutes, which I used as the mean shower duration in the SHOWER model.

Many plaintiffs also described time in bathrooms spent on personal hygiene activities other than showering (e.g., shaving). The SHOWER model includes time spent in the bathroom in addition to showering. For example, for a barracks facility used by 30 people and a mean shower duration of 20 minutes, the mean time in the shower and locker room area is 54 minutes.

7.4.2 Facility parameters

Building type: In this Report, I used the SHOWER model option for barracks with showers and locker room, toilets, and sinks.

Chemical concentrations in the facility water: The mean chemical concentrations for the chemicals of interest (**Table 2**) in Hadnot Point finished water were used to model Mr. Keller's facility-related exposures.

Facility configuration and size: I was provided a floor plan of a barracks and bathroom area in a marine barracks at New River, NC (CLJA_USMC_PWD_0000174259). This floor plan shows a bathroom divided into three sections, with five toilets and six urinals in one area, 12 sinks in a second area, and six shower heads in a third area. **(This can be compared to Mr. Keller's recollection of the number of appliances in the barracks facility: 6-8 sinks in the bathroom [Scott Keller April 15, 2024 Deposition Transcript, pg. 169] and sufficient showerheads to accommodate 10-12 people at a time [Scott Keller April 15, 2024 Deposition Transcript, pg. 171]).** The dimensions of the toilet room (toilets plus urinals), the washroom (sink area) and the shower room appear to be 16 ft x 12 ft 2 in, 16 ft x 12 ft 1 in, and 16 ft x 12 ft 2 in, respectively. Assuming a ceiling height of 8 ft, the volume in cubic inches for the three sections would be: toilet room – 2,691,072 in^3, washroom – 2,672,640 in^3, shower room – 2,691,072 in^3. In the SHOWER model, the toilet and sink areas are combined, so the total volume is 5,363,712 in^3. In cubic meters, these volumes are: bathroom – 87.9 m^3, shower room – 44.1 m^3.

These values can be used as inputs into the SHOWER model. However, the model also requires inputs for airflow (outdoor air supply flow rate, bathroom exhaust fan flow rate, and shower exhaust fan flow rate) in the facility. As noted previously, with an increasing number of people assumed to be using the facility, the numbers of appliances and the volume of the facility increase, as do the airflow values. By increasing the size of the facility using the New River floor plan dimensions and not increasing the airflow, the model may produce unrealistic air concentrations of the chemicals of interest. However, while information on facility size and appliance numbers was obtained for a New River barracks, I did not identify information on airflow rates used in the barracks at the time that plaintiffs were at Camp Lejeune and so have no supportable basis for selecting airflows for the barracks shown in the floor plan.

To determine the effect of using default versus site-specific values for the model parameters for the configuration of the facility, I compared hypothetical daily exposure concentrations using each of these two approaches. First, I used the SHOWER model (v4) default values for facility size, number of appliances, and airflow that are associated with 30 people using the facility. As noted by ATSDR (2024a, pgs. 41-42): "The default numbers of each appliance and the default area volumes are determined by the number of facility users. The program assumes an initial number of showers, toilets, and bathroom sinks based on workplace standards and building codes, and the area volumes reflect the space required to accommodate those showers, toilets, and bathroom sinks. The default air exchange rates are determined by the building type and come from the US EPA's Exposure Factors Handbook (US EPA 2011)." Second, I used site-specific values for facility size and number of appliances based on the New River barracks floor plan (CLJA_USMC_PWD_0000174259) (which appears to accommodate 30 people per shower/bathroom area) but retained the default air flow values. The input values for these two approaches are shown in **Table 4**.

Table 4. Input values for parameters associated with facility configuration and size used in this Report. Thirty people are assumed to use the facility, with 100% showering.

Parameter	SHOWER model vs 4 default values	Site-specific values
Showers (N)	3	6
Toilets (N)	2	11
Sinks (N)	1	12
Total shower room volume (m^3)	14.63	44.1
Total bathroom volume (m^3)	19.03	87.9
Outdoor air supply flow rate (L/min)	993.9	993.9
Bathroom exhaust fan flow rate (L/min)	3000	3000
Shower exhaust fan flow rate (L/min)	1800	1800

L/min=liters per minute; m^3=cubic meter; N=number

Using a hypothetical example with a PCE water concentration of 10 µg/L, the CTE and RME air concentrations with the default model inputs are: 0.058 and 0.21 µg/m^3, respectively. The CTE and RME air concentrations for the site-specific model inputs are: 0.040 and 0.14 µg/m^3, respectively (see Appendix 3.3 for supporting information). Thus, the default input values yielded more conservative results.

Given the lack of information on actual site-specific airflows, or on specific dimensions of the communal facilities used by Mr. Keller, I utilized the approach of relying on default model inputs associated with a specific number of people (40) using the facility.

<u>Outdoor air concentration and air concentration in the barracks</u>: I am unaware of outdoor air or barracks air concentration data for the chemicals of interest specific to Camp Lejeune in the time-period during which Mr. Keller was on Base. Therefore, I used the model default value of 0 µg/m^3.

7.4.3 Residence versus occupational (office) bathroom exposure

I did not identify information for Mr. Keller regarding his use of bathrooms during his days outside of the barracks. While he recalled spending days or weeks doing overnight field training at least twice a month (Scott Keller April 15, 2024 Deposition Transcript, pg. 67), his primary work was in the communications building (Scott Keller April 15, 2024 Deposition Transcript, pg. 44) and he also worked in the barracks (Scott Keller April 15, 2024 Deposition Transcript, pg. 47) (the SHOWER model accounts for more than one visit to the barracks bathroom in a given day). For his time in the communications building, I did not identify any information regarding how many people he worked with or on the configuration of the building in which he worked.

Based on results from SHOWER vs 4 modeling using default values for model inputs, exposures related to the use of office bathrooms are substantially lower than barracks facilities exposures and do not markedly contribute to overall exposures. Results from a hypothetical example are shown in **Table 5**. I compare typical exposures (CTE) for a person living in a 30-person barracks using a communal facility versus working in an office building with nine other colleagues where the water concentration of PCE is 10 µg/L in both locations (see Appendix 3.4 for supporting information). The exposure from the office setting is about 2.7% of the exposure from the barracks facilities. For this reason, as well as because I found no information on values for model inputs needed to describe Mr. Keller's office, I focus here on residential (barracks) exposures and do not estimate exposures from occupational settings (for example, the use of an office bathroom during the day).

Table 5. Comparison of CTE inhalation doses in a hypothetical residence (30-person barracks) and a 10-person office workplace (bathroom only, no shower). PCE water concentration is set equal to 10 µg/L in both locations. All other model values are set to defaults.

Building type	CTE inhalation dose (µg/kg-day)
30-person barracks	0.056
10-person office*	0.0015

*assumes average bathroom visits per person = 2

7.5 Opinion: Dermal and inhalation exposures at Camp Lejeune

I used the ATSDR SHOWER model v4.0.1 and model parameter values described in this Report to estimate chemical exposures via inhalation and dermal contact with water from Hadnot Point in a bathroom/shower facility in a barracks at Camp Lejeune.

People residing at Camp Lejeune during the time-period that Mr. Keller was there, who lived in a similar area, and engaged in similar activities could have been exposed to the following daily exposure concentrations (**Table 6**):

Table 6. CTE and RME daily exposure concentrations for all persons using the communal facility at Hadnot Point, Camp Lejeune (November 1985 - January 1987 and July 1987 – December 1987).

Exposure Type	Daily Exposure Concentration ($\mu g/m^3$)
PCE	
CTE	NA
RME	NA
TCE	
CTE	NA
RME	NA
DCE	
CTE	NA
RME	NA
VC	
CTE	NA
RME	NA
Benzene	
CTE	0.080
RME	0.18

NA = not available, mean monthly concentrations were equal to zero

Daily exposures via inhalation and dermal contact for people on Base during the time-period that Mr. Keller was there and with the scenarios described in this Report are shown in **Table 7**. Outputs from the SHOWER model – edited for length – are provided in Attachment 1.

The SHOWER model provides exposure estimates for ATSDR standard exposure groups. I assume that the default age-specific body weights, breathing rates, and skin surface areas as described by ATSDR (2022c) are used in the SHOWER model v4.0.1. For those on Base during their mid-teen years (ages 16, 17, 18 years of age), I would report model results for 16 - < 21-year-olds (**Table 7**), since mid-teens fall within this age range. For plaintiffs ages 19 and older, results for adults are reported. This is in recognition of the fact that a 19-year-old is at the high end of the 16 - < 21-year range.

Based on Mr. Keller's birth month and year ([redacted] 1966; Scott Keller April 15, 2024 Deposition Transcript, pg. 19), he would have been at least 19 years old during his time on Base. Therefore, results for adults are included in this Report.

People residing at Camp Lejeune during the time-period that Mr. Keller was there, who lived in a similar area, and engaged in similar activities could have been exposed to the following concentrations of the chemicals of interest via dermal contact with water and inhalation of chemicals volatilized from the water in communal facilities:

- Daily exposure estimates via inhalation for benzene range from 0.018 to 0.040 µg/kg/day and via dermal contact range from 0.0071 to 0.0087 µg/kg/day.

Table 7. Daily exposures to chemicals from communal facility inhalation and dermal contact with water at Hadnot Point, Camp Lejeune (November 1985 - January 1987 and July 1987 – December 1987).

Hadnot Point	Inhalation CTE (µg/kg/day)	Inhalation RME (µg/kg/day)	Dermal CTE (µg/kg/day)	Dermal RME (µg/kg/day)
PCE	NA	NA	NA	NA
TCE	NA	NA	NA	NA
DCE	NA	NA	NA	NA
VC	NA	NA	NA	NA
Benzene	0.018	0.040	0.0071	0.0087

NA = not available, mean monthly concentrations were equal to zero

8. INGESTION ROUTE OF EXPOSURE – THE PHAST MODEL

8.1 PHAST model: Background

Due to the potential presence of PCE, TCE, DCE, VC, and benzene in water during the 1950's to the 1980's at Camp Lejeune, exposure to those chemicals via ingestion of drinking water could have occurred. The approaches and equations for estimating intake of chemicals in drinking water were established decades ago (see, for example, US EPA 1989) and continue to be used to determine human exposures. Several media and exposure routes are included in PHAST; in this section of the Report, the focus is on the model developed to estimate exposures via drinking water ingestion.

The PHAST model "is based on ATSDR's exposure dose guidance (EDGs) documents, which identify the parameters that are used to estimate exposure, either as a dose from ingestion of water or soil, or exposure as an air concentration. The EDGs were sent to EPA for review before sending them through clearance at ATSDR. PHAST is based on these EDGs and on ATSDR's public health assessment guidance manual (PHAGM), which describes the PHA [public health assessment] process that ATSDR follows when investigating hazardous waste sites" (personal communication, PHAST Team; e-mail; 26 September 2023).

8.2 PHAST model: Methodology and parameters

Ingestion of water occurs from drinking the water directly (either straight or from its use in preparation of drinks such as coffee and tea) and by its use in food preparation (e.g., soups). In the case of human exposures at Camp Lejeune, chemical intakes (i.e., doses) were computed using PHAST version 2.3. PHAST includes the standard equation for estimating chemical intakes via water, as follows:

$$D = (EPC \times IR \times EF)/BW$$

Where:

D = age-specific dose (mg/kg-day), where values for body weight and intake rate vary according to age

EPC = exposure point concentration, or contaminant concentration (mg/L)

IR = intake rate of contaminated water (L/day)

BW = body weight (kg)

EF (intermediate or chronic) = exposure factor (unitless) = (F x ED)/AT

Where:

F = exposure frequency (days/week x week/year)
ED = exposure duration (year)
AT = averaging time (ED x F)

The user enters the name of the chemical of interest and the water concentration. The PHAST drinking water model estimates both the CTE and RME for different age groups. PHAST provides

the option to use default values or to modify values for certain parameters. Appendix 4 shows the PHAST model factors and options for modification.

8.3 PHAST model: Parameter default values

In the following sections, I describe the parameters included in the PHAST water ingestion model. The default values and the bases for these values are discussed. Most of these default values were used in the modeling for this Report. The following section (8.4) includes a description of the one parameter value that was modified based on site-specific information.

Population characteristics

Scenario

The PHAST model permits the user to select from one of four scenarios: residential, daycare, school, or occupational. In this Report, the residential scenario is used. The other scenarios include inputs that allow the user to model fewer days per week and weeks per year of exposure compared to a residential scenario, but these adjustments are addressed in a separate Expert Report (Expert Report of Dr. Lisa Bailey for Scott Keller). Here, the estimation of dose is for a single day of exposure.

Body weight

The PHAST model default values for age-specific body weights are derived from the US EPA Exposure Factors Handbook (described above) (ATSDR 2024b). The SHOWER model runs for this Report utilize the default SHOWER model body weight values shown in **Table 8**. For those on Base during their mid-teen years (ages 16, 17, 18 years of age), I report model results for 16 - < 21-year-olds (**Table 8**), since mid-teens fall within this age range. In addition, I report results for adults in recognition that the weights given in **Table 8** are based on national averages, but some older teens can have weights more closely resembling adults (e.g., the 75th percentile for body weights for 16 - < 21-year-olds is 80.6 kg [US EPA 2011]). Those ages 19 and older are considered adults for the purposes of the modeling in this Report and results for adults are reported. This is because a 19-year-old is at the high end of the 16 - < 21-year range and body weights are likely more closely approximated by an adult weight than a mid-teen weight.

Based on Mr. Keller's birth month and year [redacted] 1966; Scott Keller April 15, 2024 Deposition Transcript, pg. 19), he would have been at least 19 years old during his time on Base. Therefore, the body weight for adults is used in this Report.

Table 8. Default body weights for the ATSDR PHAST model.

Exposure Group	Body weight (kg)
Birth to < 1 year	7.8
1 to < 2 years	11.4

Exposure Group	Body weight (kg)
2 to < 6 years	17.4
6 to < 11 years	31.8
11 to < 16 years	56.8
16 to < 21 years	71.6
Adult	80
Pregnant/breastfeeding women	73

kg=kilogram

Water ingestion rates

The default values used in the PHAST model for estimating intake of drinking water represent the average or "typical" and 95th percentile of the distribution for water intake for the general US population (ATSDR 2023b). PHAST utilizes the drinking water ingestion rates for different age groups for both CTE and RME exposures shown in **Table 9**. The default RME values provide a conservative estimate of water intake.

For consistency with the approach taken for body weights, the water ingestion rates for adults are used in this Report.

For those involved in training exercises or other physical activities, an additional ingestion rate of 6 L/day was included (see Section 8.4 below for additional information).

Table 9. Drinking water ingestion rates in the ATSDR PHAST drinking water ingestion model used in this Report.

Exposure Group	CTE Intake Rate (L/day)	RME Intake Rate
Birth to < 1 year	0.595	1.106
1 to < 2 years	0.245	0.658
2 to < 6 years	0.337	0.852
6 to < 11 years	0.455	1.258
11 to < 16 years	0.562	1.761
16 to < 21 years	0.722	2.214

Exposure Group	CTE Intake Rate (L/day)	RME Intake Rate
Adult	1.313	3.229
Pregnant Women	1.158	2.935
Breastfeeding Women	1.495	3.061

ATSDR based its default water intake rates on the US EPA 2019 update to its Exposure Factors Handbook. The intakes rates in **Table 9** above and those from the Exposure Factors Handbook are not identical. I explain the reason for this here.

Since the time of publication of the 2011 Exposure Factors Handbook, the US EPA has updated certain chapters and made them available online (https://www.epa.gov/expobox/about-exposure-factors-handbook). The recommended default values for ingestion of water and other fluids were updated in 2019 (https://www.epa.gov/sites/default/files/2019-02/documents/efh_-_chapter_3_update.pdf). These updated values are shown in **Table 10**.

Table 10. Default water ingestion rates from the US EPA's Exposure Factors Handbook update. Reproduced from Table 3-1 in US EPA (2019).

Table 3-1. Recommended Values for Drinking Water Ingestion Rates (2-day average community water intake)[a]

Age Group	Mean		95th Percentile		Multiple Percentiles
	mL/day	mL/kg-day	mL/day	mL/kg-day	
Per Capita[b]					
Birth to <1 month	184	42	851[c]	200[c]	See Tables 3-9 and 3-13
1 to <3 months	145	25	905[c]	164[c]	
3 to <6 months	187	27	981[c]	141[c]	
6 to <12 months	269	30	988	112	
Birth to <1 year	220	29	974	137	
1 to <2 years	146	13	565	51	
2 to <3 years	205	15	778	58	
3 to <6 years	208	11	741	42	
6 to <11 years	294	10	1,071	34	
11 to <16 years	315	6	1,395	26	
16 to <21 years	436	6	1,900	28	
21 to <30 years	781	10	2,848	39	
30 to <40 years	902	11	2,967	38	
40 to <50 years	880	11	2,964	38	
50 to <60 years	956	12	2,976	37	
60 to <70 years	941	12	2,972	35	
70 to <80 years	772	10	2,273	31	
80+ years	784	11	2,122	30	
21 to <50 years	858	11	2,938	38	
50+ years	902	11	2,827	35	
All ages	711	11	2,641	37	
Consumers-Only[d]					
Birth to <1 month	581	133	938[c]	224[c]	See Tables 3-17 and 3-21.
1 to <3 months	785	136	1,224[c]	267[c]	
3 to <6 months	649	93	1,125[c]	158[c]	
6 to <12 months	554	62	1,104[c]	133[c]	
Birth to <1 year	595	79	1,106[c]	174[c]	
1 to <2 years	245	22	658	57	
2 to <3 years	332	24	901	67	
3 to <6 years	338	19	836	45	
6 to <11 years	455	15	1,258	41	
11 to <16 years	562	10	1,761	31	
16 to <21 years	722	10	2,214	31	
21 to <30 years	1,183	16	3,407	47	
30 to <40 years	1,277	16	3,278	44	
40 to <50 years	1,356	17	3,374	43	
50 to <60 years	1,419	18	3,388	42	
60 to <70 years	1,394	17	3,187	40	
70 to <80 years	1,214	16	2,641	37	
80+ years	1,087	16	2,250	33	
21 to <50 years	1,277	16	3,353	44	
50+ years	1,343	17	3,081	40	
All ages	1,096	17	2,972	44	

Table 3-1. Recommended Values for Drinking Water Ingestion Rates (2-Day Average Community Water Intake)[a] (Continued)

Age Group	Mean		95th Percentile		Multiple Percentiles
	mL/day	mL/kg-day	mL/day	mL/kg-day	
Per Capita[b]					

[a] Ingestion rates for combined direct and indirect water from community water supply. Estimates are based on the average of 2 days of water consumption reported for each NHANES respondent. If the respondent reported zero consumption on one of the 2 days and nonzero consumption on the other day, his/her average consumption would be the average of zero and nonzero consumption.

[b] Per capita intake rates are generated by averaging consumer-only intakes over the entire population (including those individuals that reported no intake).

[c] Estimates are less statistically reliable based on guidance published in the *Joint Policy on Variance Estimation and Statistical Reporting Standards on NHANES III and CSFII Reports: NHIS/NCHS Analytical Working Group Recommendations* (NCHS, 1993).

[d] Consumer-only intake represents the quantity of water consumed only by individuals that reported consuming water during the survey period.

FCID = Food Commodity Intake Database.
NCHS = National Center for Health Statistics.
NHIS = National Health Interview Survey.

Source: U.S. EPA analysis of NHANES 2005–2010 data using the FCID Consumption Calculator at http://fcid.foodrisk.org/.

For cases in which ATSDR and the US EPA organized age ranges the same way (e.g., children ages 1 < 2 years), the ATSDR (**Table 9**) and the US EPA (**Table 10**; "consumers only" data) intake values are identical. However, in certain cases ATSDR utilized different age groupings than the US EPA. So, for example, for children ages 2 to < 6 years, the ATSDR CTE and RME intake values are equivalent to the time-weighted sum of the values for children ages 2 to < 3 and 3 to < 6 in the US EPA's consumers-only data above. Similarly, the ingestion rate value for adults (21-78 years) is the time-weighted average of the US EPA age groups within that age range, as shown in **Table 10** above (these calculations are described in ATSDR 2023b, Appendix C). Therefore, even though the numbers in the Tables appear to differ, they are in fact derived from the same underlying database. The intake rates for pregnant and breastfeeding women are taken directly from the US EPA's Exposure Factors Handbook (Table 3-3).

The water intake rates in **Table 10** represent both direct ingestion (i.e., drinking water as a beverage) and indirect ingestion (e.g., intake of water that has been added during food and drink preparation) (US EPA 2019). The intake rate values are derived from NHANES and were estimated only from those NHANES participants who reported consuming water during the NHANES survey period (US EPA 2019). The values are considered to be representative of the general population in the US (https://perma.cc/5GFB-SHV9).

8.4 PHAST model: Plaintiff-specific modifications

The only parameter value that was modified from the default in the PHAST model was that for drinking water intake rates. In addition to the default values for drinking water intake described above, in this Report, I estimated exposures using additional more conservative drinking water intake values. These were included because certain populations at Camp Lejeune engaged in training or other activities that may have resulted in water consumption at rates that exceed a conservative (i.e., RME) estimate for the general population. These values are likely to capture

even the highest of average daily water intake rates at Camp Lejeune. (Note that for these high intake rates, the PHAST model includes the following warning: "The value entered exceeds the 99th percentile for drinking water ingestion rates documented in EPA's Exposure Factor Handbook.") Additional information on the use of 6 L/day as a high-end value for water consumption is given here.

First, the US EPA (2019) provided the following quote from Montain and Ely (2010): "...an individual soldier's daily water requirements to sustain hydration can range from 2 L/day to an excess of 12 L/day, depending on weather conditions, physical activity, and physical size."

Further, the US Army provides information regarding water consumption requirements developed for planning purposes (Table 3-87, US EPA 2019; **Table 11** below). Based on anticipated environmental conditions, the range of intakes is given as 7.6 to 11.4 L/day. The US EPA (2019) noted that: "[t]he advantage of using these data is that they provide conservative estimates of drinking water intake among individuals performing at various levels of physical activity in hot, temperate, and cold climates. However, the planning factors described here are based on assumptions about water loss from urination, perspiration, and respiration, and are not based on survey data or actual measurements."

Table 11. Individual water consumption for the US Army. Reprinted from US EPA (2019).

Table 3-87. Planning Factors for Individual Tap Water Consumption

Environmental Condition	Recommended Planning Factor (gal/day)[a]	Recommended Planning Factor L/day[a,b]
Hot	3.0[c]	11.4
Temperate	1.5[d]	5.7
Cold	2.0[e]	7.6

[a] Based on a mixture of activities among the workforce as follows: 15% light work; 65% medium work; 20% heavy work. These factors apply to the conventional battlefield where no nuclear, biological, or chemical weapons are used.
[b] Converted from gal/day to L/day.
[c] This assumes 1 quart/12-hour rest period/man for perspiration losses and 1 quart/day-man for urination plus 6 quarts/12 hours of light work/man, 9 quarts/12 hours of moderate work/man, and 12 quarts/12 hours of heavy work/man.
[d] This assumes 1 quart/12-hour rest period/man for perspiration losses and 1 quart/day/man for urination plus 1 quart/12 hours of light work/man, 3 quarts/12 hours of moderate work/man, and 6 quarts/12 hours of heavy work/man.
[e] This assumes 1 quart/12-hour rest period/man for perspiration losses, 1 quart/day/man for urination, and 2 quarts/day/man for respiration losses plus 1 quart/12 hours of light work/man, 3 quarts/12 hours of moderate work/man, and 6 quarts/6 hours of heavy work/man.

Source: U.S. Army (1983, 1999).

As noted by the US Armed Forces Health Surveillance Branch Communications Team (2021), "In response to previous historical cases of exertional hyponatremia in the U.S. military, the guidelines for fluid replacement during military training in hot weather were revised and promulgated in 1998...The revised guidelines were designed to protect service members from not

only heat injury, but also hyponatremia due to excessive water consumption by limiting fluid intake regardless of heat category or work level to no more than 1.5 quarts hourly and 12 quarts daily." The value of 12 quarts per day is equivalent to 11.36 L/day.

Using these data, the range of drinking water intakes for those in the military could be estimated as approximately 2 to 11.36 L/day. The value of 11.36 L/day could be considered a *maximum* intake rate. It is unlikely that this would be a long-term daily intake rate as Camp Lejeune does not experience high heat conditions every day year-round (https://www.onslowcountync.gov/DocumentCenter/View/1360/Cover-PDF?bidId=) nor do plaintiffs generally report participating in physical training every day.

While these documents suggest that under certain conditions, drinking water intakes of up to 12 L/day *could* occur, according to ATSDR (2017b, pg. 3), "A marine in training at Camp Lejeune consumes an estimated 6 liters of water per day for three days per week and 3 liters per day the rest of the week (ATSDR 2016). Under warm weather conditions, a marine may consume between 1 and 2 quarts of water per hour… (Bove et al. 2014a)." As noted above, Camp Lejeune does not experience warm weather year-round. It was noted by ATSDR (2017a) that this information was "…developed by combining information gathered from former Marines at the community assistance panel meetings and recommended military fluid replacement guidelines (Kolka et al. 2003)." Using this information, the daily water intake rate for marines in training would be 4.3 L/day on average during weeks when training was taking place. For those on Base doing activities other than training, it is likely that the CTE or RME intake rates of approximately 1 to 3 L/day are more representative.

Based on my review of depositions and the information above, the CTE and RME water intake values for adults as well as the conservative intake value of 6 L/day would be appropriate for most plaintiffs. However, Mr. Keller recalled drinking amounts of water described variously as **"Anywhere's from a gallon to two, two and a half gallons a day" (Scott Keller April 15, 2024 Deposition Transcript, pg. 141), "Two to four canteens minimum. Depends how hot it was" (Scott Keller April 15, 2024 Deposition Transcript, pg. 140), and "one to two gallons a day, if not increased during the summer. I mean, you were hydrating all the time" (Scott Keller April 15, 2024 Deposition Transcript, pg. 165). He stated that a canteen held one quart of water (Scott Keller April 15, 2024 Deposition Transcript, pg. 140). His minimum reported consumption of two canteens, or two quarts, is equivalent to 1.9 L, which is greater than the CTE intake value of 1.313 L/day but less than the RME intake value of 3.229 L/day used in the models in this Report. His reported consumption of one gallon of water is equivalent to 3.8 L, which is somewhat higher than the RME intake value of 3.229 L/day but less than the conservative value of 6 L/day. On days when he drank the reported value of 2.5 gallons, this would be equivalent to 9.5 L. It is unlikely that Mr. Keller consumed an average of 9.5 L per day over his entire time at Camp Lejeune, since some of his time there was spent in cooler seasons and in an office building. However, because he recalled very high water intake rates, in this Report, in addition to the water intake rates of 1.313, 3.229, and 6 L/day, I also estimate Mr. Keller's exposure associated with a water intake of 9.5 L/day.**

8.5 Opinion: Exposure via water ingestion at Camp Lejeune

I used the PHAST drinking water model with parameter values described in this Report to estimate chemical exposures via drinking water from the Hadnot Point and Tarawa Terrace water systems. Daily exposures via water ingestion for people on Base during the time-period when Mr. Keller was at Camp Lejeune, with the scenarios described in this Report, are shown in **Table 12**. Results are provided for both Hadnot Point and Tarawa Terrace as Mr. Keller did not specify that he only consumed water from one of these water sources.

People residing at Camp Lejeune during the time-period that Mr. Keller was there, who lived in a similar area, and engaged in similar activities could have been exposed to the following concentrations of the chemicals of interest via ingestion of water, with the ranges reflecting different likely behaviors:

- Daily exposure estimates via water ingestion for PCE range from 9.4E-05 to 0.00068 mg/kg/day.
- Daily exposure estimates via water ingestion for TCE range from 4.9E-06 to 3.6E-05 mg/kg/day.
- Daily exposure estimates via water ingestion for DCE range from 2.1E-05 to 0.00015 mg/kg/day.
- Daily exposure estimates via water ingestion for VC range from 1.5E-05 to 0.00011 mg/kg/day.
- Daily exposure estimates via water ingestion for benzene range from 4.8E-05 to 0.00034 mg/kg/day.

Table 12. Daily intakes of chemicals from drinking water at Hadnot Point and Tarawa Terrace, Camp Lejeune (November 1985 - January 1987 and July 1987 – December 1987).

Adult	Default Dose CTE (mg/kg/day)	Default Dose RME (mg/kg/day)	Dose – Conservative (mg/kg/day)	Highest Dose (mg/kg-day)
Hadnot Point				
PCE	NA	NA	NA	NA
TCE	NA	NA	NA	NA
DCE	NA	NA	NA	NA
VC	NA	NA	NA	NA
Benzene	4.8E-05	0.00012	0.00022	0.00034
Tarawa Terrace				
PCE	9.4E-05	0.00023	0.00043	0.00068
DCE	2.1E-05	5.2E-05	9.8E-05	0.00015
TCE	4.9E-06	1.2E-05	2.2E-05	3.6E-05
VC	1.5E-05	3.6E-05	6.8E-05	0.00011

CTE is based on a water intake rate of 1.313 L/day. RME is based on a water intake rate of 3.229 L/day. Conservative estimate based on water intake rate of 6 L/day.

9. CONSERVATIVE NATURE OF SELECTED MODEL INPUTS

As noted in previous sections of this Report, there are either limited or no data on various chemical (e.g., water concentrations) and behavioral (e.g., shower duration, water consumption) aspects of plaintiffs' chemical exposure during their time on Base. Some inputs for model parameters used in this Report are based on information recalled by Mr. Keller. However, plaintiffs may not always recall the details of their environment or behaviors from decades prior. Thus, while information from plaintiffs on their behaviors is used as guidance for selecting parameter input values, judgment is also used to ensure that the exposure estimates are *not likely* to underestimate overall exposures during a plaintiff's time on Base.

In this Report, I used model input values that in my view should provide conservative estimates of exposure (i.e., not result in underestimates of Mr. Keller's exposures). These are described in the following sections (these were mentioned in previous sections and are reiterated in this summary).

Overall, regarding the estimates for the mean monthly chemical concentrations in water developed by ATSDR and used in this Report, according to the Expert Report of Dr. Spiliotopoulos (pgs. 68-69): "For Hadnot Point, as with Tarawa Terrace, ATSDR assumed concentrations of contaminants in the influent to the WTP were equal to the concentrations of contaminants in the 'finished water' that was delivered to consumers...This assumption is incorrect, as treatment of the influent to the treatment plant resulted in evaporative and other losses, reducing contaminant concentrations in the 'finished' water." Further, according to the Expert Report of Dr. Hennet (pg. 5-43): "A substantial portion of [contaminants of concern] that may have been present in the treated water used to fill a water buffalo would have unavoidably been lost to evaporation during filling, use, and variations of temperature. These [contaminants of concern] reductions between the raw water and the water in the water buffaloes would have been in the order of 52% to 73% based on my estimation." *Therefore, the chemical concentrations used in this Report as well as the associated estimates of Plaintiff exposure would be overly conservative (too high)*.

9.1 Drinking water

<u>Chemical concentrations</u>: The chemical concentrations were based on monthly mean concentration data for the months that Mr. Keller was on Base. Assumptions were made that would result in conservative estimates of the number of months on Base. Specifically: (i) If a plaintiff was on Base for part of the month, I assumed that the plaintiff was there for the entire month (the exception to this was if the plaintiff was only on Base for one day for that month). (ii) Plaintiffs may have been off-Base for part of their time assigned to Camp Lejeune (e.g., leave, weekends away, time spent on parts of the Base where water was not impacted). Unless they were off Base for at least one calendar month and the exact dates were known, it was assumed that they were on Base and exposed to the chemicals of interest for the entire time-period. I recognize that while these assumptions result in conservative (longer) estimates of time on Base, they may not always yield the most conservative estimates of water concentrations.

While plaintiffs (including Mr. Keller) often noted the use of water buffaloes for obtaining drinking water during days in the field, it is not clear what water source was used for filling these tanks. Based on the expert opinion of Dr. Hennet (2024, pg. 3-1), "...the only Base water supply systems contaminated with the [contaminants of concern] were Tarawa Terrace, Hadnot Point, and Holcomb Boulevard." If the water sources were neither the Hadnot Point nor Tarawa Terrace water systems, then the modeled values for exposure from drinking water in this Report would be over-estimates.

Intake rates: It is unreasonable to expect that any individual would recall their exact water intake from their time on Base. In depositions that I reviewed, volumes of water intake were variably described using language such as "cups," "glasses," "sips," or "canteens" (and the descriptions of the size of a canteen varied). It is also unlikely that any individual would consume the same amount of water each day, and this is borne out by deposition statements in which plaintiffs note varying water consumption, depending on outdoor temperature and activities.

The model used in this Report provides an estimate of average daily water consumption over the duration of time spent on Base. Without exact information from plaintiffs on water consumption, it is reasonable to use national estimates of daily water intake (CTE equal to 1.313 L/day and RME equal to 3.229 L/day). At the same time, plaintiffs describe situations in which they consumed what they recall as large quantities of water (or fluids made from water such as "bug juice"), including physical training in hot weather. To ensure that this Report captures these high-end water consumption scenarios, a water intake rate of 6 L/day was included.

To visualize this amount of water intake, it is useful to recall that there are 8 fluid ounces in a cup. Consumption of 1.313 L as drinking water in a day is equivalent to about 44.4 ounces or about 5 and a half 8-ounce cups of water. Consumption of 3.229 L in a day is equivalent to about 13 and a half 8-ounce cups of water (or about a full glass of water every hour during the day). The higher-end estimate of 6 L in a day is equal to about 25 8-ounce glasses of water (or about two full glasses every hour during a 12-hour day).

While Mr. Keller noted that he may have consumed 9.5 L water (40 8-ounce glasses) in a day, use of this value to reflect long-term average intake is likely an overestimate of drinking water intake as high temperatures are seasonal and Mr. Keller did not describe engaging in 7-days per week physical training during his entire time on Base **(although he frequently used the gymnasium on Base) (Scott Keller April 15, 2024 Deposition Transcript, pgs. 160-161**). Thus, I consider inclusion of a drinking water intake rate of 9.5 L/day to be very conservative.

9.2 Showering

It is not expected that plaintiffs would recall their exact amount of time spent showering, nor that their shower durations would be the same from day to day. In this Report, I assumed that the average shower duration for the people in the barracks with Mr. Keller was 20 minutes. If the mean shower duration was shorter, then the inhalation and dermal results in this Report would be conservative. I also assumed a peak shower usage from 6:00 am - 7:00, which yields more conservative exposure estimates than assuming that the showers are used continuously

throughout the day. Finally, for the communal facilities, because of a lack of site-specific information on airflow, I used the default model values for facility size, configuration and airflow, which yields more conservative exposure estimates than using site-specific information with default airflow values.

10. REBUTTAL TO EXPERT REPORT BY DR. REYNOLDS

My overall approach to estimating exposures to chemicals of interest is similar to that of Dr. Reynolds in that we both provide a range of exposure estimates for each plaintiff. However, my approach differs from Dr. Reynolds' approach in several respects (described in the following paragraphs). In my opinion, and based on my training and professional experience in assessing exposures to chemicals, my assumptions are both conservative (in other words, would be unlikely to underestimate exposure) and more reasonable (i.e., supported by the scientific literature, Mr. Keller's records, and my training, experience, and professional judgment). My exposure estimates consequently provide a more appropriate picture of Mr. Keller's exposure to chemicals of interest than Dr. Reynolds' estimates.

In the following sections, I describe the general differences in approach between Dr. Reynolds' report and my Reports (Section 10.1) and differences specific to Mr. Keller (Section 10.2).

10.1 General differences in approaches

10.1.1 Exposure route differences

Dr. Reynolds' exposure estimates are based on one exposure route: consumption of drinking water. However, plaintiffs would have also been exposed via the dermal and inhalation routes of exposure. In this Report, I use models to address these routes. In addition, where relevant, I use models to assess plaintiff exposures for specific additional scenarios including swimming pools and the mess hall. These were not addressed in Dr. Reynolds' overall report. Including these other exposure routes provides a more realistic picture of plaintiffs' potential exposure based on the available evidence. My inclusion of three routes of exposure provides a more conservative (i.e., higher) estimate of exposure compared to the exposure estimate I *would* have obtained had I only included the water ingestion route of exposure (as was done by Dr. Reynolds). As discussed in the Expert Report of Dr. Lisa Bailey, including these more realistic exposure routes does not result in an unacceptable cancer risk for people who resided at Camp Lejeune during the time-period that Mr. Keller was there, who lived in similar areas, and engaged in similar activities.

10.1.2 "Cumulative consumption" versus daily intake

Dr. Reynolds provided exposure results in the form of "cumulative consumption," or the total number of micrograms of a chemical consumed by each plaintiff via drinking water over their entire time at Camp Lejeune, whereas I accounted for the body weight of the plaintiff. Generally, I used age-based default values (as described in this Report) to adjust for dose.

Inclusion of an approximate body weight (e.g., adult versus child) enhances one's ability to interpret the exposure results in a risk-based context. Generally speaking, given the same of amount of chemical intake, the lighter the person, the higher the dose. To use a familiar example, "...smaller people usually have a higher ratio of alcohol in their blood if they drink the same amount a heavier person drinks..." (https://www.stanfordchildrens.org/en/topic/default?id=understanding-alcohols-effects-1-2860).

In using this method, I employed the approach used by ATSDR in its PHAST and SHOWER models, as well as the US EPA in its Risk Assessment Guidance for Superfund (1989), and assessed average daily exposures for each plaintiff in units of mg/kg-day or µg/kg-day. Average daily exposure values are the foundation for estimating human health risks (see Expert Report of Dr. Lisa Bailey). Dr. Reynolds instead represents the exposure results in terms of cumulative consumption.

10.1.3 Water ingestion rates

Default values: The default values for CTE and RME estimates in this Report are derived from the most recent US EPA Exposure Factors Handbook (updated drinking water ingestion chapter from 2019). For example, for adults, I used values of 1.313 and 3.229 L/day for CTE and RME estimates, respectively. These values are used by ATSDR in its PHAST model.

In contrast, Dr. Reynolds used CTE and RME values of 1.227 and 3.092 L/day, respectively. According to Dr. Reynolds, these values are derived from the US EPA's Exposure Factors Handbook (2011). These values were updated by the US EPA in 2019 (US EPA 2019). I used the updated values, which are more conservative for adults and therefore would result in a more conservative exposure estimate for adults.

Other values: For plaintiffs who were marines in training on Base, Dr. Reynolds used ATSDR (2017) values to estimate drinking water intake rates: 6 L/day for 3 days per week and 3.1 L/day for 4 days per week. The overall weighted value reported by Dr. Reynolds is 4.334 L/day[5]. However, in at least one instance (Expert Report of Dr. Reynolds, pg. 126), Dr. Reynolds assumed a plaintiff consumption of 6 L/day for 3 days per week and 3 L/day for 4 days per week, for an overall weighted value of 4.29 L/day. She does not provide justification for selecting one over the other.

I used a high-end value of 6 L/day to estimate drinking water intake for marines in training. As noted in Section 8.4 of this Report, according to ATSDR (2017b), "A marine in training at Camp Lejeune consumes an estimated 6 liters of water per day for three days per week." Thus, this value is a reasonable estimate for the high end of the range of water intake rates for a marine at Camp Lejeune. The estimate of 6 L in a day is equal to about 25 8-ounce glasses of water (or about two full glasses every hour during a 12-hour day). I did not make assumptions regarding the number of days per week that a plaintiff engaged in heavy activity (see the Expert Report of Dr. Lisa Bailey).

Dr. Reynolds stated in her Report (pg. 5): "For some plaintiffs, specific information was available in their deposition detailing their training and consumption habits…if consumption data was given, for example, recall of refilling and drinking a specific number of canteens (estimated to hold 32 oz each) during training, or a specific amount of coffee or tea (5-10 oz cups), "bug juice" or glasses of water (12 oz cups), or other beverage made from the contaminated water sources, deposition-informed ingestion data was used in the exposure assessment." Dr. Reynolds utilized

[5] See, for example, pg. 26 of Dr. Reynolds' Expert Report. Based on my calculations, the weighted daily value should be: (6x3) + (3.1x4) /7 = (18 + 12.4)/7 = 30.4/7 = 4.34 L/day.

this kind of information from the depositions to develop water intake rates that appear to be very accurate, including several significant digits (e.g., "3.54882" L/day, pg. 27).

However, Dr. Reynolds' degree of implied accuracy is not supported by the record. As noted in this Report, in my professional opinion and based on my professional experience, it is unreasonable to expect that any individual could recall their *exact* daily water intake from their time on Base decades ago. Further, variations in water intake from one day to the next are expected as "...individual water requirements can vary greatly on a day-to-day basis because of differences in physical activity, climates, and dietary contents" (Armstrong and Johnson 2018, pgs. 1-2). Therefore, I did not assume that plaintiff-derived information on amounts of water (or water-based drinks such as coffee) are *exact* amounts consumed by a plaintiff every day for their entire time at Camp Lejeune. Rather, I used the plaintiff deposition water intake information to describe whether the use of the CTE and RME values are indicated (i.e., does the CTE/RME range of water intake include the water consumption amounts that were generally recalled by the plaintiff?). Thus, this degree of implied accuracy in Dr. Reynold's Report is not supported by the record.

For some plaintiffs, Dr. Reynolds also used other higher estimates of water intake than the ones that were provided by ATSDR (2017). She stated that she relied on US Army Field Manuals (FM) for information on water intakes associated with light and heavy activity to derive additional water intake values of 5.21 L/day and 8.52 L/day (Expert Report of Dr. Reynolds, pg. 7).

As noted in Section 8 of this Report, according to ATSDR (2017b, pg. 3), "A marine in training at Camp Lejeune consumes an estimated 6 liters of water per day for three days per week and 3 liters per day the rest of the week (ATSDR 2016). Under warm weather conditions, a marine may consume between 1 and 2 quarts of water per hour... (Bove et al. 2014a)." The value of "1 to 2 quarts of water per hour" is difficult to rely on as the number of hours is not provided. However, the estimate of 6 L/day is supportable given that the information is specific to marines at Camp Lejeune. Therefore, the value of 6 L/day (as used in this Report) is a reasonable and conservative value for water intake by a marine in training. Additional information on issues in using the FM-based information is given in Section 10.2.

10.2 Differences specific to Mr. Keller

In addition to differences in overall approach, two of my plaintiff-specific assumptions differ from Dr. Reynolds' assumptions. Differences specific to Mr. Keller are described here.

10.2.1 Water systems

Dr. Reynolds used data only for the Hadnot Point water system.

I assumed that Mr. Keller could have spent time on parts of the Base ingesting water from either the Hadnot Point or Tarawa Terrace water systems. I therefore modeled his exposure via drinking water consumption using data from both the Hadnot Point and Tarawa Terrace water systems.

10.2.2 Water ingestion

Dr. Reynolds assumed that Mr. Keller consumed 6.94 L/day of Kool-Aid and 2.84 L/day of drink from his canteen (Expert Report of Dr. Reynolds, pg. 120). However, this Kool-Aid consumption rate is not supported by the record. Mr. Keller was asked about his Kool-Aid consumption (Scott Keller April 15, 2024 Deposition Transcript, pg. 141):

Q. Okay. And then how often would you say you drank Kool-Aid?

A. Some meals. Not a lot, 'cause there's a lot of sugar in it.

Dr. Reynolds also relied on the values from the FMs. According to Dr. Reynolds (Expert Report, pg. 6): "FM [Field Manual] ingestion values were selected as recommended for a moderate temperature day in a tropical environment with temperatures exceeding 80°F and with differentiation between light and heavy activities. FM 1957-1983 defines light activities as desk work, guard/kitchen duties while heavy activities included forced marches, entrenching or route marches with heavy loads, or wearing protective clothing."

The Field Manuals referenced by Dr. Reynolds describe the temperature noted by Dr. Reynolds as "80°F" two different ways. The Field Manuals from 1957 and 1970 indicate that the water consumption values correspond to air temperatures below 105 °F in desert environments and below 85 °F in tropical environments (CLJA_ARMYFH_0000000532, CLJA_ARMYFH_0000000915). The Field Manuals from 1980 and 1982[6] indicate that the water consumption values correspond to air temperatures below 80° given as a Wet Bulb Globe Temperature[7], which according to the 1980 Manual is approximately equal to the temperatures in the preceding sentence (below 105 °F in desert environments and below 85 °F in tropical environments) (Official U.S. Military Field Manual, 1980, pg. 5; Official U.S. Military Final Report, 1982, pg. 36). The values from 1982 are included under the table header "Water Requirements in Hot Environments" and for 1980 the values are described as "drinking water requirements for personnel exposed to heat."

In summary, Dr. Reynolds' description of temperatures for the values that I assume[8] she used to estimate her Field Manual-based intakes of 8.52 and 5.21 L/day appear to contradict the temperatures in the Field Manual. Despite this, in my view the Field Manuals are clear that the intake values are for hot temperatures.

Dr. Reynolds' use of Field Manual values for year-round exposure estimates is not appropriate for Mr. Keller. As demonstrated in **Figure 5**, many of the months Mr. Keller was on Base likely

[6] This is dated 1983 in Dr. Reynolds' Expert Report, but the document provided is dated 1982.

[7] "The WetBulb Globe Temperature (WBGT) is a measure of the heat stress in direct sunlight, which takes into account: temperature, humidity, wind speed, sun angle and cloud cover (solar radiation)." https://perma.cc/9QU9-VNXL

[8] This is an assumption on my part as Dr. Reynolds does not specify the exact values that she relied on for her estimates. My assumption is based on my reproduction of Dr. Reynolds' values using either 5 or 6 quarts/day (average of 5.5 quarts per day or 5.2 L/day) and 9 quarts per day (or 8.82 L/day).

experience temperatures well below 80 °F and would not be described as "hot." Therefore, Dr. Reynolds' use of the Field Manual water intake values meant for hot temperatures for year-round exposure estimates is not supported by the evidence.

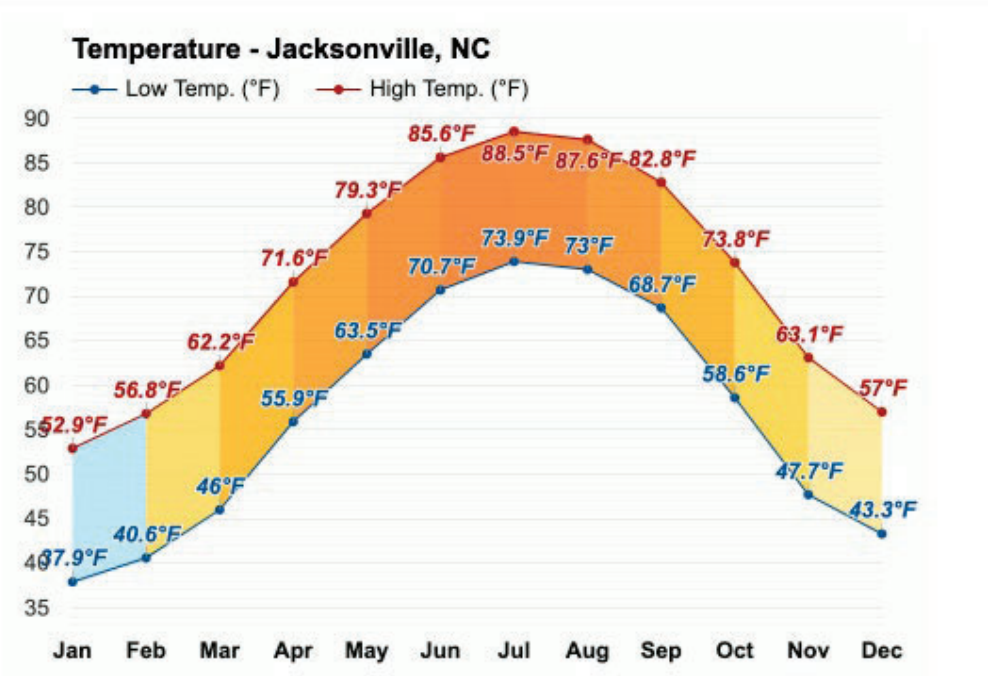

Figure 5. Average high and low monthly temperatures in Jacksonville, NC (reprinted from https://www.weather-us.com/en/north-carolina-usa/jacksonville-weather-march#google_vignette)

11. CONCLUSIONS

People living and working at Camp Lejeune from the 1950's to the 1980's may have been exposed to PCE, TCE, DCE, VC and/or benzene due to the presence of these chemicals in finished water at Camp Lejeune.

Dr. Spiliotopoulos (Expert Report, 2024, pgs. 68-69) stated that "For Hadnot Point, as with Tarawa Terrace, ATSDR assumed concentrations of contaminants in the influent to the WTP [water treatment plant] were equal to the concentrations of contaminants in the 'finished water' that was delivered to consumers...This assumption is incorrect, as treatment of the influent to the treatment plant resulted in evaporative and other losses, reducing contaminant concentrations in the 'finished' water." *Based on this opinion, the concentrations of chemicals of interest used in this Report, derived from ATSDR modeling, would be an overestimate of chemical concentrations in water used by people at Camp Lejeune*.

The routes of exposure could have included:

- Ingestion (for example, drinking the water, using the water for cooking, drinking small amounts of water during swimming)
- Inhalation (breathing the chemicals that volatilized from the water during activities such as showering, bathing, swimming, or using appliances such as washing machines)
- Skin contact (dermal exposure from contacting the water during activities such as showering, bathing, hand washing, or swimming)

The exposure assessment in this Report is intended to capture exposures experienced by people residing and/or working at Camp Lejeune during a time-period specific to the Plaintiff's actual time on Base combined with exposure-related information generally considered to be representative of people on Base (with some conservative assumptions). The exposure assessment in this Report is not a perfectly accurate representation of exposure to a specific individual because the information necessary to develop such a representation is not available. For example, no contemporaneous documentation (e.g., diaries) describing day-to-day activities was identified. However, exposures can still be assessed by making assumptions derived from information from depositions, other sources of information related to the United States population, the military in general, Camp Lejeune specifically, and my best professional judgment.

Using these existing data in conjunction with modeled water concentration data, I was able to draw conclusions about Mr. Keller's likely exposures to PCE, TCE, DCE, VC, and benzene to a reasonable degree of scientific certainty, considering my use of ATSDR's modeled chemical concentrations in water, as detailed in this Report. Where possible, conservative assumptions were made for determining model inputs. Conservative assumptions are used to avoid underestimating exposures. Therefore, Mr. Keller's actual exposures are unlikely to be higher than the exposure estimates produced by these models. These exposure estimates can be used in risk

assessments to determine whether people who resided at Camp Lejeune during the time-period that Mr. Keller was there, who lived in similar areas, and engaged in similar activities had an increased risk of disease (this is addressed in the Expert Report of Dr. Lisa Bailey for Scott Keller).

People residing at Camp Lejeune during the time-period that Mr. Keller was there, who lived in a similar area, and engaged in similar activities could have been exposed to the following concentrations of the chemicals of interest via dermal contact with water and inhalation of chemicals volatilized from the water in communal facilities:

- Daily exposure estimates via inhalation for benzene range from 0.018 to 0.040 µg/kg/day and via dermal contact range from 0.0071 to 0.0087 µg/kg/day.

People residing at Camp Lejeune during the time-period that Mr. Keller was there, who lived in a similar area, and engaged in similar activities could have been exposed to the following concentrations of the chemicals of interest via ingestion of water, with the ranges reflecting different likely behaviors:

- Daily exposure estimates via water ingestion for PCE range from 9.4E-05 to 0.00068 mg/kg/day.
- Daily exposure estimates via water ingestion for TCE range from 4.9E-06 to 3.6E-05 mg/kg/day.
- Daily exposure estimates via water ingestion for DCE range from 2.1E-05 to 0.00015 mg/kg/day.
- Daily exposure estimates via water ingestion for VC range from 1.5E-05 to 0.00011 mg/kg/day.
- Daily exposure estimates via water ingestion for benzene range from 4.8E-05 to 0.00034 mg/kg/day.

12. REFERENCES

Agency for Toxic Substances and Disease Registry (ATSDR). 2016. Exposure Dose Guidance for Body Weight. Atlanta, GA: U.S. Department of Health and Human Services, Public Health Service, October 26.

Agency for Toxic Substances and Disease Registry (ATSDR). 2017a. Public Health Assessment for Camp Lejeune Drinking Water. U.S. Marine Corps Base Camp Lejeune, North Carolina. January 20, 2017.

Agency for Toxic Substances and Disease Registry (ATSDR). 2017b. ATSDR Assessment of the Evidence for the Drinking Water Contaminants at Camp Lejeune and Specific Cancers and Other Diseases 13 January.

Agency for Toxic Substances and Disease Registry (ATSDR). 2018. Exposure Dose Guidance for Dermal and Ingestion Exposure to Surface Water. Atlanta, GA: U.S. Department of Health and Human Services, Public Health Service, Sept 25.

Agency for Toxic Substances and Disease Registry (ATSDR). 2022a. Estimating Site-Specific Ingestion and Dermal Exposure Doses. https://www.atsdr.cdc.gov/pha-guidance/conducting_scientific_evaluations/epcs_and_exposure_calculations/estimating-site-specific-ingestion-and-dermal-exposure-doses.html

Agency for Toxic Substances and Disease Registry (ATSDR). 2022b. Shower and Household Water-use Exposure (SHOWER) Model v3.0 User's Guide. September 2.

Agency for Toxic Substances and Disease Registry (ATSDR). 2022c. Technical Document for the Shower and Household Water-use Exposure (SHOWER) Model v3.0. May 4.

Agency for Toxic Substances and Disease Registry (ATSDR). 2023a. Exposure Dose Guidance for Soil/Sediment Dermal Absorption. Atlanta, GA: U.S. Department of Health and Human Services, Public Health Service, July 2023.

Agency for Toxic Substances and Disease Registry (ATSDR). 2023b. Exposure Dose Guidance for Water Ingestion. Atlanta, GA: U.S. Department of Health and Human Services, Public Health Service, Jan 31.

Agency for Toxic Substances and Disease Registry (ATSDR). 2024a. Shower and Household Water-use Exposure (SHOWER) Model v4.0 User's Guide Version 4.0. September 25, 2024.

Agency for Toxic Substances and Disease Registry (ATSDR). 2024b. PHAST User Guide. Version 2.4. April.

Armstrong LE, Johnson EC. 2018. water intake, water balance, and the elusive daily water requirement. Nutrients 10(12):1928. doi: 10.3390/nu10121928.

Baier-Anderson C, Blount BC, Lakind JS, Naiman DQ, Wilbur SB, Tan S. Estimates of exposures to perchlorate from consumption of human milk, dairy milk, and water, and comparison to current reference dose. J Toxicol Environ Health A. 2006 Feb;69(3-4):319-30. doi: 10.1080/15287390500323420. PMID: 16407090.

Chowdhury S. 2015. Predicting human exposure and risk from chlorinated indoor swimming pool: a case study. Environ Monit Assess. 187(8):502. doi: 10.1007/s10661-015-4719-8.

DeOreo WB, Mayer P, Dziegielewski B, Kiefer J. 2016. Residential End Uses of Water, Version 2. Water Research Foundation, Denver.

EarthCon. 2019. Human Health Risk Assessment, Kerr-McGee Chemical Corp - Navassa Superfund Site, Navassa, North Carolina. EPA ID #NCD980557805.

EFSA (European Food Safety Authority). 2012. SCIENTIFIC OPINION Guidance on selected default values to be used by the EFSA Scientific Committee, Scientific Panels and Units in the absence of actual measured data. EFSA Journal 10(3):2579. https://efsa.onlinelibrary.wiley.com/doi/pdf/10.2903/j.efsa.2012.2579#:~:text=In%20this%20context%2C%20the%20term,and%20be%20therefore%20scientifically%20justified.

Health Canada. 1999. Canadian exposure factors used in human health risk assessments Fact sheet series: Topics in risk assessment of substances under the Canadian Environmental Protection Act, 1999 (CEPA 1999). https://www.canada.ca/content/dam/hc-sc/documents/services/chemical-substances/fact-sheets/canadian-exposure-factors-human-health-risk-assessments/fact-sheet-exposue-factors.pdf.

Health Canada. 2021. Federal Contaminated Site Risk Assessment in Canada: Guidance on Human Health Preliminary Quantitative Risk Assessment (PQRA) VERSION 3.0.

Huerta D, Schobel T, Alexander-Ozinskas A, Hild J, Lauder J, Reynolds P, Von Behren J, Meltzer D, Ramírez-Andreotta MD. 2023. Probabilistic risk assessment of residential exposure to metal(loid)s in a mining impacted community. Sci Total Environ. 872:162228. doi: 10.1016/j.scitotenv.2023.162228.

IPCS (International Programme on Chemical Safety). 2004. IPCS Risk Assessment Terminology. IPCS harmonization project document n. 1. https://www.who.int/publications/i/item/9241562676

Jang W, Aral MM. 2008. Analyses of Groundwater Flow, Contaminant Fate and Transport, and Distribution of Drinking Water at Tarawa Terrace and Vicinity, U.S. Marine Corps Base Camp Lejeune, North Caroline: Historical Reconstruction and Present-Day Conditions – Chapter G: Simulation of Three Dimensional Multispecies, Multiphase Mass Transport of Tetrachloroethylene (PCE) and associated Degradation By-Products. Atlanta, GA: Agency for Toxic Substances and Disease Registry.

Khan K, Khan MS, Younas M, Yaseen M, Al-Sehemi AG, Kavil YN, Su C, Ali N, Maryam A, Liang R. 2024. Pathways and risk analysis of arsenic and heavy metal pollution in riverine water: Application of multivariate statistics and USEPA-recommended risk assessment models. J Contam Hydrol. 269:104483. doi: 10.1016/j.jconhyd.2024.104483.

Kolka MA, Latzka WA, Montain SJ, Corr WP, O'Brien KK, Sawka MN. Effectiveness of revised fluid replacement guidelines for military training in hot weather. Aviat Space Environ Med. 2003 Mar;74(3):242-6. PMID: 12650271.

Lowe JA, Jamall IS. 1994. Assessing health risks associated with DDT residues in soils in California: a Proposition 65 case study. Risk Anal. 14(1):47-52. doi: 10.1111/j.1539-6924.1994.tb00027.x.

Maslia ML, Sautner JB, Faye RE, Suárez-Soto RJ, Aral MM, Grayman WM, Jang W, Wang J, Bove FJ, Ruckart PZ, Valenzuela C, Green JW, Krueger AL. 2007. Analyses of Groundwater Flow, Contaminant Fate and Transport, and Distribution of Drinking Water at Tarawa Terrace and Vicinity, U.S. Marine Corps Base Camp Lejeune, North Carolina: Historical Reconstruction and Present-Day Conditions. Chapter A: Summary of Findings. Appendix A2.

Maslia ML, Suárez-Soto RJ, Sautner JB, Anderson BA, Jones LE, Faye RE, Aral MM, Guan J, Jang W, Telci IT, Grayman WM, Bove FJ, Ruckart PZ, Moore SM. 2013. Analyses and Historical Reconstruction of Groundwater Flow, Contaminant Fate and Transport, and Distribution of Drinking Water Within the Service Areas of the Hadnot Point and Holcomb Boulevard Water Treatment Plants and Vicinities, U.S. Marine Corps Base Camp Lejeune, North Carolina, Chapter A: Summary and Findings. Appendix A7.

Maslia ML, Aral MM, Ruckart PZ, Bove FJ. 2016. Reconstructing Historical VOC Concentrations in Drinking Water for Epidemiological Studies at a U.S. Military Base: Summary of Results. Water (Basel). 8(10):449. doi: 10.3390/w8100449. Epub 2016 Oct 13. PMID: 28868161; PMCID: PMC5580837.

Montain S, Ely M. 2010. Water requirements and soldier hydration. Santee WR, Friedl KE; eds. Borden Institute Monograph Series. https://ke.army.mil/bordeninstitute/other_pub/HydrationPDF.pdf.

NRC (National Research Council). 1981. US Committee on Indoor Pollutants. Indoor Pollutants. Washington (DC): National Academies Press (US). VI, Monitoring and Modeling of Indoor Air Pollution. Available from: https://www.ncbi.nlm.nih.gov/books/NBK234059/

NRC (National Research Council). 1983. *Risk Assessment in the Federal Government: Managing the Process*. Washington, DC: The National Academies Press. https://doi.org/10.17226/366.

Official U.S. Military Field Manual. 1980. Occupational and Environmental Health Prevention, Treatment and Control of Heat Injury. TB MED 507. NAVMED P-5052-5, AFP 160-1; July 1980. https://archive.org/details/50525

Official U.S. Military Final Report. 1982. Research and Development for Health and Environmental Hazard Assessment - Task Order 2. Problem Definition of R and D Requirements for Field Sanitation and Water Supply. Contract No. DAMD17-79-C-9139; September 1982. https://dn790008.ca.archive.org/0/items/DTIC_ADA122384/DTIC_ADA122384.pdf

Oregon Department of Environmental Quality Environmental Cleanup Program. 2010. Human Health Risk Assessment Guidance. October. https://www.oregon.gov/deq/filterdocs/humanhealthriskassessmentguidance.pdf

Phillips L, Moya J. The evolution of EPA's Exposure Factors Handbook and its future as an exposure assessment resource. J Expo Sci Environ Epidemiol. 2013 Jan-Feb;23(1):13-21. doi: 10.1038/jes.2012.77. Epub 2012 Jul 18. PMID: 22805985.

Ramirez-Andreotta MD, Brusseau ML, Beamer P, Maier RM. 2013. Home gardening near a mining site in an arsenic-endemic region of Arizona: assessing arsenic exposure dose and risk via ingestion of home garden vegetables, soils, and water. Sci Total Environ. 454-455:373-82. doi: 10.1016/j.scitotenv.2013.02.063.

Raychaudhuri S. 2008. Proceedings of the 2008 Winter Simulation Conference. SJ Mason, RR Hill, L Mönch, O Rose, T Jefferson, JW Fowler, eds. https://www.informs-sim.org/wsc08papers/012.pdf

Salhotra AM. 2011. Human health risk assessment for contaminated properties. Progress in Molecular Biology and Translational Science, Volume 112. ISSN 1877-1173. http://dx.doi.org/10.1016/B978-0-12-415813-9.00010-6

US Armed Forces Health Surveillance Branch Communications Team. 2021. Update: Exertional Hyponatremia, Active Component, U.S. Armed Forces, 2005–2020. 4/1/2021.

US Army Corps of Engineers CEMP-RT Washington, D.C. 20314-1000 Manual No. 200-1-4 31 December 2010 Environmental Quality Risk Assessment Handbook, Volume II: Environmental Evaluation. Department of the Army EM 200-1-4. https://perma.cc/KV4J-AQDM

US Army. 1983. Water consumption planning factors study. Directorate of Combat Developments, United States Army Quartermaster School, Fort Lee, VA. As cited in US EPA 2019.

US Army. 1999. Study Report: water consumption planning factors. Directorate of Combat Developments (Quartermaster). United States Army Combined Arms Support Command, Fort Lee, VA. As cited in US EPA 2019.

US EPA (US Environmental Protection Agency). 1989. Risk Assessment Guidance for Superfund Volume I Human Health Evaluation Manual (Part A). Office of Emergency and Remedial Response. EPA/540/1-89/002. https://www.epa.gov/sites/default/files/2015-09/documents/rags_a.pdf

US EPA (US Environmental Protection Agency). 1992a. Guidelines for Exposure Assessment. EPA/600/Z-92/001. May. https://rais.ornl.gov/documents/GUIDELINES_EXPOSURE_ASSESSMENT.pdf

US EPA (US Environmental Protection Agency). 1992b. Dermal Exposure Assessment: Principles and Applications. Office of Health and Environmental Assessment, Office of Research and Development. January.

US EPA (US Environmental Protection Agency). 2004. Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E, Supplemental Guidance for Dermal Risk Assessment). Final. EPA/540/R/99/005 OSWER 9285.7-02EP PB99-963312. July.

US EPA (US Environmental Protection Agency). 2005. Guidelines for Carcinogen Risk Assessment Risk Assessment Forum Washington, DC. EPA/630/P-03/001F March.

US EPA (US Environmental Protection Agency. 2009. Risk Assessment Guidance for Superfund. Volume I: Human Health Evaluation Manual (Part F, Supplemental Guidance for Inhalation Risk Assessment). Final. EPA-540-R-070-002. OSWER 9285.7-82. January.

US EPA (US Environmental Protection Agency). 2011. Exposure Factors Handbook: 2011 Edition. Washington, DC: National Center for Environmental Assessment. EPA/600/R-090/052A. https://perma.cc/P6D7-Z7LG

US EPA (US Environmental Protection Agency). 2019. Update for Chapter 3 of the Exposure Factors Handbook Ingestion of Water and Other Select Liquids. Washington, DC: National Center for Environmental Assessment. February. EPA/600/R-18/259F. https://www.epa.gov/sites/default/files/2019-02/documents/efh_-_chapter_3_update.pdf

US EPA (US Environmental Protection Agency). 2022. Conducting a Human Health Risk Assessment. https://www.epa.gov/risk/conducting-human-health-risk-assessment

Weed DL. 2007. The Nature and Necessity of Scientific Judgement, 15 J. L. & Pol'y. https://brooklynworks.brooklaw.edu/jlp/vol15/iss1/6

Wilczynski SM. 2017. Section IV. Professional Judgment. A Practical Guide to Finding Treatments That Work for People with Autism. ISBN 978-0-12-809480-8. Academic Press.

World Health Organization (WHO). 2020. Model documentation: WHO Household Multiple Emission Sources (HOMES) Model. https://www.who.int/publications/m/item/household-multiple-mission-sources-homes-model

APPENDIX 1: Curriculum Vitae for Judy S. LaKind, Ph.D.

Judy S. LaKind, Ph.D.
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Judy S. LaKind, Ph.D., President of LaKind Associates, LLC, and Adjunct Associate Professor, Department of Epidemiology and Public Health, University of Maryland School of Medicine is a health and environmental scientist with expertise in exposure science, assessment of human health risks, biomonitoring, scientific and technical analysis for regulatory support, and state-of-the-science and systematic reviews. She has managed a wide array of successful projects, with completion in a timely manner and within budget, and has organized and facilitated numerous workshops on a variety of scientific subjects. Dr. LaKind has spoken and published extensively on exposure- and risk-related issues, including children's exposures to environmental chemicals, the implications of uncertainty in the risk assessment process, data quality, use of environmental epidemiology research in public health decision-making, weighing potential risks and benefits related to chemical use, the presence of environmental chemicals in human milk, and time-dependence and distributional analysis of exposure. Dr. LaKind has evaluated the use of human health risk assessment in the development of water quality criteria, and has critically analyzed the environmental fate, behavior, and bioavailability of pollutants in the context of setting regulatory criteria. She has developed risk assessments for a variety of urban industrial sites, military bases, and firing ranges, and has utilized state-of-the-science models for estimating blood lead levels in adults and children.

Previously, Dr. LaKind was a geologist at the US EPA's Office of Federal Activities, where she was responsible for the evaluation of Environmental Impact Statements and legislative reports. Dr. LaKind has taught graduate level courses at The Johns Hopkins University and the University of Maryland in risk assessment and aquatic chemistry. Dr. LaKind is Insights Editor for *Environment International*. She also serves on the editorial board of the *Journal of Environmental Exposure Assessment* and is past Associate Editor for the *Journal of Exposure Science and Environmental Epidemiology* and past editorial board member of the *Journal of Toxicology and Environmental Health*.

Dr. LaKind is a Past President of the International Society of Exposure Science. She was a member of the Health Effects Institute Energy Research Committee and the Maryland Department of Health and Mental Hygiene Cancer Cluster Advisory Committee and was a Junior Councilor, Society of Toxicology's Exposure Specialty Section. She previously served on the Boards of the National Swimming Pool Foundation and the Coalition Against Childhood Lead Poisoning (with a term as president). She is a former member of Maryland's Children's Environmental Health and Protection Advisory Council, the Lead Poisoning Prevention Commission, the Maryland Pesticide Reporting and Information Workgroup, the HESI RISK21 Advisory Board, and the World Health Organization Survey Coordinating Committee for the WHO Global Survey of Human Milk for Persistent Organic Pollutants (POPs). Dr. LaKind also served on the Institute of Medicine Committee on Blue Water Navy Vietnam Veterans and Agent Orange Exposure and the US Environmental Protection Agency Science Advisory Board Panel on Perchlorate - Approaches for Deriving Maximum Contaminant Level Goals for Drinking Water.

Academic Appointments:

Fellow-by-Courtesy, The Johns Hopkins University, Department of Applied Mathematics and Statistics. February 2013 – present.

Adjunct Associate Professor, University of Maryland School of Medicine, Department of Epidemiology and Preventive Medicine, August 2003 – August 2008; August 2009 – October 2009. February 2012 – present.

Associate Professor, University of Maryland School of Medicine, Department of Epidemiology & Public Health, September 2008 – August 2009; November 2009-February 2012.

Part Time Instructor, College of Engineering & Information Technology at University of Maryland Baltimore County, January 2010 – June 2010.

Adjunct Associate Professor, University of Maryland School of Law, May 2003 – May 2004.

Adjunct Associate Professor, Penn State College of Medicine, Department of Pediatrics, Milton S. Hershey Medical Center, 2002 – 2016.

Education:

Ph.D.; The Johns Hopkins University; Geography and Environmental Engineering; 1988
M.S.; The University of Wisconsin, Madison; Geology; 1984
B.A.; The Johns Hopkins University; Earth and Planetary Sciences; 1982

Litigation Support Training, 1994
Project Manager Training, 1995
Mid-America Toxicology Course, 1995
Risk Communication, 1995
Hershey Medical College Investigator Certification for Protecting Human Subjects, 2004
CITI Course in the Protection of Human Research Subjects, 2014
CITI Course in Institutional/Signatory Official: Human Subject Research, 2022
CITI Course in Community-Engaged and Community-Based Participatory Research, 2022
CITI Course in The Protection of Human Subjects, 2022

Experience:

Human Health Risk Assessment/Product Stewardship – Developed distributional exposure analyses for body burdens of persistent organic chemicals in breastfed infants. Conducted site-specific, health-based risk assessments for urban industrial sites, military bases, and firing ranges, with emphasis on PAHs, heavy metals (including lead), and volatile organic compounds. Developed exposure scenarios, with appropriate assumptions and parameters, for on-site and off-site exposure pathways, including recreational scenarios. These assessments included determination of receptors-of-concern and the development of site-specific conceptual site models as per U.S. EPA criteria. Prepared risk assessments under Maryland's Voluntary Cleanup Program. Utilized state-of-the-science models for predicting blood lead levels in adults and children. Evaluated and utilized model developed by the American Water Works Association to predict disinfection by-product formation resulting from chlorination of drinking water for zebra mussel control. Managed the development of technical papers which utilized innovative methodologies to correlate reductions of atmospheric concentrations of lead, carbon monoxide, ozone, and air toxics with improvements in human

health. Performed literature research, prepared manuscripts and comments for the USEPA, and provided litigation and regulatory support in evaluation of toxicity and environmental impacts of ethylene glycol (EG), propylene glycol (PG), and EG and PG de-icing and anti-icing formulations.

Systematic Review: Published multiple medium- and chemical-specific systematic and critical reviews. Invited member of the Risk Of Bias In Non-randomized Studies of Exposures (ROBINS-E) Working Group and participated in the GRADE Guidance for Modelled Data Working Group. Developed instrument for assessing study quality as part of systematic review (Biomonitoring, Environmental Epidemiology, and Short-Lived Chemicals - BEES-C – instrument); approach is now used by the US Environmental Protection Agency.

Project Management – Over 30 years of project management experience with teams of scientists from both inside and outside the US; focus on team communication and meeting client expectations regarding deliverables, deadlines, and budget.

Scientific workshop/expert panel development - Developed, coordinated, and facilitated numerous expert panels and workshops on a wide range of topics including environmental chemicals in breast milk, interpretation and communication of biomonitoring data, neurodevelopmental function testing, exposure to disinfection byproducts in swimming pool environments and associated health effects, biomonitoring of chemicals with short physiologic half-lives, and disease cluster methodologies.

Criteria Development - Determined scientific issues associated with the use of bioconcentration factors for regulating hydrophobic organic chemicals (HOCs), including dioxin. Developed an alternative risk assessment formula for HOC criteria determination.

Litigation Support - Provided litigation support for pulp and paper industry counsel on issues associated with aquatic organism accumulation of dioxin. Provided seminars to pulp and paper industry counsel on dioxin bioaccumulation. Provided litigation support for chemical industry on relative toxicity and environmental fate of a group of widely used compounds. Completed Litigation Support training course.

Regulatory Review - As an invited member of the Washington State Department of Health/Department of Ecology Sediment Scientific Review Board, provided scientific evaluation of proposed method for development of marine sediment chemical criteria relative to human health. Provided regulatory review, update, and analysis of: Clean Water Act 304(l) listing and approval/disapproval process; EPA pulp and paper mill guidance documents; and states' development of dioxin water quality criteria, for the pulp and paper industry. Critiqued bioaccumulation section of EPA's Great Lakes Water Quality Initiative. Analyzed scientific basis for proposed particulate matter standard.

Lead - Former member of the Coalition Against Childhood Lead Poisoning (with a term as president) and the Maryland Lead Poisoning Prevention Commission. Managed and conducted risk assessments for sites with lead contamination. Evaluated potential for human health risks associated with lead exposure to soil, water, and air, at firing ranges, and at residential, urban, and industrial sites. Utilized state-of-the-science models for predicting blood lead levels in both adults and children and has explored the utility of these models for assessing blood lead levels in people exposed to lead-contaminated media on an episodic basis. Made presentations to the public and media on risks associated with exposure to lead and created risk communication documentation on childhood lead poisoning prevention, used by the Kennedy-Krieger Institute's Lead Poisoning Prevention Program and the Baltimore City Department of Health. Technical editor of HUD's Guidelines for the Evaluation and Control of Lead-Based Paint Hazards in Housing.

Document Review and Analysis - Conducted Record of Decision search and analysis for development of remediation strategy for mitigation of subsurface migration of DNAPL. Performed scientific review, analysis, and critique of a wide range of documents including: Environmental Impact statements associated with Federal Energy Regulatory Commission hydroelectric power projects, natural gas pipeline siting, dredging projects; legislative reports on the Arctic National Wildlife Refuge and offshore oil exploration near the Georges Bank; risk assessments on formaldehyde air emissions from a particleboard plant and aquatic organism contamination in the Sacramento River; Endangerment Assessment and RI/FS of sawmill and landfill Superfund site.

Risk Communication - Gave presentations to public and media on risks associated with exposure to lead. Created risk communication information on childhood lead poisoning prevention, including *Derek the Dinosaur's Coloring Book About Lead*, used by the Kennedy-Krieger Institute's Lead Poisoning Prevention Program and the Baltimore City Department of Health. Coloring book was also used by Lead Safe St. Louis where it was translated into Spanish, Bosnian, Somali, Dari, and Vietnamese. Assisted in the development of a decision support document and white paper outlining the health risks and benefits associated with continued use of MTBE in the U.S. Assisted in the development of a Risk Primer for a major trade association.

Teaching - University of Maryland School of Law: Environmental Law and Science. The Johns Hopkins University: graduate-level courses on aquatic chemistry and environmental risk assessment. University of Maryland Baltimore County: upper-level course on human health risk assessment.

Professional Affiliations:

American Public Health Association (APHA) (1999-2015)
Maryland Public Health Association (Board member, 2008-2009)
American Chemical Society, Environmental Division (ACS)
Int. Society for Children's Health and the Environment (ISCHE), Founding member (2009-2015)
International Society of Exposure Science (ISES)
Society for Risk Analysis (SRA)
Society of Toxicology (SOT)
SOT Exposure Specialty Section, founding member (2017-present)

Selected Publications:

Macey K, Lange R, Apel P, Poddalgoda D, Calafat AM, Kolossa-Gehring M, LaKind JS, Melnyk LJ, Nakayama SF, St-Amand A. 2025. Human biomonitoring health-based guidance values: A case study of the HB2GV Dashboard and DEHP. *International Journal of Hygiene and Environmental Health*. Vol 263. https://doi.org/10.1016/j.ijheh.2024.114490

Rule AM, Wagner FA, Negi N, Tajouoh-Daghuie CJ, Rosman L, Naiman J, Lange SS, Clougherty JE, Vorhees D, LaKind JS. Principles and elements for creating and sustaining successful public-private partnerships (PPPs) for environmental community monitoring programs: Results from a scoping review and interviews. *Environmental Justice* doi.org/10.1089/env.2024.002

Higgins JPT, Morgan RL, Rooney AA, Taylor KW, Thayer KA, Silva RA, Lemeris C, Akl EA, Bateson TF, Berkman ND, Glenn BS, Hróbjartsson A, LaKind JS, McAleenan A, Meerpohl JJ, Nachman RM, Obbagy JE, O'Connor A, Radke EG, Savović J, Schünemann HJ, Shea B, Tilling K, Verbeek J, Viswanathan M, Sterne JAC. 2024. A tool to assess risk of bias in non-randomized follow-up studies of exposure effects (ROBINS-E). *Environment International* 186:108602.

Ebelt S, Baxter L, Erickson HS, Henneman LRF, Lange S, Luben TJ, Neidell M, Rule AM , Russell AG, Wendt Hess J, Burns CJ, LaKind JS, Goodman JE. 2023. Air pollution accountability research: Moving from a chain to a web. *Global Epidemiology* 100128. ISSN 2590-1133 https://doi.org/10.1016/j.gloepi.2023.100128.

LaKind JS, Burns CJ, Johnson GT, Lange SS. 2023. Epidemiology for risk assessment: US EPA guidance and the Matrix. *Hygiene and Environmental Health Advances* 106:100059. https://doi.org/10.1016/j.heha.2023.100059

LaKind JS. 2023. Invited Perspective. PFAS and infant nutrition: Why aren't we monitoring? *Environmental Health Perspectives* 131(3): https://doi.org/10.1289/EHP12134.

LaKind JS, Naiman J, Verner M-A, Lévêque L, Fenton S. 2023. Per- and polyfluoroalkyl substances (PFAS) in breast milk and infant formula: A global issue. *Environmental Research* 219:115042.

Nakayama SF, St-Amand A, Pollock T, Ashley-Martin J, Bamai YA, Barr DB, Bessems J, Calafat A, Castaño A, Covaci A, Duca RC, Faure S, Galea KS, Hays S, Hopf NB, Ito Y, Jeddi MZ, Kolossa-Gehring M, Kumar E, LaKind JS, López ME, Louro H, Makris KC, Melnyk L, Naiman J, Nassif J, Noisel N, Quirós-Alcalá L, Rafiee A, Rambaud L, Silva MJ, Ueyama J, Verner M-A, Waras MN, Werry K. 2023. Interpreting Biomonitoring Data: Introducing the i-HBM Working Group's Guidance Value Dashboard. *International Journal of Hygiene and Environmental Health* 247:114046.

Wilson AM, Mussio I, Chilton S, Gerald LB, Jones RA, Drews FA, LaKind JS, Beamer PI. 2022. A novel application of risk–risk tradeoffs in occupational health: Nurses' occupational asthma and infection risk perceptions related to cleaning and disinfection during COVID-19. *International Journal of Environmental Research and Public Health 19*(23):16092 https://doi.org/10.3390/ijerph192316092

LaKind JS, Burns CJ, Donald R. Mattison DR. 2022. Commentary: Systematic reviews and observational epidemiology: the more things change… *Global Epidemiology* 4:100088.

LaKind JS, Burns CJ, Naiman DQ. 2,4-D and NHANES: Sources of exposure and identification of data gaps. *Hygiene and Environmental Health Advances* 4:100023.

Burns CJ, LaKind JS, Naiman J, Boon D, Clougherty JE, Rule AM, Zidek A. Research on COVID-19 and air pollution: A path towards advancing the science. *Environmental Research* 212, Part A:113240. https://www.sciencedirect.com/science/article/pii/S0013935122005679?via%3Dihub

LaKind JS, Verner M-A, Rogers R, Goeden H, Naiman DQ, Marchitti S, Lehmann G, Hines E, Fenton SE. 2022. Current breast milk PFAS levels in North America: After all this time why don't we know more? *Environmental Health Perspectives* 130(2): 25002. https://doi.org/10.1289/EHP10359

Burns CJ, LaKind JS. 2022. Elements to increase translation in pyrethroid epidemiology research: a review. *Science of the Total Environment* 813:152568 doi: 10.1016/j.scitotenv.2021.152568.

Burns CJ, LaKind JS. 2021. Using the Matrix to bridge the epidemiology/risk assessment gap: A case study of 2,4-D. *Critical Reviews in Toxicology* 51(7):591-599. doi: 10.1080/10408444.2021.1997911

LaKind JS, Burns CJ, Pottenger LH, Naiman DQ, Goodman JE, Marchitti SA. 2021. Does ozone inhalation cause adverse metabolic effects in humans? A systematic review. *Critical Reviews in Toxicology* 51(6):467-508. DOI: 10.1080/10408444.2021.1965086

Verner M-A, Salame H, Housand C, Birnbaum LS, Bouchard M, Chevrier J, Aylward L, Naiman DQ, LaKind JS. 2020. How many urine samples are needed to accurately assess exposure to non-persistent chemicals? The Biomarker Reliability Assessment Tool (BRAT) for scientists, research sponsors and risk managers. *International Journal of Environmental Research and Public Health* 17:9102; doi:10.3390/ijerph17239102.

Brozek J, Canelo C, Akl E, Bowen J, Bucher J, Chiu W, Cronin M, Djulbegovic B, Falavigna M, Guyatt G, Gordon A, Hilton Boon M, Hutubessy R, Joore M, Katikireddi S, LaKind J, Langendam M, Manja V, Magnuson K, Mathioudakis A, Meerpohl J, Mertz D, Mezencev R, Morgan R, Morgano GP, Mustafa R, Naidoo B, O'Flaherty M, Patlewicz G, Riva J, Posso M, Ringborg A, Rooney A, Schlosser P, Schwartz L, Shemilt I, Tarride J-E, Thayer K, Tsaioun K, Vale L, Wambaugh J, Wignall JA, Williams AR, Xie F, Zhang Y, Schünemann H. 2020. GRADE Guidelines 30: The GRADE Approach to Assessing the Certainty of Evidence from Models – an Overview in the Context of Health Decision-making. *Journal of Clinical Epidemiology* S0895-4356(20)31103-3.

Goodman M, Li J, Flanders WD, Mahood D, Anthony LG, Zhang Q, LaKind JS. 2020. Epidemiology of PCBs and neurodevelopment: Systematic assessment of multiplicity and completeness of reporting. *Global Epidemiology* 2:100040.

LaKind JS. 2020. Foreward. Total Exposure Health: An Introduction. Editors: Phillips KA, Yamamoto DP, Racz L. CRC Press.

LaKind JS, Naiman J, Burns CJ. Translation of exposure and epidemiology for risk assessment: A shifting paradigm. 2020. *International Journal of Environmental Research and Public Health 17*(12):4220; https://doi.org/10.3390/ijerph17124220.

LaKind JS, Burns CJ, Erickson H, Graham SE, Jenkins S, Johnson GT. 2020. Bridging the epidemiology risk assessment gap: An NO_2 case study of the Matrix. *Global Epidemiology* 2:100017.

LaKind JS, Goodman M. 2019. Methodological evaluation of human research on asthmagenicity and occupational cleaning: A case study of quaternary ammonium compounds ("quats"). *Allergy, Asthma & Clinical Immunology* 15:69.

Burns CJ, LaKind JS, Mattison DR, Alcala CS, Branch F, Castillo J, Clark A, Clougherty JE, Darney SP, Erickson H, Goodman M, Greiner M, Jurek AM, Miller A, Rooney AA, Zidek A. 2019. A Matrix for bridging the epidemiology and risk assessment gap. *Global Epidemiology* 1: 100005.

LaKind JS, Pollock T, Naiman DQ, Kim S, Nagasawa A, Clarke J. 2019. Factors affecting interpretation of national biomonitoring data from multiple countries: BPA as a case study. *Environmental Research* 173:318-329. PMID: 30951958

LaKind JS, O'Mahony C, Armstrong T, Tibaldi R, Blount BC, Naiman DQ. 2019. ExpoQual: Evaluating measured and modeled human exposure data. *Environmental Research* 171:302–312.

LaKind JS, Idri F, Naiman DQ, Verner M-A. 2019. Biomonitoring and nonpersistent chemicals – understanding and addressing variability and exposure misclassification. *Current Environmental Health Reports* 6(1):16-21.

LaKind JS, Davis M, Lehmann GM, Hines E, Marchitti SA, Alcala C, Lorber M. 2019. Infant dietary exposures to environmental chemicals and infant/child health: A critical assessment of the literature. *Environmental Health Perspectives* 126(9):96002. Highlighted in: Arnold C. 2019. Baby steps forward: Recommendations for better understanding environmental chemicals in breast milk and infant formula. *Environmental Health Perspectives* https://doi.org/10.1289/EHP4804

Lehmann GM, LaKind JS, Davis M, Hines E, Marchitti SA, Alcala C, Lorber M. 2019. Environmental chemicals in breast milk and formula: Exposure and risk assessment implications. *Environmental Health Perspectives* 126(9):96001. Highlighted in: Arnold C. 2019. Baby steps forward: Recommendations for better understanding environmental chemicals in breast milk and infant formula. *Environmental Health Perspectives* https://doi.org/10.1289/EHP4804

Metwally DE, Chain K, Stefanak MP, Alwis U, Blount BC, LaKind JS, Bearer CF. 2018. Urinary metabolites of volatile organic compounds of infants in the neonatal intensive care unit. *Pediatric Research* 83(6):1158-1164.

Goodman M, Naiman DQ, LaKind JS. 2018. Systematic review of the literature on triclosan and health outcomes in humans. *Critical Reviews in Toxicology* 48(1):1-51.

LaKind JS, Burns CJ, Naiman DQ, O'Mahony C, Vilone G, Burns AJ, Naiman JS. 2017. Critical and systematic evaluation of data for estimating human exposures to 2,4-dichlorophenoxyacetic acid (2,4-D) - quality and generalizability. *Journal of Toxicology and Environmental Health, Part B*. 20(8):423-446.

LaKind JS, Anthony LG, Goodman M. 2017. Review of reviews on exposures to synthetic organic chemicals and children's neurodevelopment: Methodological and interpretation challenges. *Journal of Toxicology and Environmental Health*, Part B 20(8):390-422.

Vincent MJ, Bernstein JA, Basketter D, LaKind JS, Dotson GS, Maier A. 2017. Chemical-induced asthma and the role of clinical, toxicological, exposure and epidemiological research in regulatory and hazard characterization approaches. *Regulatory Toxicology & Pharmacology* 90: 126-132.

LaKind JS, Overpeck J, Breysse PN, Backer L, Richardson S, Sobus J, Sapkota A, Upperman CR, Jiang C, Beard CB, Brunkard JM, Bell J, Harris R, Chretien J-P, Peltier RE, Chew GL, Blount B. 2016. Exposure science in an age of rapidly changing climate: Challenges and opportunities. *Journal of Exposure Science and Environmental Epidemiology* 26: 529-538.

Beck NB, Becker RA, Erraguntla N, Farland WH, Grant RL, Gray G, Kirman C, LaKind JS, Lewis RJ, Nance P, Pottenger LH, Santos SL, Shirley S, Simon T, Dourson ML. 2016. Approaches for describing and communicating overall uncertainty in toxicity characterizations: U.S. Environmental Protection Agency's Integrated Risk Information System (IRIS) as a case study. *Environment International* 89–90:110–128.

Weldon RH, LaKind JS. 2015. Biomonitoring of dioxins and furans: Levels and trends in humans. In: The Handbook of Environmental Chemistry. ISSN 1867-979X. Springer:Berlin Heidelberg. 23 pp. doi: 10.1007/698_2015_433. http://link.springer.com/chapter/10.1007/698_2015_433

LaKind JS, Goodman M, Makris SL, Mattison DR. 2015. Improving concordance in environmental epidemiology: A three-part proposal. *Journal of Toxicology & Environmental Health, Part B*. 18(2):105-120.

LaKind JS, Naiman, DQ. 2015. Temporal trends in bisphenol A exposure in the United States from 2003–2012 and factors associated with BPA exposure: Spot samples and urine dilution complicate data interpretation. *Environmental Research* 142:84–95.

LaKind JS, Goodman M, Barr DB, Weisel CP, Schoeters G. 2015. Lessons learned from the application of BEES-C: Systematic assessment of study quality of epidemiologic research on BPA, neurodevelopment, and respiratory health. *Environment International* 80:41-71.

LaKind JS, Sobus JR, Goodman M, Barr DB, Fürst P, Albertini RJ, Arbuckle TE, Schoeters G, Tan Y-M, Teeguarden J, Tornero-Velez R, Weisel CP. 2014. A proposal for assessing study quality: Biomonitoring, Environmental Epidemiology, and Short-Lived Chemicals (BEES-C) Instrument. *Environment International* 73C:195-207.

Mattison DR, Karyakina N, Goodman M, LaKind JS. 2014. Pharmacokinetics of selected exogenous and endogenous estrogens: A review of the data and identification of knowledge gaps. *Critical Reviews in Toxicology* 44(8):696-724.

Lehmann GM, Verner MA, Luukinen B, Henning C, Assimon SA, LaKind JS, McLanahan ED, Phillips LJ, Davis MH, Powers CM, Hines EP, Haddad S, Longnecker MP, Poulsen MT, Farrer DG, Marchitti SA, Tan YM, Swartout JC, Sagiv SK, Welsh C, Campbell JL Jr, Foster WG, Yang RS, Fenton SE, Tornero-Velez R, Francis BM, Barnett JB, El-Masri HA, Simmons JE. 2014. Improving the risk assessment of lipophilic persistent environmental chemicals in breast milk. *Critical Reviews in Toxicology* 44(7):600-17.

LaKind JS, Goodman M, Mattison DR. 2014. Bisphenol A and indicators of obesity, glucose metabolism/type 2 diabetes and cardiovascular disease: A systematic review of epidemiologic research. *Critical Reviews in Toxicology* 44(2):121–150.

Goodman M, LaKind JS, Mattison DR. 2014. Do phthalates act as obesogens in humans? A systematic review of the epidemiology literature. *Critical Reviews in Toxicology* 44(2):151–175.

Marchitti SA, Hines EP, LaKind JS, Berlin CM Jr., Fenton SE, Kenneke JF. 2013. Environmental Chemicals in Breast Milk. Reference Module in Earth Systems and Environmental Sciences. Elsevier. http://dx.doi.org/10.1016/B978-0-12-409548-9.02139-4

Goodman M, LaKind JS, Fagliano JA, Lash TL, Wiemels JL, Winn DM, Patel C, VanEenwyk J, Kohler BA, Schisterman EF, Albert P, Mattison DR. 2014. Cancer cluster investigations: Review of the past and proposals for the future. *International Journal of Environmental Research and Public Health* 11:1479-1499; doi:10.3390/ijerph110201479.

Marchitti SA, LaKind JS, Naiman DQ, Berlin CM, Kenneke JF. 2013. Improving infant exposure and health risk estimates: Using serum data to predict polybrominated diphenyl ether concentrations in breast milk. *Environmental Science & Technology* 47:4787−4795.

LaKind JS, Goodman M, Naiman DQ. 2012. Use of NHANES data to link chemical exposures to chronic diseases: a cautionary tale. *PLoS ONE* 7(12):e51086. doi:10.1371/journal.pone.0051086

LaKind J. 2013. Can coatings for foods and beverages: Issues and options. *International Journal of Technology, Policy and Management.* 13(1):80-95.

Dórea JG, Fenton SE, LaKind JS, Berlin CM Jr. 2012. Researching chemicals in human milk can be conducted without discouraging breastfeeding. *Bosnian Journal of Basic Medical Sciences* 12(2):137-138.

Goodman M, Naiman JS, Goodman D, LaKind JS. 2013. Response to Condon et al. comments on "Cancer clusters in the USA: What do the last twenty years of state and federal investigations tell us?" *Critical Reviews in Toxicology* 43(1)75-76.

Goodman M, Naiman JS, Goodman D, LaKind JS. Cancer clusters in the US – What do the last twenty years of state and federal investigations tell us? 2012. *Critical Reviews in Toxicology* 42(6):474-490. *Number 50 of "Most read articles in Critical Reviews in Toxicology"

LaKind JS, Levesque J, Dumas P, Bryan S, Clarke J, Naiman DQ. 2012. Comparing United States and Canadian population exposures from national biomonitoring surveys: Bisphenol A intake as a case study. *Journal of Exposure Science and Environmental Epidemiology* 22:219-226.

Youngstrom E, Kenworthy L, Lipkin PH, Goodman M, Squibb K, Mattison DR, Anthony LG, Makris SL, Bale A, Raffaele KC, LaKind JS. 2011. A proposal to facilitate weight-of-evidence assessments: Harmonization of neurodevelopmental environmental epidemiology studies (HONEES). *Neurotoxicology and Teratology*. 33:354–359.

LaKind JS, Berlin CM Jr., Fenton SE. 2011. Environmental Chemicals in Breast Milk. In: Nriagu, J.O. (ed.) *Encyclopedia of Environmental Health*, volume 2, pp. 347–356 Burlington: Elsevier.

Blount BC, Backer LC, Aylward LL, Hays SM, LaKind JS. 2011. Human Exposure Assessment for DBPs: Factors Influencing Blood Trihalomethane Levels. In: Nriagu, J.O. (ed.) *Encyclopedia of Environmental Health*, volume 3, pp. 100–107 Burlington: Elsevier.

LaKind JS, Naiman DQ. 2011. Daily intake of bisphenol A (BPA) and potential sources of exposure – 2005-2006 NHANES. *Journal of Exposure Science and Environmental Epidemiology* 21:272-279.

Blount BC, McElprang DO, Chambers DM, Waterhouse MG, Squibb KS, LaKind JS. 2010. Methodology for collecting, storing, and analyzing human milk for volatile organic compounds. *Journal of Environmental Monitoring* 12:1265-1273.

Goodman M, Youngstrom E, Gutermuth Anthony L, Kenworthy L, Lipkin PH, Squibb K, Mattison DR, LaKind JS. 2010. Using systematic reviews and meta-analyses to support regulatory decision-making for neurotoxicants: Lessons learned from a case study of PCBs. *Environmental Health Perspectives* 118:727-734. doi:10.1289/ehp.0901835.

LaKind JS, Richardson SD, Blount BC. The good, the bad, and the volatile – Can we have both healthy pools and healthy people? 2010. *Environmental Science & Technology* 44:3205–3210.

Youngstrom E, LaKind JS, Kenworthy L, Lipkin PH, Goodman M, Squibb K, Mattison DR, Anthony BJ, Gutermuth Anthony L. 2010. Advancing the selection of neurodevelopmental measures in epidemiological studies of environmental chemical exposure and health effects. *International Journal of Environmental Research and Public Health* 7:229-268.

LaKind JS, Birnbaum LS. 2010. Out of the frying pan AND out of the fire: the indispensable role of exposure science in assessing replacement chemicals. *Journal of Exposure Science and Environmental Epidemiology* 20:115–116.

LaKind JS, Naiman DQ, Hays SM, Aylward LL, Blount BC. 2010. Public health interpretation of trihalomethane blood levels in the U.S.: NHANES 1999-2004. *Journal of Exposure Science and Environmental Epidemiology* 20(3):255-262. Featured article.

LaKind JS, Berlin CM Jr., Sjödin A, Turner W, Wang RY, Needham LL, Paul IM, Stokes JL, Naiman DQ, Patterson DG Jr. 2009. Do human milk concentrations of persistent organic chemicals really decline during lactation? Chemical concentrations during lactation and milk/serum partitioning. *Environmental Health Perspectives* 117(10):1625–1631.

LaKind JS, Berlin CM Jr., Mattison DR. 2009. Response to Geraghty et al. Letter to the Editor. *Breastfeeding Medicine* 4(2):127.

LaKind JS, Fenton SE, Dórea JG. 2009. Human milk biomonitoring of phthalates: Expanding our understanding of infant exposure is compatible with supporting breastfeeding. Letter to the editor. *Environment International* 35:994-995.

LaKind JS, Hays SM, Aylward LL, Naiman DQ. 2009. Perspective on serum dioxin levels in the United States: An evaluation of the NHANES data. *Journal of Exposure Science and Environmental Epidemiology* 19:435-441.

Weisel CP, Richardson SD, Nemery B, Aggazzotti G, Baraldi E, Blatchley ER III, Blount BC, Carlsen K-H, Eggleston PA, Frimmel FH, Goodman M, Gordon G, Grinshpun S, Heederik DJJ, Kogevinas M, LaKind JS, Nieuwenhuijsen MJ, Piper FC, Sattar SA. 2009. Conclusions and key research recommendations from the Workshop on Advancing the Science: Childhood Asthma and Environmental Exposures at Swimming Pools. *Environmental Health Perspectives* 117:500–507. doi:10.1289/ehp.11513. Science Selections: Widening the Pool of Factors Studies Needed to Assess Asthma–Swimming Link p. A162.

LaKind JS, Berlin CM Jr., Mattison DR. 2008. The heart of the matter on breast milk and environmental chemicals: Essential points for health care providers and new parents. *Breastfeeding Medicine* 4(3):251-259. New York Times summary: Despite Worries Over Toxins, Breast-Feeding Still Best for Infants. http://www.nytimes.com/2008/12/20/health/research/20breast.html?scp=6&sq=lakind&st=cse

LaKind JS, Naiman DQ. 2008. Bisphenol A (BPA) Daily intakes in the United States: Estimates from the 2003-2004 NHANES urinary BPA data. *Journal of Exposure Science and Environmental Epidemiology* 18:608–615. Featured Article.

Aylward LL, LaKind JS, Hays SM. 2008. Derivation of biomonitoring equivalent (BE) values for 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD) and related compounds: a screening tool for interpretation of biomonitoring data in a risk assessment context. *Journal of Toxicology and Environmental Health A* 71(22):1499-1508.

Aylward LL, LaKind JS, Hays SM. 2008. Biomonitoring Equivalents (BEs) for interpretation of whole blood biomonitoring data for trihalomethanes. *Regulatory Toxicology and Pharmacology* 51(3, Suppl 1):S68-S77.

Hays SM, Aylward LL, LaKind JS, Bartels MJ, Barton HA, Boogaard PJ, Brunk C, DiZio S, Dourson M, Goldstein DA, Lipscomb J, Kilpatrick ME, Krewski D, Krishnan K, Nordberg M, Okino M, Tan Y-M, Viau C, Yager JW. 2008. Guidelines for the Derivation of Biomonitoring Equivalents: Report from the Biomonitoring Equivalents Expert Workshop. *Regulatory Toxicology and Pharmacology* 51(3, Suppl 1):S4-S15.

LaKind JS, Aylward LL, Brunk C, DiZio S, Dourson M, Goldstein DA, Kilpatrick ME, Krewski D, Bartels M, Barton HA, Boogaard PJ, Lipscomb J, Krishnan K, Nordberg M, Okino M, Tan Y-M, Viau C, Yager JW,

Hays SM. 2008. Guidelines for the Communication of Biomonitoring Equivalents: Report from the Biomonitoring Equivalents Expert Workshop. *Regulatory Toxicology and Pharmacology* 51(3, Suppl 1):S16-S26.

Hays SM, Aylward LL, LaKind JS. 2008. Introduction to the Biomonitoring Equivalents Pilot Project: Development of guidelines for the derivation and communication of Biomonitoring Equivalents. *Regulatory Toxicology and Pharmacology* 51(3, Suppl 1):S1-S2.

LaKind JS, Berlin CM Jr., Stokes JL, Naiman DQ, Paul IM, Patterson DG Jr., Jones RS, Niehüser S., Park A, Wang RY, Needham LL, Lorber MN, Sjödin A. 2008. Lifestyle and polybrominated diphenyl ethers (PBDEs) in human milk in the United States: A pilot study. *Toxicological & Environmental Chemistry* 70(9):61-63.

LaKind JS, Barraj L, Tran N, Aylward LL. 2008. Guest Commentary: Environmental chemicals in people: Challenges in interpreting biomonitoring information. *Journal of Environmental Health* 70(9):61-64.

LaKind JS, Holgate ST, Ownby DR, Mansur AH, Helms PJ, Pyatt D, Hays SM. 2008. Authors' response to 'Critical review of Clara cell protein: sound science?' *Biomarkers* 13(3):244-245.

LaKind JS, Holgate ST, Ownby DR, Mansur AH, Helms PJ, Pyatt D, Hays SM. 2007. A critical review of the use of Clara Cell secretory protein CC16 as a biomarker of exposure and effect. *Biomarkers* http://www.informaworld.com/smpp/title~content=t713693137~db=all~tab=issueslist~branches=12 - v1212:445-467.

LaKind JS, Wilkins AA, Bates MN. 2007. Breast biomonitoring and environmental chemicals: Use of breast tissues and fluids in breast cancer etiologic research. *Journal of Exposure Science and Environmental Epidemiology* 17:525-540.

LaKind JS. 2007. Recent global trends and physiologic origins of dioxins and furans in human milk. *Journal of Exposure Science and Environmental Epidemiology* 17:510–524. Featured Article.

Needham LL, Naiman DQ, Patterson DG, Jr., LaKind JS. 2007. Assigning concentration values for dioxin and furan congeners in human serum when measurements are below limits of detection: an observational approach. *Chemosphere* 67:439–447.

Amler RW, Barone S Jr., Belger **A,** Berlin CM Jr., Cox C, Frank H, Goodman M, Harry J, Hooper SR, Ladda R, LaKind JS, Lipkin PH, Lipsitt LP, Lorber MN, Myers G, Mason AM, Needham LL, Sonawane B, Wachs TD, Yager JW. 2006. Hershey Medical Center Technical Workshop Report: Optimizing the design and interpretation of epidemiologic studies for assessing neurodevelopmental effects from *in utero* chemical exposure. *NeuroToxicology* 27(5):861-874.

Baier-Anderson C, Blount B, LaKind JS, Naiman DQ, Wilbur SB, Tan S. 2006. Estimates of exposure to perchlorate from consumption of human milk, dairy milk, and water and comparison to current reference dose. *Journal of Toxicology and Environmental Health* 69(4):319-330.

Sjödin A, LaKind JS, Patterson D, Needham LL, Wang R, Berlin CM Jr., Paul IM, Stokes J. 2005. Current concentrations and changes in concentrations of PBDEs, persistent pesticides, and PCBs in human milk. *Organohalogen Compounds* 67:1745-1748.

Bates MN, Hamilton JW, LaKind JS, Langenberg P, O'Malley M, Snodgrass W. 2005. Biomonitoring Workshop Report: Biomonitoring study design, interpretation, and communication - Lessons learned and path forward. *Environmental Health Perspectives* 113:1615–1621.

LaKind JS, Brent RL, Dourson ML, Kacew S, Koren G, Sonawane B, Tarzian AJ, Uhl K. 2005. Human milk biomonitoring data: interpretation and risk assessment issues. *Journal of Toxicology and Environmental Health* 68(20):1713-1770.

Fenton SE, Condon M, Ettinger A, Mason A, McDiarmid M, Qian Z, Selevan SG, LaKind JS. 2005. Collection and use of exposure data from human milk biomonitoring in the Unites States. *Journal of Toxicology and Environmental Health* 68(20):1691-1712.

Berlin CM Jr., LaKind JS, Fenton SE, Wang RY, Bates MN, Brent RL, Condon M, Crase BL, Dourson ML, Ettinger AS, Foos B, Fürst P, Giacoia GP, Goldstein DA, Haynes SG, Hench KD, Kacew S, Koren G, Lawrence RA, Mason A, McDiarmid MA, Moy G, Needham LL, Paul IM, Pugh LC, Qian Z, Salamone L, Selevan SG, Sonawane B, Tarzian AJ, Tully MR, Uhl K. 2005. Conclusions and recommendations of the Expert Panel: Technical Workshop on Human Milk Surveillance and Biomonitoring for Environmental Chemicals in the United States. *Journal of Toxicology and Environmental Health* 68(20):1825-1831.

Berlin CM Jr, Crase BL, Fürst P, Moy G, Needham LL, Pugh L, Tully MR, LaKind JS. 2005. Methodologic considerations for improving and facilitating human milk research. *Journal of Toxicology and Environmental Health* 68(20):1803-1824.

LaKind JS, Berlin CM Jr., Bates MN. 2005. Overview: Technical Workshop on Human Milk Surveillance and Biomonitoring for Environmental Chemicals in The United States. *Journal of Toxicology and Environmental Health* 68(20):1683-1690.

LaKind JS, Wilkins AA, Berlin CM. 2004. Environmental chemicals in human milk: a review of concentrations, determinants, infant exposures and health, and guidance for future research. *Toxicology and Applied Pharmacology* 198:184-208.

Berlin CM, LaKind JS, Selevan SG. 2003. Human Milk Monitoring for Environmental Chemicals: Guidance for Future Research. ABM News and Views. 9(17):23-24.

LaKind JS, Bates MN, Wilkins AA. 2003. How useful is measurement of environmental chemicals in human milk in investigations of breast cancer etiology? *Organohalogen Compounds* 65:346-349. Presented at Dioxin2003, Boston, MA. August.

LaKind JS. 2003. Interpreting and Communicating Health and Risk Information on Chemicals in Breast Milk: DDT as a Case Study. In: Reviews in Food and Nutrition, Volume 1. V.R. Preedy and R.R. Watson, eds. Taylor & Francis. Pp. 230-242.

Borgert CJ, LaKind JS, Witorsch RJ. 2003. A critical review of potential methods for comparing hormonal activity of endogenous and exogenous chemicals in human milk and infant formula. *Environmental Health Perspectives* 111(8):1020-1036.

Aylward LL, Hays SM, LaKind JS, Ryan JJ. 2003. Partitioning of persistent lipophilic compounds including dioxins between human milk lipid and blood lipid: An initial assessment. *Journal of Toxicology and Environmental Health* 66(1):1-5.

Berlin CM, Kacew M, Lawrence R, LaKind JS, Campbell R. 2002. Criteria for chemical selection for programs on human milk surveillance and research for environmental chemicals. *Journal of Toxicology and Environmental Health* 65(22):1839-1852.

LaKind JS, Berlin CM. 2002. Technical Workshop on Human Milk Surveillance and Research on Environmental Chemicals in the United States: An Overview. *Journal of Toxicology and Environmental Health* 65(22):1829-1838.

LaKind JS, Birnbach N, Borgert CJ, Sonawane BR, Tully MR, Friedman L. 2002. Human milk surveillance and research of environmental chemicals: Concepts for consideration in interpreting and presenting study results. *Journal of Toxicology and Environmental Health* 65(22):1909-1928.

Berlin CM, LaKind JS, Sonawane BR, Kacew S, Borgert CJ, Bates MN, Birnbach N, Campbell R, Dermer A, Dewey KG, Ellerbee SM, Fürst P, Giacoia GP, Gartner L, Groer M, Haynes SG, Humenick SS, Lawrence RA, Lorber M, Lovelady C, Mason A, Needham LL, Picciano MF, Plautz J, Ryan JJ, Selevan SG, Sumaya CV, Tully MR, Uhl K, Vesell E, Wilson JT. 2002. Conclusions, research needs and recommendations of the expert panel: technical workshop on human milk surveillance and research for environmental chemicals in the United States. *Journal of Toxicology and Environmental Health* 65(22):1929-1935.

LaKind JS, Berlin C, Naiman DQ. 2001. Infant exposure to chemicals in breast milk in the United States: What we need to learn from a breast milk monitoring program. *Environmental Health Perspectives* 109:75-88.

LaKind JS, Berlin CM. 2000. PDBEs in breast milk: Where do we go from here? Presented at Dioxin2000, Monterey, California, August 13-17, 2000. *Organohalogen Compounds* 47:241-244.

Pandak CA, Mason AM, LaKind JS. 2000. Tools for improving community health: Community health assessment and asset evaluation. *Environmental Regulation and Permitting*. 9(3):15-23.

LaKind JS, Berlin CM, Naiman DQ, Park CN. 2000. Characterization of dose distributions of selected breast milk contaminants to nursing infants: DDE and TCDD. *Organohalogen Compounds* 48:223-226.

LaKind JS, Berlin C, Park C, Naiman DQ, Gudka NJ. 2000. Methodology for characterizing distributions of incremental body burdens of 2,3,7,8-TCDD and DDE from breast milk in North American nursing infants. *Journal of Toxicology and Environmental Health, Part A* 59:605-639.

Peddicord RK, LaKind JS. 2000. Ecological and human health risks at an outdoor firing range. *Environmental Toxicology & Chemistry* 19(10):2602–2613.

LaKind JS, Filser JG. 1999. Dietary exposure to PCBs and dioxins. *Environmental Health Perspectives* 107(10):A9.

LaKind JS, McKenna EA, Hubner RP, Tardiff RG. 1999. A review of the comparative mammalian toxicity of ethylene glycol and propylene glycol. *Critical Reviews in Toxicology* 29(4):331-365.

LaKind JS, Ginevan ME, Naiman DQ, James AC, Jenkins RA, Dourson ML, Felter SP, Graves CG, Tardiff RG. 1999. Distribution of exposure concentrations and doses for constituents of environmental tobacco smoke. *Risk Analysis: An International Journal* 19(3):375-390.

LaKind JS, Jenkins RA, Naiman DQ, Ginevan ME, Graves CG, Tardiff RG. 1999. Use of environmental tobacco smoke (ETS) constituents as markers for ETS exposure. *Risk Analysis: An International Journal* 19(3):359-376.

LaKind JS, Graves CG, Ginevan ME, Jenkins RA, Naiman DQ, Tardiff RG. 1999. Exposure to environmental tobacco smoke in the workplace and the impact of away-from-work exposure. *Risk Analysis: An International Journal* 19(3):349-358.

LaKind JS. 1998. Comparison of three models for predicting blood lead levels in children: Episodic exposures to lead. *Journal of Exposure Analysis and Environmental Epidemiology* 8(3):399-406.

LaKind JS. 1998. Forward to Special Issue on Environmental Risk Assessment: Issues and Methods. *International Journal of Environment and Pollution* 9(1):1-2.

Bartell SM, LaKind JS, Moore JA, Anderson P. 1998. Bioaccumulation of hydrophobic organic chemicals by aquatic organisms: A workshop summary. *International Journal of Environment and Pollution* 9(1):3-25.

LaKind JS. 1994. Sediment dioxin levels as the basis for risk assessment and human health criteria. *International Journal of Environment and Pollution* 3(4):226-232.

LaKind JS, Naiman DQ. 1993. Comparison of predicted and observed dioxin levels in fish: Implications for risk assessment. *Risk: Issues in Health & Safety* 4(3):253-262.

Rifkin E, LaKind JS. 1991. Dioxin bioaccumulation: Key to a sound risk assessment methodology. *Journal of Toxicology and Environmental Health* 33:103-112.

LaKind J. 1991. Bioconcentration – letter to the editor. *Environmental Science & Technology* 25(1):6.

LaKind JS, Rifkin E. 1990. Current method for setting dioxin limits in water requires reexamination. *Environmental Science & Technology*. 24:963-965.

LaKind JS, Stone AT. 1989. Reductive Dissolution of Goethite by Phenolic Reductants. *Geochimica Cosmochimica Acta*. 53:961-971

LaKind JS, Stone AT. 1988. Reductive Dissolution of Goethite by Substituted Phenols. EOS 69(16):369.

LaKind JS, Stone AT. 1986. Reductive Dissolution of Goethite and Hematite by Substituted Phenols. EOS 67(44):948.

Gieskes JM, Elderfield H, Lawrence JR, LaKind J. 1984. Interstitial Water Studies, Leg 78a, *Initial Reports of the Deep Sea Drilling Project*. LXXVIII:377-384.

Gieskes JM, Sirocky FX, LaKind JS. 1984. Interstitial Water Studies, Leg 73, *Initial Reports of the Deep Sea Drilling Project*. LXXIII:539-541.

Gieskes JM, Sirocky FX, LaKind JS. 1983. Interstitial Water Studies, Leg 72, *Initial Reports of the Deep Sea Drilling Project*. LXXII:391-394.

Selected Presentations:

LaKind JS. (with A.M. Rule and F. Wagner). 2024. Creating and Sustaining Successful Public-Private Partnerships (PPPs) for Environmental Monitoring Programs: Principles and Elements. Webinar. 25 July.

Keynote speaker. 2023. Epidemiology and risk assessment: Reflections on working together to improve public health. International Conference on Using Epidemiological Studies in Health Risk Assessments: Relevance, Reliability and Causality. Berlin, Germany. 9 November.

Invited lecturer. 2023. Everything you wanted to know about consulting* - *but were afraid to ask. Lecture, Applied Mathematics and Statistics, The Johns Hopkins University. 15 February. 14 September.

Invited speaker. 2022. "Forever Chemicals" (PFAS) in Breast Milk and Infant Formula: A Global Issue. International Clean-up Conference. Adelaide, Australia. 12 September.

Invited speaker. 2022. PFAS and breast milk: What we don't know, what we should know. 3rd National PFAS Meeting**:** Highly Fluorinated Compounds – Environmental Justice and Scientific Discovery. Wilmington, NC. 16 June.

Invited speaker. 2022. PFAS in breast milk in the US and Canada: Mom/infant exposure data gaps. Health Canada Environmental Health Science and Research Bureau. 25 May.

Invited speaker. 2022. Chemical exposures and health effects: Exposure assessment and interpreting epidemiology research. Center for Food Safety and Applied Nutrition (CFSAN). Division of Risk and Decision Analysis. U.S. Food and Drug Administration. 25 March.

Invited speaker. 2022. Epidemiology and exposure assessment: What toxicologists need to know (or remember). The Toxicology Forum—2022 Virtual Winter Meeting. 25 January.

LaKind JS. 2021. Current breast milk PFAS levels in the US and Canada: After all this time why don't we know more? International Society for Exposure Science Annual Meeting (virtual). 1 September.

LaKind JS. 2020. The Matrix: Bridging the gap between epidemiology and risk assessment. International Society for Exposure Science Annual Meeting. Webinar. 22 September.

LaKind JS, Burns CJ. 2020. The Matrix: Bridging the gap between epidemiology and risk assessment. Series of invited webinars (e.g., US EPA OPPP/OPPT, 9 September; Environmental and Occupational Health Sciences (EOHS) Research Seminar Series at The University of Texas Health Science Center at Houston, School of Public Health, 11 September; Johns Hopkins Bloomberg School of Public Health Current Topics in Epidemiology seminar series, 30 September; Department of Environmental and Occupational Health, Dornsife School of Public Health, Drexel University, 9 November).

LaKind JS. 2020. Environmental Chemicals in Breast Milk and Formula: Exposure and Risk Assessment Implications. The Society for Birth Defects Research & Prevention Virtual 60th Annual Meeting. 30 June.

LaKind JS, Burns CJ. 2020. Epidemiology, exposure and risk assessment. Texas Commission on Environmental Quality. Webinar. 18 June.

LaKind JS. 2019. Exposure Data Quality Assessments: Why and How? Society for Risk Analysis Annual Conference. Arlington, VA. 11 December.

LaKind JS, Burns CJ. 2019. The Matrix: Bridging the gap between epidemiology and risk assessment. Health Canada. Ottawa, Canada. 4 November.

Invited speaker. 2019. Biomonitoring and epidemiology research on personal care products: We're not in Kansas anymore. Personal Care Products Council Annual Safety Seminar. Philadelphia, PA. 30 October.

Invited lecture (with Dr. Heidi S. Erickson and Dr. Carol Burns). 2019. The University of Texas Medical Branch at Galveston/ Chronic Disease Epidemiology Course. 23 April.

Invited lecture. 2019. Conflicts of Interest and Environmental Research. Bioethics, Honors College of Florida Atlantic University. Jupiter, FL. 20 March.

LaKind JS, Burns CJ. 2019. Evidence-based environmental decisions: Bridging the gap between epidemiology and risk assessment. SOT RASS/ISES Webinar. 13 February.

LaKind JS. 2018. Exposure data quality assessments: ExpoQual. International Society of Exposure Science/International Society of Environmental Epidemiology. Ottawa, Canada. 28 August.

Invited speaker. 2018. How to assess and interpret biomonitoring data once you have it.

Workshop on the Feasibility of Addressing Environmental Exposure Questions Using Department of Defense Biorepositories. The National Academies of Sciences, Engineering and Medicine. Washington, DC. 15 June.

Invited speaker. 2018. Chemical exposures and human health: What can we take away from epidemiology research? Occupational Medicine, Clinical Public Health & Epidemiology Army Public Health Center. Aberdeen Proving Ground, MD. 6 June.

Invited speaker. 2018. Evidence-based environmental decision-making: Problems and progress. Bundesinstitut für Risikobewertung. Berlin, Germany. 24 May.

Invited speaker. 2018. Exposure data quality and environmental epidemiology: Implications for systematic reviews and weight of evidence. Environmental Health Science and Research Bureau (EHSRB) Seminar Series. Health Canada. 21 February. Ottawa, Canada.

Invited speaker. 2018. Exposure data quality in environmental epidemiology: Bad habits and remedies. Université de Montréal Public Health Research Institute. 20 February. Montreal, Canada.

Invited speaker. 2017. Exposure data in environmental epidemiology: limitations and quality assessments. European Food Safety Authority Scientific Conference on the Use of Epidemiological findings in Regulatory Pesticide Risk Assessment. 21 November. Parma Italy.

LaKind JS. 2017. Critical and systematic evaluation of 2,4-dichlorophenoxyacetic acid (2,4-D) exposure data: quality and generalizability for human assessments. International Society of Exposure Science Annual Meeting. 18 September. Durham NC.

LaKind JS. 2017. Transparent and systematic reviews of exposure data in environmental epidemiology: Approaches and case studies. International Society of Exposure Science Annual Meeting. 17 September. Durham NC.

LaKind JS. 2017. Evaluating strengths and limitation of the exposure data using the Biomonitoring, Environmental Epidemiology, and Short-Lived Chemicals (BEES-C) Instrument: Implications for science and policy. American College of Epidemiology Annual Conference. 25 September. New Orleans, LA.

Invited speaker. 2017. Chemical exposures and health effects: What we know and what we don't know from epidemiology research. Mid-Atlantic Regional Conference in Occupational and Environmental Medicine. 23 September. Baltimore, MD.

Invited speaker. 2017. Chemical exposures and health effects: What we know and what we don't know from epidemiology research. Occupational and Environmental Residency Program, Johns Hopkins Bloomberg School of Public Health. 18 September. Baltimore, MD.

LaKind JS. 2017. Human exposure to 2,4-D: What do the data tell us? American Chemical Society 254th Annual Meeting. 21 August. Washington DC.

Invited speaker. 2016. Quality matters in environmental epidemiology: The exposure data we collect versus the data we need. Grand Rounds, University of Maryland School of Medicine. 17 November. Baltimore, MD.

Invited speaker. 2016. Can coating complexities. Workshop - Identifying and Evaluating Alternative Materials: The Case of BPA-Free Can Linings. 4 November. UC Berkeley. Berkeley, CA. https://www.youtube.com/watch?v=UqNXi1qNXHQ

Invited speaker. 2016. Biomonitoring and environmental epidemiology: Implications for personal care products. Personal Care Products Council Safety Workshop. 26 October. Alexandria, VA.

LaKind JS. 2016. Assessing Biomonitoring Data Quality**:** The Biomonitoring, Environmental Epidemiology, and Short-Lived Chemicals (BEES-C) Instrument. International Society of Exposure Science Annual Meeting. 12 October. Utrecht, The Netherlands.

LaKind JS. 2016. Harmonization, transparency, and access: Why we need these in environmental epidemiology [exposure science]. International Society of Exposure Science Annual Meeting. 10 October. Utrecht, The Netherlands.

Invited speaker. 2016. Cleaning, environmental exposures and respiratory health effects: Issues, challenges and opportunities. 17 June. Advancing the Science Webinar Series. Sponsored by the American Cleaning Institute (ACI), in collaboration with the Toxicology Excellence for Risk Assessment (TERA) Center, University of Cincinnati and Endorsed by the Society of Toxicology.

Invited speaker. 2016. Environmental Epidemiology: The importance of exposure assessment. CropLife America and RISE Spring Conference. 14 April. Arlington, Virginia.

LaKind JS. 2016. Quality Matters in Environmental Epidemiology: The data we collect versus the data we need. 14 March. Society of Toxicology. New Orleans, LA.

Invited speaker. 2016. Biomonitoring and temporality in environmental epidemiology: The data we collect versus the data we need. U.S. Environmental Protection Agency. Temporal Exposure Issues for Environmental Pollutants: Health Effects and Methodologies for Estimating Risk. 27–29 January. Research Triangle Park, NC

LaKind JS. 2015. Biomonitoring Data in Cumulative Risk Assessment: The Biomonitoring, Environmental Epidemiology, and Short-Lived Chemicals (BEES-C) Instrument. Society for Risk Analysis. 9 December. Arlington, Virginia.

LaKind JS, Naiman DQ. 2015. Temporal trends in BPA exposure in the US from 2003–2012 and factors associated with BPA exposure: Spot samples and urine dilution complicate data interpretation. International Society for Exposure Science. 19 October. Henderson, Nevada.

Invited speaker/panelist. 2015. Exposure Science and Environmental Epidemiology: Problems and Proposed Solutions. ICCA-LRI & US EPA Workshop. What Will Work? Application of New Approaches for Chemical Safety Assessment. June 16-17. New Orleans, Louisiana**.**

Invited poster presentation. 2015. Issues with quality and harmony in environmental epidemiology: PCBs, BPA and phthalates. ICCA-LRI & US EPA Workshop. What Will Work? Application of New Approaches for Chemical Safety Assessment. June 16-17. New Orleans, Louisiana**.**

Invited speaker. 2015. Institute of Medicine Workshop on the Interplay between Environmental Exposures and Obesity. March 2-3. Research Triangle Park, NC.

Invited speaker. 2014. The need for more robust data in environmental epidemiology: BPA as a case study. Toxicology Forum. July 9. Aspen, Colorado.

Invited panelist. 2014. What Is Safe? Integrating Multi-Disciplinary Approaches for Decision Making about the Human Health and Environmental Impacts of Chemicals. ICCA-LRI & JRC Workshop. June 17-18, Lugano, Switzerland.

Speaker. 2014. PCBs and related chemicals in breast milk: What do the data mean for mothers, infants, doctors, regulators and others? Society of Toxicology Annual Meeting. 26 March. Phoenix, Arizona.

Invited speaker. 2013. Endocrine disruptors and obesity, diabetes and heart disease: What does epidemiological research tell us? 15th Cefic-LRI Annual Workshop. 21 November. Brussels, Belgium.

Invited speaker. 2013. Uncertainties in Epidemiology: The Example of Bisphenol A. 2013 Center for Advancing Risk Assessment Science And Policy Workshop. 6 November. Washington DC.

Invited speaker. 2013. Urine and Pool Water: Exposure and Health. World Aquatic Health Conference. 18 October. Indianapolis, Indiana.

Invited speaker. 2013. Cancer Clusters in the USA: What Do the Last 20 Years of State and Federal Investigations Tell Us? DHMH Workgroup on Cancer Clusters and Environmental Causes of Cancer. September 10, Baltimore, Maryland.

Invited speaker/panelist. 2013. What is Normal? Biomarkers of Exposure & Effect. ICCA-LRI & NCATS Workshop: What Is Normal? Implications for Chemical Safety Assessment. June 11-12, Santa Fe, New Mexico.

Guest lecturer. 2013. Human Health Risk Assessment Primer. University of Maryland, College Park. 30 April.

Invited speaker. 2012. 21st Century Solutions for 20th Century Problems: Lessons from 4 decades of environmental epidemiology research. CropLife America & RISE. Spring Conference. Arlington, Virginia. 5 April.

Invited speaker. 2011. Endocrine disruption and risk assessment: The controversial case of bisphenol A. Grand Rounds. Division of Endocrinology, Diabetes and Nutrition, University of Maryland School of Medicine. 31 October.

LaKind JS, Levesque J, Dumas P, Bryan S, Clarke J, Naiman DQ. 2011. Can We Compare United States and Canadian Population Exposures from National Biomonitoring Surveys? Bisphenol A (BPA) as a Case Study. International Society for Exposure Science. Baltimore, Maryland. 27 October.

Invited speaker. 2011. Swimming and asthma: What does the current research say? ACI Asthma Science Forum. Arlington, VA. 10 May.

Invited speaker. 2010. Are the kids alright? Strengthening regulatory decision-making in the uncertain world of children's health research. 12th Cefic LRI Annual Workshop. Brussels, Belgium. 18 November.

Guest Lecturer. 2010. Human Health Risk Assessment Primer. University of Maryland, College Park. 8 November.

Speaker. 2010. The Good, the Bad, and the Volatile: Can We Have Both Healthy Pools and Healthy People? World Aquatic Health Conference. Colorado Springs, CO. 8 October.

Invited speaker. 2010. A Multidisciplinary Approach to Advancing the Science of Neurodevelopmental Testing in Cohorts of Infants and Young Children. Teratology Society's 50th Annual Meeting. Louisville, Kentucky. Joint TS/Neurobehavioral Teratology Society Symposium on Advancing Neurodevelopmental Evaluation in Children. June 29. Citation: LaKind JS, Youngstrom E, Goodman M, Squibb K, Lipkin PH, Anthony LG, Kenworthy L, Mattison D. 2010. A multidisciplinary approach to advancing the science of neurodevelopmental testing in cohorts of infants and young children. NBTS 34 *Neurotoxicology and Teratology 32:*505.

Kenworthy L, Anthony LG, Goodman M, LaKind JS, Lipkin PH, Mattison D, Squibb K, Youngstrom E. 2010. Getting the biggest bang for your buck: Choosing neurodevelopmental tests that maximize power. NBTS35 Neurotoxicology and Teratology 32:506.

Anthony LG, Youngstrom E, Kenworthy L, LaKind JS, Goodman M, Squibb K, Lipkin PH, Mattison D. 2010. Threats to study validity: The Flynn Effect, examiner drift, confounders, lost in translation, and other important considerations. NBTS36 Neurotoxicology and Teratology 32:506.

Invited speaker. 2010. Environmental fate of chemicals: Bring babies into the food web. University of Maryland Baltimore County. 10 March.

Invited participant/speaker. 2009. Human milk biomonitoring: data interpretation and risk assessment issues. International Atomic Energy Agency. Vienna, Austria. 16 February.

Invited speaker. 2008. Grand Rounds. Environmental chemicals and breastfeeding infants. The Johns Hopkins School of Medicine. February 6. Baltimore, Maryland.

LaKind JS, Squibb KS, McElprang DO, Blount BK. Methodologic pilot study of volatile organic compounds (VOCs) in human milk. 2007. 17th Annual Conference of the International Society for Exposure Analysis. October. Durham, North Carolina.

LaKind JS, Aylward LL, Brunk C, DiZio S, Dourson M, Goldstein DA, Kilpatrick ME, Krewski D, Bartels M, Barton HA, Boogaard PJ, Lipscomb J, Krishnan K, Nordberg M, Okino M, Tan Y-M, Viau C, Yager JW, Hays SM. 2007. Guidelines for the Communication of Biomonitoring Equivalents: Report from the Biomonitoring Equivalents Expert Workshop. 17th Annual Conference of the International Society for Exposure Analysis. October. Durham, North Carolina.

Speaker. 2007. Workshop on Childhood Asthma and Environmental Exposures at Indoor Swimming Pools. Advancing the Science. Fourth Annual World Aquatic Health™ Conference. 3 October. Cincinnati, Ohio.

LaKind JS, Berlin CM Jr., Stokes JL, Naiman DQ, Paul IM, Patterson DG Jr., Jones RS, Niehüser S, Park A, Wang RY, Needham LL, Lorber MN, Sjödin A. 2007. Lifestyle and polybrominated diphenyl ethers (PBDEs) in human milk in the United States: A pilot study. 17th Annual Conference of the International Society for Exposure Analysis. October. Durham, NC.

Invited speaker. 2007. Environmental chemicals and breastfeeding infants (update). La Leche League

International's 50th Anniversary Conference. July 23. Chicago.

Invited speaker. 2006. Women's & Children's Health and the Environment. Talking about Environmental Chemicals in Human Milk: Why "Breast is Best." April 24. Baltimore, Maryland.

Invited speaker. 2006. Grand Rounds. What is in mother's milk and what does it mean? Environmental chemicals and breastfeeding infants. Children's Hospital at Sinai. February 14. Baltimore, Maryland.

LaKind JS, Berlin CM Jr. 2005. Workshop on Human Milk Surveillance and Biomonitoring for Environmental Chemicals in the United States. 15th Annual International Society of Exposure Analysis Annual Meeting. November. Tucson, Arizona.

Invited speaker. 2005. Grand Rounds. Interpretation and communication of information from biomonitoring studies. What physicians should know. Maryland General Hospital. October 10. Baltimore, Maryland.

Invited speaker. 2005. Biomonitoring Panel Report: Biomonitoring study design, interpretation, and communication. International Society of Regulatory Toxicology and Pharmacology Workshop: Understanding Human Biomonitoring. June 16. Sacramento, California.

Invited speaker. 2005. What is in mother's milk and what does it mean? Environmental chemicals and breastfeeding infants. Pediatric Academic Societies' Annual Meeting, Perinatal Nutrition and Metabolism Club. May 16. Washington, DC. Invited speaker. 2005. Interpretation and communication of information from biomonitoring studies. Ethics & Sustainability Dialogue Group. May 12. Alexandria, Virginia.

Invited speaker. 2004. Breast Feeding Promotion Task Force. June 7. Baltimore, Maryland.

Invited speaker. 2004. What is in mother's milk and what does it mean? A discourse on environmental chemicals and breastfeeding infants. Institute of Pharmacology and Toxicology, Section of Developmental and Environmental Toxicology, University of Zurich, April 22, Lausanne, Switzerland; World Health Organization, April 26, Geneva, Switzerland.

LaKind JS, Susten A, Mistry K. 2003. Uses and interpretation of human biomonitoring data. Society for Risk Analysis Annual Meeting. December 10. Baltimore, Maryland.

Invited speaker. 2003. Environmental chemicals in human milk. Sixth National Environmental Public Health Conference. December 4. Atlanta, Georgia.

LaKind JS, Bates MN, Wilkins AA. 2003. How useful is measurement of environmental chemicals in human milk in investigations of breast cancer etiology? Dioxin2003. August. Boston, MA.

Invited speaker. 2003. Department of Health and Human Services, Office on Women's Health. Workshop on Breast Cancer and the Environment. June 26. Washington, DC.

Invited speaker. 2003. Chemicals and Risk: What You Should Know, What Patients May Ask. Grand Rounds, Hershey Medical Center, Penn State College of Medicine. April 8. Hershey, Pennsylvania.

LaKind JS, Susten A, Mistry K. 2003. Society for Risk Analysis Annual Meeting. Uses and Interpretation of Human Biomonitoring Data. December 10. Baltimore, Maryland.

Invited speaker. 2003. US Environmental Protection Agency's Children's Health Protection Advisory Committee. Research and surveillance of environmental chemicals in human milk. March 19. Washington, DC.

Invited speaker. 2002. The Johns Hopkins University Bloomberg School of Public Health Education and Research Center Lecture Series. Environmental Chemicals in Human Milk. 2 December. Baltimore, Maryland.

Invited speaker. 2002. US Environmental Protection Agency Children's Health and Protection Advisory Council Science and Regulatory Work Group. 15 October. Washington, DC. Invited speaker. 2002. Breast milk monitoring for environmental chemicals in the U.S. Summary Expert Panel Workshop, Hershey, PA. Workshop on Chemicals and Drugs in Breast Milk. National Institutes of Health. April 24. Bethesda, Maryland.

Pittinger CA, LaKind JS. 2001. Weighing ecological risks and societal benefits: Pharmaceuticals and personal care products in the environment. 22nd Annual Society of Environmental Toxicology and Chemistry Meeting. November 15. Baltimore, Maryland.

Invited speaker. 2001. Protocol for breast milk monitoring for environmental chemicals. Toxic Chemicals in Breast Milk: A National Workshop to Assess Hazards to Children's Health of Chemical Contaminants in Breast Milk. Center for Children's Health and the Environment, Mt Sinai School of Medicine. October 5. New York City, New York.

LaKind JS, Berlin CM. 2001. Developing a protocol for breast milk monitoring for environmental chemicals: Workshop overview. International Society of Exposure Analysis Annual Meeting. November 4-8. Charleston, South Carolina.

LaKind JS, Berlin CM, Naiman DQ. 2001. Infant exposure to chemicals in breast milk in the United States: What we need to learn from a breast milk monitoring program. Presented at the Children's Environmental Health II: A Global Forum for Action. September 8. Washington, DC.

LaKind JS, Berlin CM. 2000. PDBEs in breast milk: Where do we go from here? Dioxin2000. August 13-17. Monterey, California.

LaKind JS, Berlin CM, Naiman DQ, Park CN. Characterization of dose distributions of selected breast milk contaminants to nursing infants: DDE and TCDD. American Public Health Association Annual Meeting, November, 1999; Society for Risk Analysis Annual Meeting, December, 1999; and Dioxin2000, Monterey, California, August 13-17, 2000.

Invited speaker. 1998. Principles of toxicology. School Nurse Institute. August 5. Towson, Maryland.

Invited speaker. 1998. Alchemy, risk assessment, and other phenomena. Lawrence University Science Colloquium. April 17. Appleton, Wisconsin.

Invited speaker. 1997. Managing risk in the face of scientific uncertainty. The Center for Technology, Environment, and Development (CENTED). Clark University. September 26. Worcester, Massachusetts.

Williams LG, Fendick E, LaKind JS, Stern B, Strand JA, Tardiff RG. 1995. Risk-based water quality criteria for treated mine-tailings effluent. Second World Congress of the Society of Environmental Toxicology and Chemistry.

Invited speaker. 1994. Comparison of human health risk assessment modeled data with observed data: Dioxin and lead. University of Guelph Department of Statistics. Guelph, Canada.

Invited speaker. 1993. Morgan State University Chemistry Department. Lecture on aquatic chemistry concepts and environmental and regulatory applications.

Invited speaker. 1992. Contradictions between Predictions and the Real World. National Association of Health Professionals Annual Conference. Norfolk, VA.

LaKind JS, Naiman DQ. 1991. Comparison of predicted and observed dioxin levels in fish: Implications for risk assessment. Society for Risk Analysis Annual Meeting.

LaKind JS, Rifkin E. 1991. A coordinated approach to dioxin regulation. Presented at Dioxin: National Conference on Establishing Multimedia Controls. May, 1991. Washington, DC.

Invited speaker. 1991. Use of the BCF in criteria development for hydrophobic compounds. Virginia Water Pollution Control Association Annual Conference.

LaKind JS, Rifkin E. 1990. Current method for setting dioxin limits in water requires reexamination. Dioxin and PCBs: National Conference on Approaches to Address Human Health Risks and Aquatic Life Impacts. May 10-11, 1990. Washington, DC.

LaKind JS, Rifkin E. 1990. Alternative approach for developing criteria for hydrophobic substances. 11th Annual Meeting of the Society of Environmental Toxicology and Chemistry.

LaKind JS, Stone AT. 1988. Reductive dissolution of goethite by substituted phenols. Annual Meeting of the American Geophysical Union.

LaKind JS, Stone AT. 1986. Reductive dissolution of goethite and hematite by substituted phenols. Annual Meeting of the American Geophysical Union.

LaKind JS, Brown PE. 1984. Characterization of the gold-bearing fluid at Red Lake, Ontario. Annual Meeting of the Geological Association of Canada- Mineralogical Association of Canada.

Professional Activities/Recognition:

Special Issue Guest Editor (with J. Domino). 2024. *Journal of Environmental Exposure Assessment*. Guest editorial: Domingo JL. LaKind JS. Environmental chemicals in breast milk and infant formula: measurements, interpretation, and communication. *J. Environ. Expo. Assess.* 2024, *3*, 25. http://dx.doi.org/10.20517/jeea.2024.49.

Insights Editor (founder). 2024 - present. *Environment International.*

Special Issues Editor. 2023-2024. *Environment International.*

Member. 2022 – 2024. Justice, Equity and Risk Specialty Group, Society for Risk Analysis.

Society of Toxicology. Junior Councilor, SOT Exposure Specialty Section. 2022-2023.

Mentor. 2021 – present. The Johns Hopkins University Mentoring Program.

Invited panelist. National Academies Committee on Guidance on PFAS Testing and Health Outcomes Information Gathering Session. 2021.

Member, Peer Consultation on Biomonitoring Data and Reverse Dosimetry to Estimate Chemical Exposures. 2021. FDA/CFSAN/Versar.

Member, Technical Organizing Committee. 2021. International Society of Exposure Science Annual Meeting.

ISES. 2020 - 2022. Ethics Committee.

EPA Grant Review Panel. 2020.

Steering Committee, 2020-present. i-HBM (International Human Biomonitoring) Working Group, ISES.

Session chair. 2020. Epidemiology, Exposure Science, and Risk Assessment: We need each other. International Society of Exposure Science. 22 September.

Member, HESI Assembly. 2019-2020.

Member, 2019 - 2020. Core Science Panel of the Beyond Science and Decisions Workshop Series.

Special issue editor. 2019. International Journal of Environmental Research and Public Health. Special Issue: Environmental Health Study with Remote Sensing Technologies: Exposure Assessment and Health Outcomes.

Appointed member. Health Effects-Energy Research Committee. December 18, 2017-2023.

ISES Committee member, Diversity, General Scientific Meetings Committees. January -December 2019.

ISES Vice Chair, Finance Committee, January-December 2019.

ISES Past President. January-December 2019.

ISES President. 2017-2018.

Session co-chair. 2018. Society Presidents' Call for Discussion: Intersection of Epi, Exposure and Decision-Making: Data Quality for Public Health Protection. International Society of Exposure Science/International Society of Environmental Epidemiology. Ottawa, Canada. 29 August.

Session co-chair. 2018. Exploring Current Worker Exposure Tools and Their Capability to Support Risk Evaluations of Chemicals under Amended TSCA. International Society of Exposure Science/International Society of Environmental Epidemiology. Ottawa, Canada. 28 August.

Session co-chair. 2018. Strengthening Exposure Assessment in Environmental Epidemiology: Problem Identification and Suggestions for Path Forward International Society of Exposure Science/International Society of Environmental Epidemiology. Ottawa, Canada. 28 August.

Invited member. 2018. Organizing Committee of the Conference on Uncertainty in Risk Analysis, 2019, Berlin, Germany.

Invited member. 2018. Technical Advisory Board, Total Exposure Health Conference and Workshop "Total Exposure Health: Bridging Exposure Science and Precision Medicine".

ISES Committees. ex officio member, all Committees, 2017-2018.

Founder, ISES Newsletter, 2017. Editorial Board, ISES Newsletter, 2017-2019.

Invited member. 2017. HESI Epidemiology "Best Practices" Project.

Session co-chair. 2017. International Society of Exposure Science Annual Meeting. 18 September. Durham NC. Exposure Assessment and Epidemiology for Regulatory Decision Making- Challenges and Opportunities (with June Yan). Durham, NC. 18 October.

Session co-chair. 2017. International Society of Exposure Science Annual Meeting. 2,4-D – A Case Study of Decades of Exposure Science; A Discussion of Quality, Quantity, and Harmonization (with Carol Burns). Durham, NC. 19 October.

Session Organizer. 2017. 2,4-D Human Exposure Data: Lessons from Decades of Study. American Chemical Society 254th Annual Meeting. Washington DC. 21 August.

Invited reviewer. 2017. Research-Practice Grants. Gulf Research Program of the National Academies of Sciences, Engineering, and Medicine. Washington DC. 12 September.

Invited reviewer. 2017. Minnesota Department of Health (MDH) revised health-based values for water. PFOS and PFOA.

Invited member. 2017. GRADE Guidance for Modelled Data Working Group. Hamilton, Ontario. 15-16 May.

Invited member. 2017. Risk Of Bias In Non-randomized Studies of Exposures (ROBINS-E) Working Group. Bristol, UK. 30-31 January.

HESI RISK21 Science Advisory Board. 2017-2020.

2017 SOT Regulatory and Safety Evaluation Specialty Section Award: Best Paper Contributing to the Field of Regulatory and Safety Evaluation in Toxicology. Beck et al. Approaches for describing and communicating overall uncertainty in toxicity characterizations: U.S. Environmental Protection Agency's Integrated Risk Information System (IRIS) as a case study. *Environment International* 89–90:110–128.

Member, Technical Organizing Committee. 2017. International Society of Exposure Science Annual Meeting.

Reviewer. 2017. Using 21st Century Science to Improve Risk-Related Evaluations. The National Academies Press.

Symposium Chair (with M. Mortensen). 2016. Biomonitoring: The Genie is out of the Bottle: Challenges in Data Quality and Interpretation. International Society of Exposure Science. Utrecht, The Netherlands. 12 October.

Symposium Chair (with D. Mattison). 2016. Harmonization, access, transparency: improving environmental epidemiology for public health decision-making. International Society of Exposure Science. Utrecht, The Netherlands. 10 October.

Invited member. 2016. National Institutes of Health Working Group - Risk Of Bias In Non-randomized Studies of Exposures. 2016.

Invited member. Epidemiology and Risk Assessment Expert Panel. 8 April 2016.

Invited member. EPA Expert Workshop on Aggregate Exposure Pathway: A Conceptual Framework to Advance Exposure Science Research and Complete the Source-to-Outcome Continuum for Risk Assessment. May 9-11, 2016. Research Triangle Park, North Carolina.

Invited member, Maryland Department of Health and Mental Hygiene (DHMH) Cancer Cluster Advisory Committee. 2016.

Membership Committee, Society for Risk Analysis. 2016.

President-Elect, International Society of Exposure Science. 2016.

Member, Technical Organizing Committee. 2016 International Society of Exposure Science Annual Meeting.

EPA Scientific and Technological Achievement Award (STAA) Level III for 2015 for: Providing Critical Models and Information Needed for Exposure and Risk Assessments of Environmental Chemicals in Infants.

Invited member, Review Panel, National Cancer Institute Laboratory of Metabolism (LM) of the NCI Intramural Program. September 16-18, 2015. Bethesda MD.

Jury member, ISES representative. 2015 LRI Innovative Science Award.

Invited participant. 2015. Institute of Medicine's Roundtable on Environmental Health Sciences, Research, and Medicine Workshop: The Interplay between Environmental Exposures and Obesity. March 2-3, Research Triangle Park, NC.

Co-Chair (with Dr. Benjamin Blount, CDC), 2015 Annual Meeting, International Society of Exposure Science. Henderson, NV. 18-22 October.

Founder, ISES Women's Networking Event. 2014.

Member, Diversity Committee. 2015 - present. International Society of Exposure Science.

Member, Nominations Committee. 2014 - present. International Society of Exposure Science.

Member, General Scientific Meetings Committee. 2014 - present. International Society of Exposure Science.

External Peer Reviewer. 2013. America's Children and the Environment. Third Edition. Environmental Protection Agency. EPA 240-R-13-001.

Grant Proposal Review. Health Canada's Chemicals Management Monitoring and Surveillance Fund. 2013.

Appointed member. Maryland Pesticide Reporting and Information Workgroup. June 2013.

Grant Proposal Review. Research Foundation - Flanders (Fonds Wetenschappelijk Onderzoek - Vlaanderen, FWO). April 2013.

Facilitator, Best Practices for Obtaining, Interpreting and Using Human Biomonitoring Data in Epidemiology and Risk Assessment: Chemicals with Short Biological Half-Lives. April 10-12, 2013. Baltimore, MD.

Facilitator, Advancing Cancer Cluster Assessments: Starting the Dialogue. April 3-5, 2013. Baltimore, MD.

Editorial Board. 2013. *Environment International*. February 2013- present.

Scientific Program Committee, 2013. Environmental Health Conference, Basel, Switzerland. 19-23 August. Joint conference of the International Society of Environmental Epidemiology (ISEE), International Society of Exposure Sciences (ISES) and International Society of Indoor Air Quality (ISIAQ).

Councilor, International Society of Exposure Science. 1 January 2013 – 31 December 2015.

Board of Directors, National Swimming Pool Foundation. 1 November 2012 – 28 October 2015.

Invited participant. 2012. Expert Workshop on Approaches to Improving the Risk Assessment of Persistent, Bioaccumulative and Toxic (PBT) Chemicals in Breast Milk. Environmental Protection Agency, Research Triangle Park, North Carolina. October 24-26.

Discussion Leader. 2012. Swimming Pools: Chemistry and Respiratory Effects, Gordon Research Conference, Drinking Water Disinfection Byproducts. Mount Holyoke College, August 5-10.

Panel member. 2012. US Environmental Protection Agency Science Advisory Board Panel on Perchlorate - Approaches for Deriving Maximum Contaminant Level Goals for Drinking Water.

Invited participant. Experts panel on exposure to swimming pool disinfection by-products and asthma and allergy effects. Porto, Portugal. 15 March 2011.

Mentor. 2011 - present. International Society of Exposure Science Mentor Program.

Facilitator, Children's Environmental Health & Protection Advisory Council: Feasibility of Biomonitoring in Maryland: An Open Meeting & Discussion. 1 April 2011. Laurel, MD.

Grant Proposal Review. Health Canada's Chemicals Management Monitoring and Surveillance Fund. 2011.

Grant Proposal Review. Health Canada's Chemicals Management Plan Monitoring & Surveillance Fund. 2011.

Grant Proposal Review. Human and Social Sciences, Epidemiology and Public Health, National Cancer Institute, France. 2011.

Institute of Medicine Committee on Blue Water Navy Vietnam Veterans and Agent Orange Exposure. May 2010 - 2011.

Graduate Council, UMBC. Associate member. April 2010 – present.

Grant Proposal Review: NIEHS. Superfund Basic Research and Training Program. October 2009.

Environmental Health Advisor, Maryland Department of the Environment Science Services Administration. June 2008-June 2009.

Grant Proposal Peer Review: NIEHS R21. Research to Action: Assessing and Addressing Community Exposures to Environmental Contaminants. July 2009.

Grant Proposal Peer Review: AAAS Research Competitiveness Service; Washington State's Life Sciences Discovery Fund. 2009.

Society of Toxicology Risk Assessment Specialty Section 2008 Top Ten Publications Advancing the Science of Risk Assessment awarded to Hays, S.M., Aylward, L.L., LaKind, J.S., et al. 2008. Guidelines for the Derivation of Biomonitoring Equivalents: Report from the Biomonitoring Equivalents Expert Workshop. *Regulatory Toxicology and Pharmacology* 51(3, Suppl 1):S4-S15.

Society of Toxicology Risk Assessment Specialty Section 2008 Top Ten Publications Demonstrating an Application of Risk Assessment awarded to Aylward LL, LaKind JS, et al., *J Toxicol Environ Health A* 71(22):1499-1508.

Board of Directors, U.S. – Montenegro Business Council. January -September, 2009.

Project Committee. 2008. *Maryland's Children and the Environment*. August. http://www.dhmh.state.md.us/reports/pdf/MDChildrenEnv08.pdf

Associate Editor. *Journal of Exposure Science and Environmental Epidemiology* 2008-2014.

Aquatics International Power 25. 2008. http://www.aquaticsintl.com/2008/feb/0802_power.html

Workshop Facilitator. 2007. Workshop on Childhood Asthma and Environmental Exposures at Indoor Swimming Pools. Advancing the Science. 21-24 August. Leuven, Belgium.

Associate Editor. 2006. Environmental and Neurodevelopmental Disorders. Special Issue of NeuroToxicology, vol 27, Issue 5.

Invited participant. 2006. WHO Consultation to Develop a Strategy to Estimate the Global Burden of Foodborne Diseases. 25-27 September. Geneva, Switzerland.

Workshop Co-Instructor (D. Barr, A. Calafat, L. Needham). 2005. Exposure Assessment for Environmental Chemicals Using Biomonitoring. International Society for Exposure Analysis. Tucson, Arizona. November, 2005.

Symposium Chair (with B. Blount). 2005. Environmental Chemicals in Human Milk. International Society for Exposure Analysis. Tucson, Arizona. November, 2005.

Organizing Committee. 2005. Twenty-Second International Neurotoxicology Conference. Environment and Neurodevelopmental Disorders. Research Triangle Park, NC. 11-14 September.

Workshop Steering Committee and Organizer. 2005. Hershey Medical Center Technical Workshop: Optimizing the Design and Interpretation of Epidemiologic Studies for Assessing Neurodevelopmental Effects from In Utero Chemical Exposure. Research Triangle Park, NC. 14 September, 2005.

Session Co-chair (with L.L. Needham). Body Burden and Dietary Intake, Dioxin2005. Toronto, Canada. August, 2005.

Invited Participant: International Biomonitoring Workshop, ILSI Health and Environmental Sciences Institute, Research Triangle Park, NC, September, 2004.

Member, World Health Organization Survey Coordinating Committee for the WHO Global Survey of Human Milk for Persistent Organic Pollutants (POPs). Since 2004.

Workshop Organizer (with C.M. Berlin): Second Technical Workshop on Human Milk Surveillance and Biomonitoring Research on Environmental Chemicals in the United States. Milton S. Hershey Medical Center, Pennsylvania State University College of Medicine, 24-26 September 2004.

Symposium Chair. 2003. Society for Risk Analysis Annual Meeting. Uses and Interpretation of Human Biomonitoring Data. Baltimore, MD. December 7-10.

Technical Program Committee. 2003. Dioxin2003, Boston, MA. Session Chair: Public Health Decision-Making and Resource Allocation: Dioxin and Other PBTs as a Case Study.

Guest Editor. 2002, 2005. *Journal of Toxicology and Environmental Health*, issues on the Technical Workshop on Human Milk Surveillance and Research on Environmental Chemicals in the United States.

Workshop Organizer (with C.M. Berlin): Technical Workshop on Human Milk Surveillance and Research on Environmental Chemicals in the United States. Milton S. Hershey Medical Center, Pennsylvania State University College of Medicine, 15-17 February 2002.

Appointed Member: Maryland's Children's Environmental Health and Protection Advisory Council, December 2000 – July 2008.

Appointed Member: Maryland Lead Poisoning Prevention Commission, January 2000 – February 2002.

Invited Award Selection Panel Member: USEPA Science Achievement Award in Water Quality. 1998.

Guest Editor: *International Journal of Environment and Pollution*. Special Issue on Environmental Risk Assessment: Issues and Methods. Vol. 9, No. 1. 1998.

Session Organizer and Chair: Emerging EPA Guidance: Implications for the Pulp and Paper Industry. Annual TAPPI Environmental Division Conference, May 5-7, 1997.

TAPPI, Technical Program Committee Member. 1996 - 1997.

Technical Editor: Guidelines for the Evaluation and Control of Lead-Based Paint Hazards in Housing (1995 Edition). US Department of Housing and Urban Development.

Symposium Chair: Society for Risk Analysis Annual Meeting. Organized session on *Predicting Blood Lead Levels: Models and Applications*. December, 1994.

Invited Participant: Alliance for the Chesapeake Bay Roundtable on the Toxics Reduction Strategy of the Chesapeake Bay Program. Baltimore, May, 1994.

Invited Participant: Washington State Departments of Health and Ecology Sediment Scientific Review Board. Seattle, 1993.

Participant: Scientific Working Conference on Bioaccumulation of Hydrophobic Organic Chemicals. Institute for Evaluating Health Risks, Washington DC, June 1992.

Editorial Board: *Journal of Toxicology and Environmental Health*. 1992-2024.

Editorial Board: *Environmental Toxicology and Chemistry*. 1996-1998.

Peer Reviewer: *Environmental Health Perspectives, Journal of Exposure Science and Environmental Epidemiology, Chemosphere, Risk Analysis: An International Journal, Public Health Reports, Environmental Research, Journal of Pediatric Gastroenterology and Nutrition: An International Journal of Clinical, Experimental and Developmental Investigation, Toxicology and Applied Pharmacology, Integrated Environmental Assessment and Management, Reproductive Toxicology, Food and Chemical Toxicology, Environment International, Environmental Pollution, Reviews on Environmental Health, Toxicology and Industrial Health, Critical Reviews in Toxicology, International Journal of Hygiene and Environmental Health*

Member of Board of Directors, Advisory Board, and past President: Baltimore Coalition Against Childhood Lead Poisoning, Inc., Coalition for a Lead Safe Environment. 1992-1994.

Guest Editor: *Journal of Toxicology and Environmental Health*, 1991.

American Chemical Society Graduate Student Award in Environmental Chemistry. 1987.

On-line media:

ROBINS-E Development Group (Higgins J, Morgan R, Rooney A, Taylor K, Thayer K, Silva R, Lemeris C, Akl A, Arroyave W, Bateson T, Berkman N, Demers P, Forastiere F, Glenn B, Hróbjartsson A, Kirrane E, LaKind J, Luben T, Lunn R, McAleenan A, McGuinness L, Meerpohl J, Mehta S, Nachman R, Obbagy J,

O'Connor A, Radke E, Savović J, Schubauer-Berigan M, Schwingl P, Schunemann H, Shea B, Steenland K, Stewart T, Straif K, Tilling K, Verbeek V, Vermeulen R, Viswanathan M, Zahm S, Sterne J). Risk Of Bias In Non-randomized Studies - of Exposure (ROBINS-E). Launch version, 1 June 2022. Available from: https://www.riskofbias.info/welcome/robins-e-tool.

LaKind JS. 2018. Webinar: Chemical exposures and health effects: What we know and what we don't know from epidemiology research. CME through Accreditation Council for Continuing Medical Education (ACCME). Johns Hopkins Bloomberg School of Public Health, Johns Hopkins Education and Research Center for Occupational Safety and Health. https://www.jhsph.edu/research/centers-and-institutes/johns-hopkins-education-and-research-center-for-occupational-safety-and-health/ce/ChemicalEpiCME

LaKind JS. 2106. Webinar: Environmental Contributions to Asthma Prevalence: Assessing the Link between Exposure and Disease. Advancing the Science Webinar Series: Chemical-Induced Asthma. University of Cincinnati College of Medicine. 17 June.

LaKind JS. 2013. Soapbox Science, Nature.com Guest blog. Environmental chemicals in our bodies – we know they are in there, but what does it mean? http://blogs.nature.com/soapboxscience/2013/01/02/environmental-chemicals-in-our-bodies-we-know-they-are-in-there-but-what-does-it-mean 2 January.

Exposure science video for the International Society of Exposure Science. "Get connected - join the International Society of Exposure Science!!" https://www.youtube.com/watch?v=Qcx65X5Davo

Research/Grants:

Investigator: Pilot Study on Concentrations of PBDEs in Human Milk (with Drs. C. M. Berlin, Jr. and I. Paul, Milton S. Hershey Medical Center, Penn State College of Medicine, and Dr. D. Patterson, Centers for Disease Control and Prevention). Cooperative Agreement CR-83150601-0 from the US Environmental Protection Agency. 2003.

Investigator: Partitioning and Elimination Kinetics Study of Human Milk and Blood (with Drs. C. M. Berlin, Jr. and I. Paul, Milton S. Hershey Medical Center, Penn State College of Medicine, and Drs. A. Sjödin and D. Patterson, Centers for Disease Control and Prevention). 2004.

Investigator: Human Milk Biomonitoring For Environmental Chemical (Volatile Organic Compound) Exposures (with Dr. K Squibb, University of Maryland School of Medicine and Dr. B. Blount, Centers for Disease Control and Prevention). 2005.

Principle Investigator. Review of Neurodevelopmental Function Tests in Children (with Drs. Eric Youngstrom, Michael Goodman, Katherine Squibb, Paul H. Lipkin, Laura Gutermuth Anthony, Lauren Kenworthy, Donald R. Mattison). Cefic/LRI Research Grant. 2009.

Principle Investigator. Development of Guidelines for Addressing Contamination and Associated Toxicity in Freshwater/Marine/Estuarine Sediments. Maryland Department of the Environment. 2009-2010.

Principle Investigator. Critical review of epidemiological evidence for the potential association between endocrine active chemicals and obesity, diabetes and cardiovascular disease (with Drs. Donald Mattison, Michael Goodman). Cefic/LRI Research Grant. 2013.

Principle Investigator. Exploring the Design Elements for Successful Public-Private Partnerships (PPPs) for Community Environmental Monitoring Programs (with Drs. Ana Rule and Fernando Wagner). Foundation for Chemistry Research and Initiatives Research Grant. 2022.

MPI (with Dana Boyd Barr [Emory] and Daniel Q. Naiman [Johns Hopkins]). Does NHANES underestimate true population-based exposures to pesticides? Exploring bias in NHANES human biomonitoring data." NIEHS RO3. 2023.

Selected Co-Authored Reports/Articles:

LaKind JS, Naiman J. 2022. White Paper: Review of the PFAS Personal Intervention Literature, Appendix E. In: National Academies of Sciences, Engineering, and Medicine 2022. Guidance on PFAS Exposure, Testing, and Clinical Follow-Up. Washington, DC: The National Academies Press. https://doi.org/10.17226/26156.

HEI Energy Research Committee. Rosofsky A, Dunn-Norman S, Ebelt S, Hornberger G, Hu H, LaKind JS, Russell AG, Thorne PS, Adelsheim LA, Vorhees DJ. 2022. Recommendations for epidemiologic research to inform environmental health policy for unconventional oil and gas development.

HEI Energy Research Committee. 2020. Human Exposure to Unconventional Oil and Gas Development: A Literature Survey for Research Planning (FINAL COMMUNICATION). Communication 1. June 2020.

HEI Energy Research Committee. 2019. Potential Human Health Effects Associated with Unconventional Oil and Gas Development: A Systematic Review of the Epidemiology Literature (FINAL REPORT). Special Report 1. September 2019

Environmental Protection Perchlorate Advisory Panel. 2013. SAB Advice on Approaches to Derive a Maximum Contaminant Level Goal for Perchlorate Final Report.

IOM Committee. 2011. Blue Water Navy Vietnam Veterans and Agent Orange Exposure. The National Academies Press. Washington DC.

LaKind JS, Blatchley ER. 2011. The ABCs of DBPs. Aquatics International. February.

http://www.aquaticsintl.com/2011/feb/1102_tech.html

University of Maryland. 2009. Standard Operating Procedures for Fish and Shellfish Collection and Analysis. For: Maryland Department of the Environment Science Services Administration. 22 May.

University of Maryland. 2009. Technical Support Document for Establishing Fish and Shellfish Consumption Advisories in Maryland. For: Maryland Department of the Environment Science Services Administration. 23 March.

LaKind Associates, LLC (with Dr. E.J. Bouwer). 2003. Investigation of the Removal of Formaldehyde and Phenol by Funeral Home Septic Systems. Prepared for the National Funeral Directors Association. May 2003.

LaKind Associates, LLC and ENVIRON International Corporation. 2002. Assessment of Triclosan Residues In Breast Milk Based on Available Data: Final Report.

LaKind Associates, LLC. Human Health Risk Evaluation of the Windsor Terminal Site, Baltimore, Maryland. December, 2000.

LaKind Associates, LLC. Onsite Human Health Risk Evaluation of TCE at the Sparks, Maryland Leica, Inc. Site. October, 1999.

The Sapphire Group, Inc. Distributions of Exposures Among Workers to Selected ETS-Related Chemicals in Indoor Workplace Air Using Data from the Oak Ridge 16-City Study. March, 1998.

The Sapphire Group, Inc. Critical Review of the USEPA's Proposed Rule for National Ambient Air Quality Standards for Particulate Matter. February, 1997.

EA Engineering, Science, and Technology, Inc. Ethylene Glycol: Scientific Rationale for Continued Listing on EPA's Toxics Release Inventory (TRI). Prepared for ARCO Chemical Company, February, 1996.

EA Engineering, Science, and Technology, Inc. Comparative Toxicity and Environmental Impacts of Ethylene Glycol and Propylene Glycol: A Review. Prepared for ARCO Chemical Company, February, 1996.

EA Engineering, Science, and Technology, Inc. Decision Support Document on Health Benefits and Health and Safety Associated with the Use of Methyl Tertiary Butyl Ether (MTBE) in Gasoline. Prepared for ARCO Chemical Company, December, 1995.

EA Engineering, Science, and Technology, Inc. Report on Toxins Analysis and Assessment (Phase I). Prepared for International Paper Company, November, 1995.

EA Engineering, Science, and Technology, Inc. Phase II Site Investigation Camp Buckner Skeet and Trap Range, U.S. Military Academy, West Point, New York. Prepared for U.S. Army Corps of Engineers - Baltimore District, November, 1995.

EA Engineering, Science, and Technology, Inc. Technical Papers on MTBE and Human Health. Health Benefits Analyses. Prepared for ARCO Chemical Company, October, 1995.

EA Engineering, Science, and Technology, Inc. Human Health Risk Assessment of Manufactured Gas Plant Residuals and Other Chemicals at Baltimore Gas & Electric Company's (BGE) Spring Gardens Facility — Evaluation of the Need for Additional Offsite Information to Conduct an Offsite Risk Assessment. Prepared for Baltimore Gas & Electric Company, March, 1995.

EA Engineering, Science, and Technology, Inc. Preliminary Analysis of Health Risk for the Proposed Kensington Mine Submarine Discharge. Prepared for confidential client. 1994.

EA Engineering, Science, and Technology, Inc. Human Health Risk Assessment of Manufactured Gas Plant Residuals and Other Chemicals to Construction Workers at Baltimore Gas & Electric Company's (BGE) Spring Gardens Facility. Prepared for Baltimore Gas & Electric Company, November, 1994.

EA Engineering, Science, and Technology, Inc. Environmental Impact Analysis: Blue Mountain Sportsman's Center. Prepared for Westchester County, September, 1994.

EA Engineering, Science, and Technology, Inc. Modeled Predictions of Disinfection By-Products for the Baltimore Water Supply System After Implementation of Zebra Mussel Control. Prepared for KCI Engineers, February, 1994.

Student Mentoring:

2024-present: International Society for Exposure Science Mentor Program. Alexandra Del Favero-Campbell, Ph.D. candidate, Dalhousie University.

2024-present: mentoring Melissa Vendramini, student at Lakewood Ranch High School, Bradenton, Florida.

2021. Facilitator. International Society for Exposure Science Webinar: Top tips for Writing an Academic and Industrial Curriculum Vitae. 8 November.

2021-present. Johns Hopkins Engineering Mentoring Program.

2018-present. Dissertation committee member. Cecilia Alcala, Tulane University Ph.D. candidate. Awarded Ph.D. in 2020.

2014-2017. Doctoral defense committee member. Huan Xia, UMBC Ph.D. candidate. Awarded Ph.D. in 2017.

2012-2013 International Society for Exposure Science Mentor Program. Satori Marchitti, Ph.D., US Environmental Protection Agency, National Exposure Research Laboratory.

2012. Eric Sewell, summer intern, Johns Hopkins University Department of Applied Mathematics and Statistics.

2011-2012. International Society for Exposure Science Mentor Program. Liesel M. Seryak, Ph.D. candidate, The Ohio State University College of Public Health.

2011. Doctoral defense committee member. Piuly Paul, UMBC Ph.D. candidate.

2009. Mentor, Maryland Department of the Environment, Chunxiao Zhu, MS candidate, Department of Geography & Environmental Engineering, Johns Hopkins University.

2009. Mentor, Maryland Department of the Environment, Edward Berg, MS candidate, Department of Geography & Environmental Engineering, Johns Hopkins University.

Employment History:

Employer: LaKind Associates, LLC
Employed: June 1998 - present
Title: Founder, President

Employer: University of Maryland Baltimore County
Employed: January 2010 – May 2010
Title: Part Time Instructor, College of Engineering & Information Technology

Employer: University of Maryland School of Medicine
Employed: September 2008 – 2009
Title: Associate Professor

Employer: University of Maryland School of Medicine
Employed: July 2008 – June 2009

Title: Environmental Health Advisor, Maryland Department of the Environment

Employer: University of Maryland School of Medicine
Employed: May 2003 – present
Title: Adjunct Associate Professor

Employer: University of Maryland School of Law
Employed: May 2003 – May 2004
Title: Adjunct Associate Professor

Employer: The Sapphire Group
Employed: January 1997 - May 1998
Title: Co-founder, Vice President, and Managing Principal

Employer: EA Engineering, Science and Technology, Inc.
Employed: September 1993 - December 1996
Title: Senior Scientist

Employer: The Johns Hopkins University
Employed: September 1991 - 1994
Title: Instructor, Aquatic Chemistry

Employer: The Johns Hopkins University
Employed: September 1993 - December 1994
Title: Instructor, Environmental Risk Assessment

Employer: Self-employed, JSL Consulting
Employed: June 1991 - August 1993
Title: Environmental Consultant

Employer: Rifkin & Associates, Inc.
Employed: October 1988 - May 1991
Title: Senior Associate

Employer: U.S. Environmental Protection Agency, Office of Federal Activities
Employed: 1988
Title: Geologist

APPENDIX 2: SHOWER model factors and options for modifications

The following table shows the ATSDR SHOWER model parameters and the exposure assessment factors that can be modified. Unless otherwise specified in the main body of this Report, default values were used. The chemical information characterizes the water concentration of the chemical; the other factors describe the people and residence being modeled (ATSDR 2022c). There are also model parameters related to properties of chemicals; the values for these were not modified and these are not shown here.

Residential	
Factor type	**Options**
Chemical information[1]	
Concentration in water	User-specified
Units	User-specified (e.g., ppm, ppb)
Concentration in air[2]	User-specified
Units	User-specified (e.g., ppm, μg/m^3)
Report units	ppb or μg/m^3
Household scenarios	
Number of people	From 1 to 8
Number and time of showers/baths	Morning or evening
Exhaust fan when bathroom occupied	Open or closed
Bathroom door when bathroom occupied	Open or closed
Exposure group[3]	9 standard ATSDR groups
House information	
Number of bathrooms with showers	1 or 2
Shower/bath layout	Bathtub with shower or separate[4]
Clothes washer location	Main house or bathroom[5]
Exhaust fan location	Bathroom or shower[6]
Area volumes (house, bathroom, shower)	Default or user-specified[7]
Appliance parameters	
Main house compartment parameters (Kitchen sink flow rate and duration per use, utility sink volume per person, dishwasher volume per cycle, cycle duration and start time)	Default or user-specified[8]

Clothes washer parameters (location, volume per cycle, cycle duration, start time)	Default or user-specified[9]
Bathroom compartment parameters (sink flow rate, duration per use, toilet volume per flush)	Default or user-specified
Exhaust fan parameters (location and flow rate)	Default or user-specified
Bathtub volume	Default or user-specified
Shower flow rate	Default or user-specified
Activity Patterns	
Shower pattern for each modeled person (morning or evening; duration, time in bathroom after showering; optional bathtub setting)	Default or user-specified
Time between bathroom stay and next shower	Default or user-specified
Number of bathroom visits separate from shower	0 to 5
Kitchen sink uses	Maximum of 30
Activity start and end times (time when all morning/evening showers begin and are completed)	Default or user-specified
Time away from home for each resident	Default or user-specified.
Communal Facilities	
Building type	Commercial gym (non-school), commercial daycare, dorm or barracks, office, school, other
Facility type	Shower and locker room, bathroom (no shower)
Number of people using the facility	Can input up to 1,000 people
Chemical name and water concentration	User inputs
Concentration in outdoor air	Use site-specific data if available, otherwise use zero (default)
Concentration in air in main building	Input if available, otherwise use zero (default)

Time each person spends in building	Standard distribution or custom
Average bathroom visits per person	Modifiable, default = 5
Percent of people taking showers	Default depends on type of building
Operating hours	Open all day or closed for part of day (times can be input)
Usage pattern	Constant or peak (times can be input for peak)
Number of showers, toilets, bathroom sinks	Number corresponds to number of people using facility. Other values can be used.
Total shower and total bathroom volumes	Default values are based on numbers of appliances. Ceiling height assumed to be 8'. Other values can be used.
Shower and bathroom air exchange rates	Default depends on type of building. Other values can be used.
Toilet volume per flush and sink and shower flow rates	Defaults included. Other values can be used.
Outdoor air supply flow rate, bathroom exhaust fan flow rate, shower exhaust fan flow rate	Default depends on numbers of showers and toilets and floor area. Other values can be input.
Activity durations: shower duration, times in locker room before and after shower, time in bathroom for bathroom-only visits, sink use duration	Defaults given. Other values can be input.

[1] The sources for the chemical properties (e.g., molecular weight, f values, permeability coefficients) are described in ATSDR (2022c).

[2]If outdoor air concentration is not known, the default value is zero.

[3]A customized group can be added if total body surface area, hand surface area, body weight, daily breathing rate and shower and bathroom breathing rate are known.

[4]Bathtub with shower used in this Report as there was no indication in depositions of availability of separate bathtubs.

[5]Main house location used in this Report.

[6]Default of bathroom used in this Report; no other information available.

[7]See main Report for information on user-specified inputs.

[8]Defaults used in the Report; default start time of 9:00 pm would be conservative as people are modeled to be back in residence by that time.

[9] Defaults used in the Report; default start time of 7:00 pm would be conservative as people are generally modeled to be back in residence by that time.

APPENDIX 3: Supporting information

Appendix 3.1 Effect of number of people using communal facilities on indoor daily air concentration (partial reports)

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

30-Person Shower and Locker Room Facility

Site and Model Input Information

Information	Report Setting
Site Information	—
Site name:	Number of people and concentrations
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	Custom Shower and Locker Room Scenario
Building type:	Dorm or Barracks
Number of persons using the facility:	30

Information	Report Setting
Percent of people taking showers:	100%

Scenario Description

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 30 persons in a dorm or barracks. It provides information about the scenario parameters and the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 30 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Quick Summary

Table S1 shows the CTE and RME average daily exposure concentrations for all persons using the facility. These daily exposure concentrations are 24-hour exposure concentrations derived from the exposure people experience while in the building.

Table S1. CTE and RME daily exposure concentrations for all persons using the facility

Exposure Type	Percentile Exposure	Daily Exposure Concentration (µg/m³)
CTE	50th	0.058
RME	95th	0.21

Abbreviations: CTE = central tendency exposure; µg/m³ = micrograms chemical per cubic meter air; RME = reasonable maximum exposure

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

60-Person Shower and Locker Room Facility

Site and Model Input Information

Information	Report Setting
Site Information	—
Site name:	Number of people and concentrations
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	Custom Shower and Locker Room Scenario
Building type:	Dorm or Barracks
Number of persons using the facility:	60
Percent of people taking showers:	100%

Scenario Description

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 60 persons in a dorm or barracks. It provides information about the scenario parameters and

the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 60 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Quick Summary

Table S1 shows the CTE and RME average daily exposure concentrations for all persons using the facility. These daily exposure concentrations are 24-hour exposure concentrations derived from the exposure people experience while in the building.

Table S1. CTE and RME daily exposure concentrations for all persons using the facility

Exposure Type	Percentile Exposure	Daily Exposure Concentration (µg/m³)
CTE	50th	0.044
RME	95th	0.14

Abbreviations: CTE = central tendency exposure; µg/m^3 = micrograms chemical per cubic meter air; RME = reasonable maximum exposure

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

120-Person Shower and Locker Room Facility

Site and Model Input Information

Information	Report Setting
Site Information	—
Site name:	Number of people and concentrations
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	Custom Shower and Locker Room Scenario
Building type:	Dorm or Barracks
Number of persons using the facility:	120
Percent of people taking showers:	100%

Scenario Description

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 120 persons in a dorm or barracks. It provides information about the scenario parameters and

the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 120 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Quick Summary

Table S1 shows the CTE and RME average daily exposure concentrations for all persons using the facility. These daily exposure concentrations are 24-hour exposure concentrations derived from the exposure people experience while in the building.

Table S1. CTE and RME daily exposure concentrations for all persons using the facility

Exposure Type	Percentile Exposure	Daily Exposure Concentration (µg/m³)
CTE	50th	0.039
RME	95th	0.10

Abbreviations: CTE = central tendency exposure; µg/m^3 = micrograms chemical per cubic meter air; RME = reasonable maximum exposure

Appendix 3.2 Constant versus peak facility usage (partial reports).

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

30-Person Shower and Locker Room Facility

Site and Model Input Information

Information	Report Setting
Site Information	—
Site name:	constant versus peak (constant)
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	**Modified Custom Shower and Locker Room Scenario**
Building type:	Dorm or Barracks
Number of persons using the facility:	30

Information	Report Setting
Percent of people taking showers:	100%

Scenario Description

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 30 persons in a dorm or barracks. It provides information about the scenario parameters and the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 30 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Quick Summary

Table S1 shows the CTE and RME average daily exposure concentrations for all persons using the facility. These daily exposure concentrations are 24-hour exposure concentrations derived from the exposure people experience while in the building.

Table S1. CTE and RME daily exposure concentrations for all persons using the facility

Exposure Type	Percentile Exposure	Daily Exposure Concentration (µg/m^3)
CTE	50th	0.25
RME	95th	0.56

Abbreviations: CTE = central tendency exposure; µg/m^3 = micrograms chemical per cubic meter air; RME = reasonable maximum exposure

Operating Hours	—
Schedule type	Open all day

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

30-Person Shower and Locker Room Facility

Site and Model Input Information

Information	Report Setting
Site Information	—
Site name:	constant versus peak (peak 6:00 – 7:00 am)
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	**Modified Custom Shower and Locker Room Scenario**
Building type:	Dorm or Barracks
Number of persons using the facility:	30

Information	Report Setting
Percent of people taking showers:	100%

Scenario Description

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 30 persons in a dorm or barracks. It provides information about the scenario parameters and the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 30 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Quick Summary

Table S1 shows the CTE and RME average daily exposure concentrations for all persons using the facility. These daily exposure concentrations are 24-hour exposure concentrations derived from the exposure people experience while in the building.

Table S1. CTE and RME daily exposure concentrations for all persons using the facility

Exposure Type	Percentile Exposure	Daily Exposure Concentration (µg/m³)
CTE	50th	0.28
RME	95th	0.62

Abbreviations: CTE = central tendency exposure; µg/m³ = micrograms chemical per cubic meter air; RME = reasonable maximum exposure

Operating Hours	—
Schedule type	Open all day
Peak Periods	—
Peak period 1 start time	**12:06 a.m.**
Peak period 1 end time	**12:07 a.m.**

Abbreviations: hr = hours; % = percent; min/day = minutes per day

Table 9. Activity duration distribution parameters

Parameter	Value
Shower duration mean	20 min
Shower duration standard deviation	5.3 min

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

30-Person Shower and Locker Room Facility

Site and Model Input Information

Information	Report Setting
Site Information	—
Site name:	continuous versus peak/ 6:00-7:00 pm
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	**Modified Custom Shower and Locker Room Scenario**
Building type:	Dorm or Barracks
Number of persons using the facility:	30
Percent of people taking showers:	100%

Scenario Description

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 30 persons in a dorm or barracks. It provides information about the scenario parameters and the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 30 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Quick Summary

Table S1 shows the CTE and RME average daily exposure concentrations for all persons using the facility. These daily exposure concentrations are 24-hour exposure concentrations derived from the exposure people experience while in the building.

Table S1. CTE and RME daily exposure concentrations for all persons using the facility

Exposure Type	Percentile Exposure	Daily Exposure Concentration (µg/m³)
CTE	50th	0.28
RME	95th	0.62

Abbreviations: CTE = central tendency exposure; µg/m^3 = micrograms chemical per cubic meter air; RME = reasonable maximum exposure

Operating Hours	—
Schedule type	Open all day
Peak Periods	—

Peak period 1 start time	**12:18 a.m.**
Peak period 1 end time	**12:19 a.m.**

Abbreviations: hr = hours; % = percent; min/day = minutes per day

Table 9. Activity duration distribution parameters

Parameter	Value
Shower duration mean	20 min
Shower duration standard deviation	5.3 min

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

30-Person Shower and Locker Room Facility

Site and Model Input Information

Information	Report Setting
Site Information	—
Site name:	model default values
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	Custom Shower and Locker Room Scenario
Building type:	Dorm or Barracks
Number of persons using the facility:	30
Percent of people taking showers:	100%

Scenario Description

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 30 persons in a dorm or barracks. It provides information about the scenario parameters and the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 30 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Quick Summary

Table S1 shows the CTE and RME average daily exposure concentrations for all persons using the facility. These daily exposure concentrations are 24-hour exposure concentrations derived from the exposure people experience while in the building.

Table S1. CTE and RME daily exposure concentrations for all persons using the facility

Exposure Type	Percentile Exposure	Daily Exposure Concentration (µg/m³)
CTE	50th	0.058
RME	95th	0.21

Abbreviations: CTE = central tendency exposure; µg/m^3 = micrograms chemical per cubic meter air; RME = reasonable maximum exposure

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

30-Person Shower and Locker Room Facility

Site and Model Input Information

Information	Report Setting
Site Information	—
Site name:	model site-specific values
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	**Modified Custom Shower and Locker Room Scenario**
Building type:	Dorm or Barracks
Number of persons using the facility:	30
Percent of people taking showers:	100%

Scenario Description

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 30 persons in a dorm or barracks. It provides information about the scenario parameters and the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 30 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Quick Summary

Table S1 shows the CTE and RME average daily exposure concentrations for all persons using the facility. These daily exposure concentrations are 24-hour exposure concentrations derived from the exposure people experience while in the building.

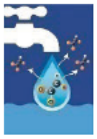
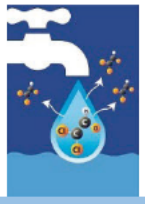


Table S1. CTE and RME daily exposure concentrations for all persons using the facility

Exposure Type	Percentile Exposure	Daily Exposure Concentration (µg/m³)
CTE	50th	0.040
RME	95th	0.14

Abbreviations: CTE = central tendency exposure; µg/m^3 = micrograms chemical per cubic meter air; RME = reasonable maximum exposure

Appendix 3.4 Comparison of inhalation doses for 30-person barracks setting and 10-person office building (partial reports).



ATSDR SHOWER Model Report

Custom Office Scenario

10-Person Bathroom-only Facility

Information	Report Setting
Site Information	—
Site name:	office
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	Custom Bathroom Scenario
Building type:	Office
Number of persons using the facility:	10

This report is for a custom scenario Monte Carlo simulation of a bathroom facility used by 10 persons in an office or similar setting. It provides information about the scenario parameters

and the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 10 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Table 3. CTE and RME daily inhalation doses for all persons using the facility

Exposure Group	CTE Inhalation Dose (µg/kg/day)	RME Inhalation Dose (µg/kg/day)
Full–time worker	0.0015	0.0036

Abbreviations: CTE = central tendency exposure; µg/kg/day = micrograms chemical per kilograms body weight per day; RME = reasonable maximum exposure

The following tables and figures present the parameter values that were used to run this custom scenario. These tables are provided for reference and generally are not reported in your public health documents. References for the default parameters used in this scenario can be found in ATSDR's Guidance for Evaluating Inhalation and Dermal Exposure Using the SHOWER Model (ATSDR 2024a).

In Table 4, the term f value refers to the percentage of a chemical that will be released from a water source (e.g., shower water) to air. Chemical f values are both chemical- and appliance-specific, such that the same chemical will have different f values for different appliances. More

information about f values and their derivation can be found in the SHOWER model technical document (ATSDR 2024b).

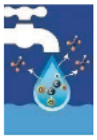


Table 6. Facility information

Parameter	Value
Number of Appliances	—
Toilets	1
Bathroom sinks	1

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

30-Person Shower and Locker Room Facility

Information	Report Setting
Site Information	—
Site name:	barracks
Address:	—
Application:	Version 4.0.0
CASRN:	127-18-4
Contaminant:	Tetrachloroethylene
Synonym:	1,1,2,2-perchloroethylene 1,1,2,2-tetrachloroethylene PCE PERC Perchloroethylene Tetrachloroethene
Model Input Information	—
Chemical name:	Tetrachloroethylene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	10 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	**Modified Custom Shower and Locker Room Scenario**
Building type:	Dorm or Barracks
Number of persons using the facility:	30
Percent of people taking showers:	100%

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 30 persons in a dorm or barracks. It provides information about the scenario parameters and the central tendency exposure (CTE) and reasonable maximum exposure (RME)

results of the Monte Carlo simulation for 30 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Table 3 shows the CTE and RME daily inhalation doses in µg/kg/day, which are derived from age-specific breathing rates and the average daily exposure concentrations calculated for each person in each Monte Carlo iteration.

Table 3. CTE and RME daily inhalation doses for all persons using the facility

Exposure Group	CTE Inhalation Dose (µg/kg/day)	RME Inhalation Dose (µg/kg/day)
Adult	0.056	0.12

Abbreviations: CTE = central tendency exposure; µg/kg/day = micrograms chemical per kilograms body weight per day; RME = reasonable maximum exposure

Abbreviations: kg = kilograms body weight; L/min = liters air breathed per minute; cm^2 = square centimeters

Table 6. Facility information

Parameter	Value
Number of Appliances	—
Showers	3
Toilets	2
Bathroom sinks	1
Area Volumes	—
Shower area volume	15 m^3 (e.g., 65 ft^2 x 8 ft)
Locker room and bathroom volume	19 m^3 (e.g., 84 ft^2 x 8 ft)
Air Exchange Rates	—
Shower area air exchange rate when facility is open	1.4 ACH
Locker room and bathroom air exchange rate when facility is open	1.4 ACH

Abbreviations: ACH = air changes per hour; m^3 = cubic meters air

Table 9. Activity duration distribution parameters

Parameter	Value
Shower duration mean	**20 min**
Shower duration standard deviation	5.3 min
Average time in locker room before shower	5 min
Average time in locker room after shower	10 min
Average time in bathroom for bathroom-only visits	5 min
Bathroom sink use duration mean	0.61 min
Bathroom sink use duration standard deviation	0.57 min

Abbreviations: min = minute

APPENDIX 4: PHAST model parameters and options for modifications

The following table includes the inputs to the ATSDR PHAST model for drinking water ingestion and the exposure assessment factors that can be modified.

Factor type	Options
Chemical information	
Concentration in water	User-specified
Units	User-specified (e.g., ppm, ppb)
Type	User-specified (e.g., arithmetic mean, geometric mean)
Exposure groups and body weights	
Exposure group	Residential, daycare, school, or occupational
Age groups and body weights	Default and/or customized
Intake rates	
Drinking water intake rate	Default (CTE, RME) or and/or site-specific intake rates

ATSDR SHOWER Model Report

Custom Dorm or Barracks Scenario

40-Person Shower and Locker Room Facility

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Site and Model Input Information

Information	Report Setting
Site Information	—
Site name:	KELLER HP BARRACKS v4.0.1
Address:	—
Application:	Version 4.0.1
CASRN:	71-43-2
Contaminant:	Benzene
Synonym:	—
Model Input Information	—
Chemical name:	Benzene
Chemical properties:	Default
Exposure routes available:	Inhalation and Dermal
Water concentration:	2.9 µg/L
Outdoor air concentration:	0 µg/m³
Main building air concentration:	0 µg/m³
Scenario type:	**Modified Custom Shower and Locker Room Scenario**
Building type:	Dorm or Barracks
Number of persons using the facility:	40
Percent of people taking showers:	100%

Scenario Description

This report is for a custom scenario Monte Carlo simulation of a shower and locker room facility used by 40 persons in a dorm or barracks. It provides information about the scenario parameters and the central tendency exposure (CTE) and reasonable maximum exposure (RME) results of the Monte Carlo simulation for 40 persons. This report uses scientific notation for numbers greater than 10,000 (1.0E+4) or less than 0.001 (1.0E-3). Parameter values in this custom scenario that differ from the default parameters for the standard scenario are highlighted in **red**.

Unlike residential scenarios, which predefine resident activity patterns and specify a target person for calculating results, communal shower and bathroom scenarios do not predefine facility user activity patterns or specify a target person. Instead, they use Monte Carlo methods to randomly generate activity patterns based on input parameter distributions, and then derive CTE and RME statistics from the inhalation concentration and dermal dose distributions in the simulation output. The CTE result is the exposure for the person with the 50th percentile (median) exposure across all Monte Carlo iterations, and the RME result is the exposure for the person with the 95th percentile exposure.

The Monte Carlo simulation conducted for this report generated 1,000 iterations of the scenario. Because each of the 1,000 iterations is randomly generated, the report output may change slightly from simulation to simulation.

Quick Summary

Table S1 shows the CTE and RME average daily exposure concentrations for all persons using the facility. These daily exposure concentrations are 24-hour exposure concentrations derived from the exposure people experience while in the building.

Table S1. CTE and RME daily exposure concentrations for all persons using the facility

Exposure Type	Percentile Exposure	Daily Exposure Concentration (µg/m³)
CTE	50th	0.080
RME	95th	0.18

Abbreviations: CTE = central tendency exposure; µg/m³ = micrograms chemical per cubic meter air; RME = reasonable maximum exposure

Table S2 presents the CTE and RME dermal doses from contact with water for all persons using the facility. This contact occurs from showering and from hand washing while in the locker room.

Table S2. CTE and RME daily administered dermal dose for all persons using the facility

Exposure Group	CTE Dermal Dose (µg/kg/day)	RME Dermal Dose (µg/kg/day)
Adult	0.0071	0.0087

Abbreviations: CTE = central tendency exposure; µg/kg/day = micrograms chemical per kilograms body weight per day; RME = reasonable maximum exposure

Supplemental Information

The following sections of the report provide supplemental information about the scenario beyond what is provided in the Quick Summary section. They include additional summary statistics for the Monte Carlo simulation, one or more histograms of model results, and multiple tables showing the input parameters used in the scenario. Including this information in your public health document is optional.

Monte Carlo Simulation Summary Statistics

Table 1 provides summary statistics about the randomly generated facility usage patterns for this scenario across all Monte Carlo iterations. The values in the 50th and 95th percentile columns represent statistics for each parameter individually and do not correspond with the CTE and RME results from the Quick Summary section. For example, the 50th percentile shower duration is the median shower time for all people that take showers, not the shower duration for the person with the 50th percentile

exposure. For persons that shower in the facility, the time in shower and locker room accounts for the time that people spend in the locker room getting ready to shower and the time that they spend in the locker room getting dressed after their shower, in addition to time spent using the shower, toilet, and bathroom sink. For persons that do not shower in the facility, the time in locker room accounts for time spent using the toilet and bathroom sink.

Table 1. Facility usage summary statistics for all persons using the facility

Parameter	Mean	50th percentile	95th percentile
Shower duration	20 min	19 min	30 min
Time in shower and locker room	53 min	53 min	66 min
Time using bathroom sink	2.5 min	2.4 min	4.5 min
Time in building	17 hr	16 hr	24 hr

Abbreviations: hr = hours; min = minute

Scenario Results – Figures

Figure 1 shows a histogram of the inhalation exposure concentrations for all persons using the facility. The figure also shows the CTE (50th percentile) and RME (95th percentile) average daily exposure concentrations for persons using the facility.

Figure 1. Histogram of inhalation exposure concentrations for all persons using the facility

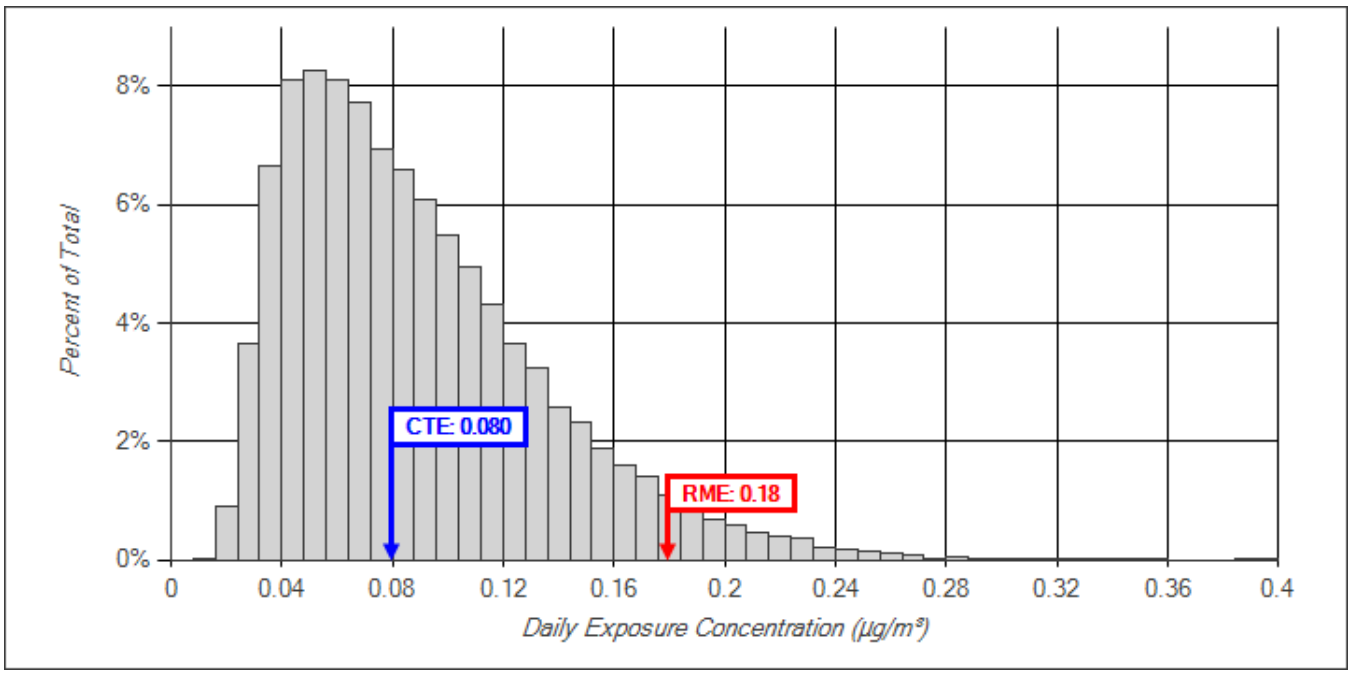
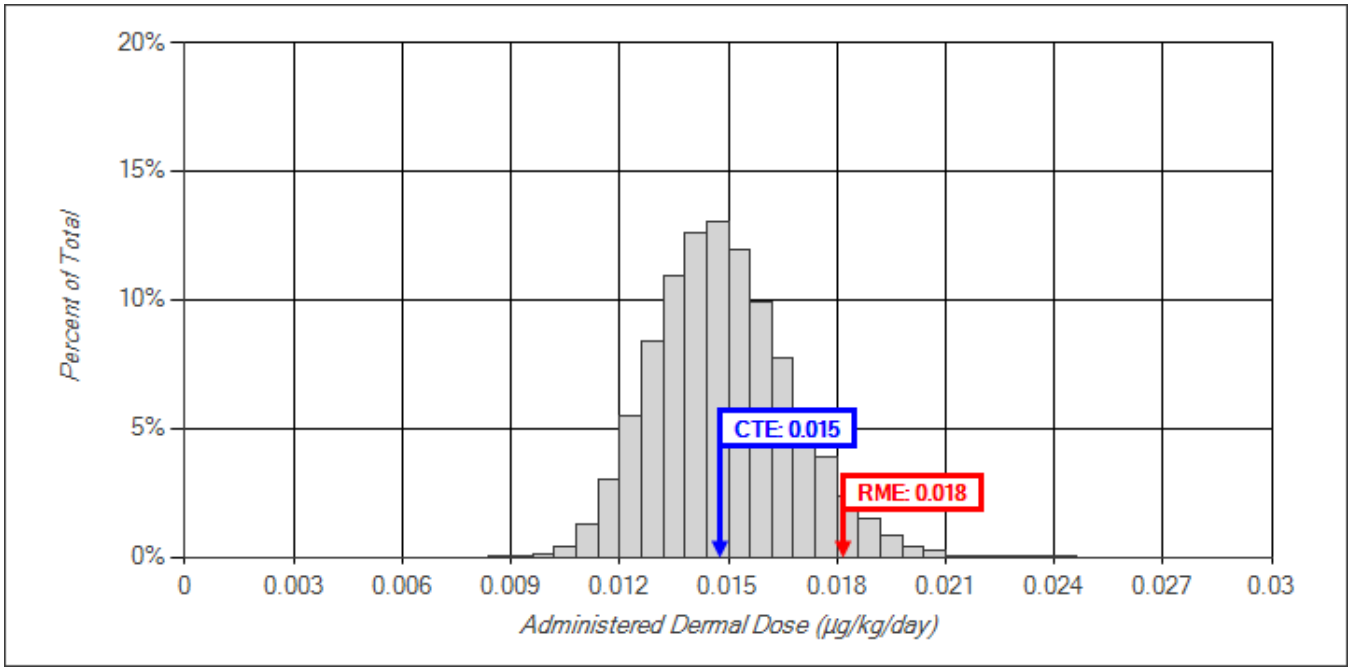


Figure 2 shows a histogram of dermal doses for all persons using the facility. Multiple doses were calculated for each person based on the exposure groups in the scenario, but the histogram shows only the highest dose calculated for each person. The CTE and RME flags identify the 50th and 95th percentile doses in the histogram.

Figure 2. Histogram of administered dermal doses for all persons using the facility



Scenario Results – Inhalation Concentration and Dose Tables

Peak Exposure By Location (Table 2)

The locations simulated in this scenario consist of a shower area, a locker room and bathroom, and the main building. Table 2 shows the 50th and 95th percentile time-weighted average exposure concentrations in each of these locations. Similar to the statistics reported in Table 1, the values in the 50th and 95th percentile columns represent statistics for each value individually and do not correspond with the CTE and RME results from the Quick Summary section.

The exposure from taking a shower and being in the locker room can be much higher (but for shorter periods) than the exposure from being in the main building. Knowledge of this brief exposure to high levels in the shower and locker room might be useful when evaluating whether harmful effects might be possible from acute exposure to high concentrations. This acute exposure to high levels might be particularly important for irritant chemicals, such as formaldehyde, 2-butanone, and acetone. Some irritants, however, cannot be run using the model because parameters are lacking. Health assessors should evaluate this acute exposure duration if the acute EMEG is exceeded. More information about evaluating acute exposure can be found in ATSDR's Guidance for Evaluating Inhalation and Dermal Exposure Using the SHOWER Model (ATSDR 2024a). **Health assessors should consult with the Associate Director of Science (ADS) when evaluating brief exposure to high levels.**

Table 2. 50th and 95th percentile exposure time and time-weighted average exposure concentration by location for all persons using the facility

Location	50th Percentile Exposure Time (min)	50th Percentile Time-weighted Average Exposure Concentration (µg/m³)	95th Percentile Exposure Time (min)	95th Percentile Time-weighted Average Exposure Concentration (µg/m³)
Shower area	20.0	5.5	30.0	11
Locker room	33.0	0.20	41.0	0.29
Main building	952.0	0	1,398.0	0

Abbreviations: µg/m³ = micrograms chemical per cubic meter air; min = minute

Inhalation Doses (Table 3)

Table 3 shows the CTE and RME daily inhalation doses in µg/kg/day, which are derived from age-specific breathing rates and the average daily exposure concentrations calculated for each person in each Monte Carlo iteration.

Table 3. CTE and RME daily inhalation doses for all persons using the facility

Exposure Group	CTE Inhalation Dose (µg/kg/day)	RME Inhalation Dose (µg/kg/day)
Adult	0.018	0.040

Abbreviations: CTE = central tendency exposure; µg/kg/day = micrograms chemical per kilograms body weight per day; RME = reasonable maximum exposure

Model Parameters

The following tables and figures present the parameter values that were used to run this custom scenario. These tables are provided for reference and generally are not reported in your public health documents. References for the default parameters used in this scenario can be found in ATSDR's Guidance for Evaluating Inhalation and Dermal Exposure Using the SHOWER Model (ATSDR 2024a).

In Table 4, the term f value refers to the percentage of a chemical that will be released from a water source (e.g., shower water) to air. Chemical f values are both chemical- and appliance-specific, such that the same chemical will have different f values for different appliances. More information about f values and their derivation can be found in the SHOWER model technical document (ATSDR 2024b).

Table 4. Chemical properties

Parameter	Value
µg/m³ to ppb conversion factor	1 µg/m³ = 0.313 ppb
Inhalation Parameters	—
Shower f value	0.4878
Bathroom sink f value	0.2324
Toilet f value	0.232
Henry's law constant	0.2307
Dermal Parameters	—
Chemical type	Organic
Molecular weight (MW)	78.12 g/mol
Dermal permeability coefficient (K_p)	0.0149 cm/hr
Fraction absorbed through skin (FA)	1
Fraction absorbed in gastrointestinal tract (ABS_{GI})	1
Permeability coefficient ratio (B)	0.051
Lag time per event (τ_{event})	0.29 hr/event
Time to reach steady state (t*)	0.69 hr

Abbreviations: cm/hr = centimeters per hour; g/mol = grams chemical per mole; hr = hours; hr/event = hours per event; µg/m³ = micrograms chemical per cubic meter air; ppb = parts chemical per billion parts air

Table 5. Mean parameters used to calculate inhalation and dermal doses

Exposure Group	Body Weight (kg)	Daily Breathing Rate (L/min)	Shower and Bathroom Breathing Rate (L/min)	Total Body Surface Area (cm^2)	Hand Surface Area (cm^2)
Adult	80.0	10.53	12.34	2.0E+4	980

Abbreviations: kg = kilograms body weight; L/min = liters air breathed per minute; cm^2 = square centimeters

Table 6. Facility information

Parameter	Value
Number of Appliances	—
Showers	4
Toilets	3
Bathroom sinks	1
Area Volumes	—
Shower area volume	20 m^3 (e.g., 86 ft^2 x 8 ft)
Locker room and bathroom volume	25 m^3 (e.g., 112 ft^2 x 8 ft)
Air Exchange Rates	—
Shower area air exchange rate when facility is open	1.4 ACH
Locker room and bathroom air exchange rate when facility is open	1.4 ACH

Abbreviations: ACH = air changes per hour; m^3 = cubic meters air

Table 7. Appliance parameters

Appliance	Parameter	Value
Shower area	—	—
Shower	Flow rate	7.6 L/min
Shower exhaust fan	Total exhaust rate	2,400 L/min
Locker room and bathroom	—	—
Bathroom sink	Flow rate	3.34 L/min
Toilet	Volume per flush	8.7 L/flush
Bathroom exhaust fan	Total exhaust rate	4,500 L/min
Supply vent	Outdoor air supply rate	1,325 L/min

Abbreviations: L/min = liters per minute; L/flush = liters water per flush

Table 8. Daily facility usage parameters

Parameter	Value
Activity Parameters	—
Time each person spends in the building	Standard distribution
Average bathroom visits per person	5
Percent of people taking showers	100%
Time in Building Standard Distribution Percentiles	—
Minimum	8.0 min/day (0.13 hr/day)
5th percentile	575 min/day (9.6 hr/day)
25th percentile	795 min/day (13.3 hr/day)
50th percentile	985 min/day (16.4 hr/day)
75th percentile	1,235 min/day (20.6 hr/day)
90th percentile	1,395 min/day (23.3 hr/day)
95th percentile	1,440 min/day (24 hr/day)
98th percentile	1,440 min/day (24 hr/day)
99th percentile	1,440 min/day (24 hr/day)
Maximum	1,440 min/day (24 hr/day)
Operating Hours	—
Schedule type	Open all day
Peak Periods	—
Peak period 1 start time	**6:00 a.m.**
Peak period 1 end time	**7:00 a.m.**

Abbreviations: hr = hours; % = percent; min/day = minutes per day

Table 9. Activity duration distribution parameters

Parameter	Value
Shower duration mean	**20 min**
Shower duration standard deviation	5.3 min
Average time in locker room before shower	5 min
Average time in locker room after shower	10 min
Average time in bathroom for bathroom-only visits	5 min
Bathroom sink use duration mean	0.61 min
Bathroom sink use duration standard deviation	0.57 min

Abbreviations: min = minute

Facility Visit Statistics

Table 10 provides statistics on the facility visits simulated across all Monte Carlo iterations. The expected number of facility visits equals the product of the number of facility users, the average bathroom visits per person, and the number of Monte Carlo iterations. The actual number of simulated visits should be close to the expected number. Successful visits were those in which a person was able to use the facility, and unsuccessful visits were those in which the person was not, for one of the reasons described in the table. If the percentage of unsuccessful visits is 10% or more, a significant fraction of the simulated visits did not occur. This situation can occur for various reasons. For example, if you run a scenario where 1,000 people try to use a facility with only 1 shower, toilet, and bathroom sink, most of them will not be able to use the facility because it is at capacity, and the unsuccessful visit rate will be much higher than 10%. If the unsuccessful visit rate is 10% or more, you should review the reasons that the visits were unsuccessful and confirm that they are physically meaningful for your scenario. If they are not, review the inputs to your scenario for potential errors. See the ATSDR technical document (ATSDR 2024b) for more information, and for additional assistance, contact showermodel@cdc.gov.

Table 10. Facility Visit Statistics

Statistic	Value	Percent of Simulated Visits
Expected Visit Parameters	—	—
Facility users	40	NA
Average bathroom visits per person	5.0	NA
Monte Carlo iterations	1,000	NA
Expected simulated visits	200,000	NA
Simulated Visit Statistics	—	—

Statistic	Value	Percent of Simulated Visits
Simulated visits	199,702	NA
Successful visits	177,567	89%
Unsuccessful visits	22,135	11%
Reasons Visits Were Unsuccessful	—	—
Insufficient time in bathroom to use toilet and sink	0	0%
Facility visit overlapped with an earlier visit	1	5.0E-4%
Facility visit ended after midnight	31	0.016%
Part of the facility visit occurred when the facility was closed	0	0%
The facility was already at maximum occupancy	40	0.020%
A shower was not available	235	0.12%
A toilet was not available	62	0.031%
A bathroom sink was not available	21,766	11%
Total facility visit time was greater than total time in building	0	0%

Abbreviations: NA = not applicable; % = percent

References

[ATSDR] Agency for Toxic Substances and Disease Registry. 2024a. Guidance for Evaluating Inhalation and Dermal Exposure Using the Shower and Household Water-use Exposure (SHOWER) Model v4.0. Atlanta, GA: U.S. Department of Health and Human Services, Public Health Service.

[ATSDR] Agency for Toxic Substances and Disease Registry. 2024b. Technical Document for the Shower and Household Water-use Exposure (SHOWER) Model v4.0. Atlanta, GA: U.S. Department of Health and Human Services, Public Health Service.